

Distr. GENERAL



# Convention on Biological Diversity

UNEP/CBD/SBSTTA/17/INF/16 12 October 2013

ORIGINAL: ENGLISH

SUBSIDIARY BODY ON SCIENTIFIC, TECHNICAL AND TECHNOLOGICAL ADVICE Seventeenth meeting Montreal, 14-18 October 2013 Item 3 of the provisional agenda\*

# **REVIEW OF THE USE OF REMOTELY-SENSED DATA FOR MONITORING BIODIVERSITY CHANGE AND TRACKING PROGRESS TOWARDS THE AICHI BIODIVERSITY TARGETS**

### Note by the Executive Secretary

1. The Executive Secretary is circulating herewith, for the information of participants in the seventeenth meeting of the Subsidiary Body on Scientific, Technical and Technological Advice, the "Review of the use of remotely-sensed data for monitoring biodiversity change and tracking progress towards the Aichi Biodiversity Targets".

2. The report has been prepared by the United Nations Environment Programme – World Conservation Monitoring Centre in association with Biodiversity Indicators Partnership and the Group on Earth Observations Biodiversity Observations Network (GEO BON) with the financial support of the European Commission and the Federal Office for the Environment (FOEN), Switzerland.

3. The preliminary conclusions of the report are presented in document UNEP/CBD/SBSTTA/17/2, paragraph 39.

4. The report is presented in the form and language in which it was received by the Secretariat.

5. It is envisaged that, following further review of this report, a revised version will be published in the CBD Technical Series. Comments on the report are therefore invited and should be sent to secretariat@cbd.int.

<sup>\*</sup> UNEP/CBD/SBSTTA/17/1.

In order to minimize the environmental impacts of the Secretariat's processes, and to contribute to the Secretary-General's initiative for a C-Neutral UN, this document is printed in limited numbers. Delegates are kindly requested to bring their copies to meetings and not to request additional copies.

Review of the use of remotelysensed data for monitoring biodiversity change and tracking progress towards the Aichi Biodiversity Targets

#### Authors

**Lead authors:** Cristina Secades, Brian O'Connor, Claire Brown and Matt Walpole, United Nations Environment Programme World Conservation Monitoring Centre (UNEP-WCMC)

**Contributing authors:** Andrew Skidmore, Tiejun Wang, Thomas Groen, Matt Herkt and Aidin Niamir (University of Twente); Amy Milam (independent consultant); Zoltan Szantoi, Evangelia Drakou, Juliana Stropp, Joysee M. Rodriguez and Aymen Chartef, Joint Research Centre (JRC); Alexander Held, AusCover Facility of the Terrestrial Ecosystem Research Network (TERN) and Commonwealth Scientific and Industrial Research Organization (CSIRO); Heather Terrapon, South Africa National Biodiversity Institute (SANBI); Nicholas Coops, University British Columbia (UBC); Michael Wulder, Canadian Forest Service (CFS); Trisalyn Nelson, University of Victoria (UVic) and Margaret Andrew, Murdoch University with the support of Ryan Powers, Jessica Fitterer and Shanley Thompson; and, Jose Carlos Epiphano, Brazil National Institute for Space Research (INPE).

#### Acknowledgements

The authors wish to express deep gratitude to the following experts: Bob Scholes, South Africa Council for Scientific and Industrial Research (CSIR); Edward Mitchard, Edinburgh University; France Gerard, Centre for Ecology and Hydrology (NERC); Hervé Jeanjean, French Space Agency (CNES); Marc Paganini, European Space Agency (ESA); Woody Turner, National Aeronautics and Space Administration (NASA); Mark Spalding, The Nature Conservancy (TNC); Matthew Hansen, University of Maryland; Peter Fretwell, British Antarctic Survey (BAS); Rob Rose, Wildlife Conservation Society (WCS); Ruth de Fries, Columbia University; Ruth Swetnam, Stafforshire University; Colette Wabnitz, Secretariat of the Pacific Community (SPC); Susana Baena, Kew Royal Botanic Gardens; Gregoire Dubois (JRC); Gilberto Camara (INPE); Chen Jun, National Geomatics Center of China; Yichuan Shi, International Union for Conservation of Nature (IUCN); Andreas Obrech, Federal Office for the Environment, Swiss Government; Natalie Petorelli, Zoological Society of London (ZSL); Martin Wegmann, Committee on Earth Observation Satellites (CEOS); David Cooper and Robert Höft, Secretariat of the Convention on Biological Diversity (CBD); Jon Hutton, Lera Miles, Neil Burgess, Max Fancourt and Jan-Willem (UNEP-WCMC).

UNEP-WCMC would like to gratefully acknowledge the financial support of the European Commission and the Federal Office for the Environment (FOEN), Switzerland.

#### **Legal notice**

The views reported in this review do not necessarily represent those of UNEP-WCMC, the CBD, or those of other contributing organizations, authors or reviewers. The designations employed and the presentations do not imply the expressions of any opinion whatsoever on the part of UNEP-WCMC concerning the legal status of any country, territory, city or area and its authority, or concerning the delimitation of its frontiers or boundaries.



## **Key messages**

- 1. The potential for remotely sensed earth observation data to support biodiversity policy is growing, but is yet to be fully realised. While many remote sensing products are demonstration activities and thus lack longer time series that would allow for temporal change analysis, there are an increasing number of robust environmental time series data sets available from remote sensing platforms. The value of remote sensing is dependent upon sustained observations over the longer term.
- 2. There are clear opportunities presented by existing and emerging remote sensing technologies to support monitoring of the Aichi Biodiversity Targets. Key areas of development surround land cover change and water/air quality (Aichi Targets 5 and 8), although innovations in other areas offer additional opportunities including helping to fill some of the key gaps for Targets for which is has proven difficult to develop indicators using only *in situ* data (such as Aichi Target 9 and 14), and assessing effectiveness of conservation actions (Aichi Target 11). However, *in situ* data and statistical modelling are also frequently required to create comprehensive indicators, and these are not always available.
- 3. Remotely sensed data, when processed, packaged and communicated appropriately, can have impacts on policy and practice that yield positive biodiversity outcomes. Current scientific understanding, computational power and web architecture create the possibility for automated products providing spatially explicit change analyses and alerts in 'near real time', in particular for forest cover.
- 4. However, the use of remotely sensed earth observation data is often constrained by access to data and processing capacity. Whilst some data of appropriate spatial and temporal coverage and resolution are freely available, access to other potentially valuable and complementary data incurs a financial cost. Free and open access to all taxpayer-funded satellite remote sensing imagery would address this significant constraint. However, even where data are accessible, the considerable amounts of data being generated are not being effectively used. Significant computational power and human resources are required on an ongoing basis to process the data and create the kinds of periodically updated analytical products suitable to inform indicators and assessments of progress towards the Aichi Targets.
- 5. Priorities for future development of remote sensing products should be driven by end users needs. Given resource constraints, priorities must be established. An agreed set of minimum requirements and common standards to focus the efforts of the earth observation experts would be valuable. Moreover, a significant requirement is for a long-term, consistent and repeteable land cover change product. Monitoring land cover change over time can identify where pressures are occurring and how likely they are to impact current status and future trends in global biodiversity.
- 6. Creating a dialogue between data providers and users is key to realising the potential of remotely sensed data. To date, this dialogue has been limited. A closer relationship between the earth observation community and potential users in the biodiversity policy and management communities would help to enhance understanding, align priorities, identify opportunities and overcome challenges, ensuring data products more effectively meet user needs.

# **Table of Contents**

1.	Introduction	. 5
	1.1 Background and purpose	. 5
	1.2 Scope and definitions	. 7
	1.3 Approach	. 7
	1.4 Structure of the review	. 8
2.	Remote sensing opportunities for monitoring the Aichi Targets	10
	2.1 Overview	10
	2. Target by target assessment	13
3.	Lessons learnt from national experiences	36
	3.1 Remote sensing as a surveillance tool: fire monitoring in Australia	36
	3.2 The effectiveness of free open access data. The Brazilian example	37
	3.3 Using remote sensing for Protected Area planning in Canada	38
	3.4 Use of remote sensing in data creation for use in biodiversity indicators in South Africa	41
4.	Limitations and challenges	44
	4.1 What has limited the use of remote sensing in developing indicators?	44
	4.2 Key challenges in the use of remote sensing for indicator development	51
5.	Conclusions	53
6.	References	56
Aı	nnex 1. The basics of remote sensing in biodiversity monitoring	69
Aı	nnex 2. Overview of available remote sensing / Earth Observation products	78
Aı	nnex 3. Emerging applications of remote sensing in the context of the Convention	87
Aı	nnex 4. Detailed mapping of databases, remote sensing sensors, targets and indicators	93
Aı	nnex 5. Relative costs of using remote sensing for biodiversity monitoring	22

## **1. Introduction**

#### 1.1 Background and purpose

At the 10<sup>th</sup> meeting of the Conference of the Parties to the Convention on Biological Diversity (CBD COP 10) Parties, through decision X/2, adopted a Strategic Plan for Biodiversity 2011-2020, including twenty Aichi Biodiversity Targets. Parties committed to using these as a framework for setting national targets and to report on progress using indicators. During COP 11 an Indicator Framework for the Strategic Plan for Biodiversity 2011-2020 was adopted (Decision XI/3). It contains an indicative list of 98 indicators providing a flexible basis for Parties to assess progress towards the Aichi Biodiversity Targets.

Biodiversity indicators are a fundamental part of any monitoring system providing the mechanism for determining whether policies and actions are having the desired effect. They are also designed to communicate simple and clear messages to decision makers. Indicators use quantitative data to measure aspects of biodiversity, ecosystem condition, ecosystem services, and drivers of change, and aim to enhance understanding of how biodiversity is changing over time and space.

The CBD-mandated Biodiversity Indicators Partnership (BIP) is the global initiative to promote and coordinate development and biodiversity indicators in support of the Convention. The Partnership brings together over forty organizations working internationally on indicator development to provide the most comprehensive information on biodiversity trends. Established in 2007 to support monitoring of the 2010 Biodiversity Target, its mandate was renewed during CBD COP 11 (October 2012), becoming the principle vehicle for coordinating the development of biodiversity indicators at global, regional and national scales, and for delivery of indicator information for monitoring progress towards the Aichi Targets.

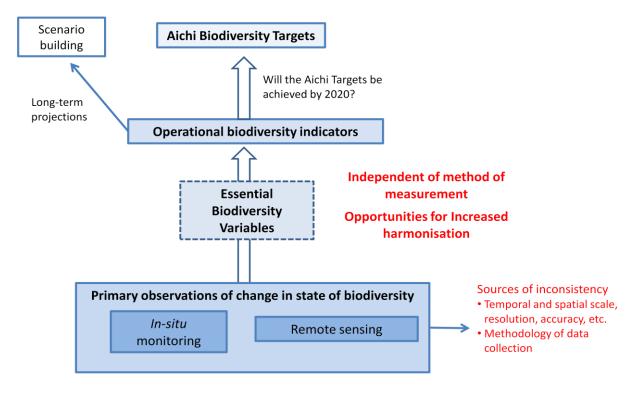
To create indicators requires observations, the collection of which may be guided by a set of agreed common variables, such as the proposed Essential biodiversity Variables (EBVs, Pereira *et al.*, 2013). The EBVs are being developed upon the request of the CBD with the aim to help prioritize by defining a minimum set of essential measurements to capture major dimensions of biodiversity change, and facilitate data integration by providing an intermediate linkbetween primary observations and indictors (Pereira *et al.* 2013). In the context of the Aichi Targets, the EBVs could offer a way to harmonize monitoring efforts carried out by different observation communities, helping the development of a global earth observation system. A number of candidate EBVs have been proposed to guide biodiversity observations. Such observations may be obtained *in situ* by direct, field measurements of individuals, populations, species, habitats, etc., or they may be collected at a distance using specialised instruments for *remote sensing* (Fig. 1).

*In situ* measurements offer the potential of extracting precise information on the existence and distribution of species. However, since field measurements are particularly time-consuming and expensive they are more practical for small scale, discrete data collection at sample sites rather than extensive, large scale monitoring. In addition, for certain highly variable ecosystems such as wetlands, or those located in remote areas, field-based observation might be difficult.

Remote sensing data, derived from both airborne and satellite sensors, promise a repeatable and cost effective manner to cover spatially extended areas contributing to biodiversity monitoring. However,

despite the wealth of remotely sensed data along a spectrum of sensors, wavelengths and resolutions, some of which are available free-of-charge, there is still limited use of remote sensing data for biodiversity monitoring that can detect biodiversity change in time as well as in space. Whilst in part this may be due to data and analytical constraints, it may also in part be due to a lack of adequate connection between user needs (including the specification of standards for each indicator) and opportunities provided by remotely sensed data.

Figure 1. The pathway to biodiversity indicators for the Aichi Biodviersity Targets from remotelysensed data and the role of EBVs.



Biodiversity scientists together with the world's major space agencies are beginning to explore the challenges and opportunities for the use of satellite remote sensing for biodiversity research applications. However, explicit policy needs such as biodiversity indicators have to date received little direct attention.

The present review of the use of remotely sensed data for monitoring biodiversity aims to contribute to fill this gap in the context of the CBD and the Aichi Biodiversity Targets. It has been produced on the request of the CBD Secretariat as a contribution to a developing effort to facilitate and expand the uptake of Earth Observations (EO) in the framework of the Convention. Its objectives are to:

1. Understand the main obstacles to, and identify opportunities for, greater use of remotely-sensed data and products in biodiversity monitoring and assessment.

2. Promote and facilitate enhanced, productive dialogue between the remote sensing community and policy end users through a shared understanding of needs and opportunities.

### **1.2 Scope and definitions**

This document is not intended to constitute a systematic or exhaustive review of all existing remote sensing technology, neither to be a highly technical discourse on their advantages and disadvantages. It aims to offer an accessible overview of the possibilities remotely-sensed data offers to track progress towards the Aichi Biodiversity Targets. Therefore, the content of the core body of the review has been developed with non-specialist policy-users in mind, with additional technical detail contained in the Annexes.

In the context of this review we have adopted the definition of remote sensing proposed by the United Nations in 1986 which defines the term *Remote Sensing* as "the sensing of the Earth's surface [...] by making use of the properties of electromagnetic wave emitted, reflected or diffracted by the sensed objects, for the purpose of improving natural resource management, land use and the protection of the environment" (UNGA A/RES/41/65). The review focuses on space-borne (satellite) sensors as they offer the greatest current potential for accessible global data coverage. However, the potential of airborne and ground-based sensors is also considered, as some ongoing developments could offer novel applications for biodiversity monitoring. A brief description of the different remote sensing technologies and how they can be used to monitor biodiversity can be found in Annex 1.

Spatial resolution is an important attribute of any digital image, describing the level of spatial detail which can be seen in the image. However, a balance must be struck between spatial detail in a satellite image and the field of view of the sensor recording the information conveyed in the image. Generally higher spatial detail requires a sensor with anarrower field of view hence less spatial coverage per image scene. Satellite sensors with a smaller field of view are generally constrained by low revisit times. Coarser spatial resolution sensors tend to image larger areas in one overpass of the satellite sensor with more regular repeat cycles. For the purposes of this report four categories of spatial resolution (in metres) have been defined:

- Very high resolution (<1m)
- High resolution (1- <15m)
- Medium resolution (15-30m)
- Low resolution (>30m)

## 1.3 Approach

The review was based on a desk study of available literature on remote sensing alongside an expert consultative process. An initial list of relevant literature was compiled by consultation with a small group four specialists in the application of remotely-sensed data, which was expanded afterwards following a thematic approach based in the literature referenced in the initial list of publications and by consultation with a larger group of 15 experts.

The expert consultation was conducted through a series of qualitative semi-structured surveys to compile expert knowledge. A group of around 30 specialists consisting of appropriate representatives from the major space agencies and remote sensing scientists/analysists and indicator specialists from

the international biodiversity policy community were selected to take part in the process. A questionnaire was specifically developed, structured in three sections: (1) technical and analytical section which focused on collecting information on ecological parameters and EO products currently used, how remotely-sensed data is produced, processed and consumed, and existing obstacles in each step; (2) indicators section, in which challenges in the use remotely-sensed data to develop indicators were discussed, and existing indicators derived from remote sensing recorded; and (3) future development section, in which interviewees had the opportunity to indicate up to three remote sensing priorities that could realistically be developed or improved within a 5-years framework that would significantly enhance the potential use of remote sensing for monitoring biodiversity. The survey was conducted in person or by telephone when possible, and through completion of the questionnaire in written in other case.

#### **1.4 Structure of the review**

The review is organized into an accessible main report of five sections supported by technical annexes.

Section 2 maps remote sensing products against each of the Aichi Biodiversity Targets. Opportunities, as well as gaps and limitations for the use of remote sensing to develop indicators for each target are highlighted.

Section 3 contains a number of case studies illustrating different approaches, methods and products used at national level to monitor diverse aspects of biodiversity, and their impact in decision-making and policy implementation.

Section 4 outlines the key limitations that have hindered the use of remotely-sensed data in indicator development to date, and the main challenges encountered. For most of them improvements and possible solutions are suggested using practical examples.

Section 5 summarises the key conclusions of the review and offers final thoughts and recommendations.

Annex 1 gives the reader a brief introduction to remote sensing methods and terminology, and compares these against traditional *in situ* measurements as a tool to monitor biodiversity. It answers common questions about what remote sensing is and how it is used.

Annex 2 analyses existing operational EO products according to their applications in biodiversity monitoring, and specifically in the framework of the CBD. Their potential for supporting the Strategic Plan for Biodiversity 2011-2020 and tracking progress towards the Aichi Biodiversity Targets is discussed.

Annex 3 introduces emerging applications of remote sensing for both marine and terrestrial environments relevant for biodiversity monitoring and outlines new areas of work and potential for future directions in the use of remote sensing in the context of the CBD.

Annex 4 contains a series of detailed tables mapping the various remote sensing products against the Aichi Targets and the EBVs in support of section 2. Information on spatial and temporal resolution suitable for global, regional and national levels, type of data and appropriate sensors required to develop each of the indicators contained in the indicative list of indicators (Decision XI/3) is described.

Potentially appropriate sensors for each Aichi Biodiversity Target and details of their characteristics are also provided (e.g. host organization, repeat viewing frequency, availability, data products).

Annex 5 provides a view on some of the costs involved in using remotely-sensed data that policy end users should take into account when planning to incorporate remote sensing in their monitoring systems.

## 2. Remote sensing opportunities for monitoring the Aichi Targets

#### 2.1 Overview

The field of remote sensing is a discipline in fast and constant evolution, with an increasing number of operational EO products that could be used for biodiversity monitoring. The choice of product can be daunting, as it is difficult to keep up-to-date with the latest developments and improvements in different areas. Nonetheless, the choice of product is in first instance determined by what is to be monitored. A detailed summary of operational EO products according to their applications in biodiversity monitoring and their potential to support the Convention can be found in Annex 2.

Most of the work done to date to use remotely-sensed data for biodiversity monitoring has been focused on the status and trends of selected habitats and species, and on ecosystem integrity, through the use of land cover and land use information. However, research is continuously evolving and opening new possibilities. A summary of emerging applications of remote sensing for both marine and terrestrial environments relevant for tracking progress towards the Aichi Biodiversity Targets can be found in Annex 3.

In order to support Parties to monitor the Aichi Biodiversity Targets this review analyses the potential use of remote sensing per Target. On the following pages a series of factsheets can be found, in which operational EO products have been mapped against each target and its operational indicators<sup>1</sup>. Only operational indicators that can be supported by an EO-based approach are listed. These are suggested products only and end users are encouraged to explore the strengths and weakness of the operational EO products and select those which might be best suited to develop a particular indicator in their own context.

A summary of the possibilities currently offered by remote sensing that can be applied by Parties is offered as well as an overview of current limitations. In addition, upcoming EO applications that could be used by Parties in the near future are discussed.

A traffic light system has been adopted to assess the adequacy of remotely-sensed data to monitor progress towards each of the Aichi Biodiversity Targets. As table 2.1 shows, this varies greatly. Potential applications for Strategic Goal A and E are limited, opportunities to contribution to Strategic Goal B and C have already proven to be extensive, whilst recent developments hold promising options for Strategic Goal D.

<sup>&</sup>lt;sup>1</sup> The Ad Hoc Technical Expert Group on Indicators for the Strategic Plan for Biodiversity 2011-2020 identified three categories of operational indicators. Indicators which are ready for use at the global level are denoted by the letter (A). Indicators which could be used at the global level but which require further development to be ready for use are denoted by the letter (B). Additional indicators for consideration for use at the national or other sub-global level are denoted by the letter (C) and given in italics. The set of (A) and (B) indicators are those which should be used to assess progress at the global level, while the (C) indicators are illustrative of some of the additional indicators available to Parties to use at the national level, according to their national priorities and circumstances.

Table 2.1 Mapping of the current adequacy of remote sensing to support tracking progress towards the Aichi Biodiversity Targets. Currently not observable by EO-based approach but maybe technically feasible in the future; Could be partially derived from EO-based information or EO-based approaches currently in development; Can be totally or partially derived from existing EO-based information

Strategic Goal	Aichi Biodiversity Target	Current remote sensing adequacy
A	1. Awareness of biodiversity values	•
	2. Integration of biodiversity values	•
	3.Incentives	•
	4.Sustainable production and consumption	•
В	5. Habitat loss, fragmentation and degradation	•
	<ol> <li>Sustainable explotation of marine resources</li> </ol>	•
	7.Biodiversity-friendly agriculture, forestry and aquaculture	•
	8. Pollution reduction	•
	9.Control of invasive alien species	0
	10. Coral reefs and other vulnerable ecosystems	•
С	11. Protected areas	
	12. Prevented extinction of threatened species	•
	13. Genetic diversity of socio- economically and culturally valuable species	•
D	14. Ecosystem services	
U	15. Ecosystem resilience	$\bigcirc$
	16. Access and benefit sharing	•
E	17. NBSAPs	
	18. Traditional Knowledge and customary	
	use	-
	19. Biodiversity knowledge improvement and transfer	
	20. Resource mobilisation	•

In addition to the summary factsheets a range of more detailed information can be found in Annex 4. A more detailed mapping of Aichi Biodiversity Targets that operational EO products mentioned in this review could support, with a summary of key features and various available datasets, can be found in Annex 4, Table 4.3. In addition, an in-depth mapping of each of the 98 indicators included in the indicative list of indicators, providing information on spatial and temporal resolution suitable for global, regional and national levels, type of data and appropriate sensors required to develop the indicator can be found through tables 4.4A-E, also in Annex 4. It should be noted this mapping does not mean to be absolute. It should be regarded as a guideline, and is subject to review and refinement. To complement these, a description of existing remote sensing sensors characteristics and their potential use for each Aichi Biodiversity Target can be found in Table 4.5.

## 2.2 Target by target assessment

Target	<ol> <li>Awareness of biodiversity values</li> <li>By 2020, at the latest, people are aware of the values of biodiversity and the steps they can take to conserve and use it sustainably.</li> </ol>
	Currently not measurable by an EO-based approach
Operational Indicators that can be (partly) derived from remotely- sensed data	None
Limitations	While it is expected that awareness leads to positive gains for biodiversity including measurable environmental factors such as reforestation, sustainable agriculture, increased fish stocks, restored habitats and the preservation of species diversity, there is no way to directly correlate human awareness with a change in environmental conditions using remote sensing.

Target	2. Integration of biodiversity values By 2020, at the latest, biodiversity values have been integrated into national and local development and poverty reduction strategies and planning processes and are being incorporated into national accounting, as appropriate, and reporting systems.
	Currently not measurable by an EO-based approach
Operational Indicators that can be (partly) derived from remotely- sensed data	None
Limitations	Green infrastructure such as ecological networks, forest corridors, viaducts, natural water flows and other realisations of the integration and implementation of biodiversity values into spatial planning are potentially possible to measure with remote sensing, if they are represented by visible features on the surface of the Earth. Whilst monitoring these might inform national accounting, it says little about actual integration into accounting, planning and development strategies.

Target	3. Incentives By 2020, at the latest, incentives, including subsidies, harmful to biodiversity are eliminated, phased out or reformed in order to minimize or avoid negative impacts, and positive incentives for the conservation and sustainable use of biodiversity are developed and applied, consistent and in harmony with the Convention and other relevant international obligations, taking into account national socio economic conditions.
	Currently not measurable by an EO-based approach
Operational Indicators that can be (partly) derived from remotely- sensed data	None
Limitations	Although the impacts of subsidy reform (for example on land cover and ecological condition) may be partly assessed via remote sensing, subsidy reform cannot be directly measured with remotely sensed data.

Target	4. Sustainable consumption and production
<b>Q</b>	By 2020, at the latest, Governments, business and stakeholders at all levels have taken steps to achieve or have implemented plans for sustainable production and consumption and have kept the impacts of use of natural resources well within safe ecological limits.
•	EO-based products can contribute to this Target but must be combined with other sources of data for a more comprehensive overview
Operational Indicators that can be (partly) derived from remotely- sensed data	<ol> <li>Trends in population and extinction risk of utilized species, including species in trade (A)</li> <li>Trends in ecological footprint and/or related concepts (C)</li> <li>Ecological limits assessed in terms of sustainable production and consumption (C)</li> <li>Trends in biodiversity of cities (C)</li> </ol>
Relevant Operational EO products	Landcover, NDVI
Current EO- based approaches	Carbon parameters are one of the newest remote sensing metrics for monitoring sustainable production within ecological limits. Archived data levels of atmospheric concentrations of carbon dioxide and other greenhouse gas emissions (GHGs) have been acquired through ground-based methods, dating from the Ice Age to the Industrial Revolution to present day. These estimates can be combined with modern satellite measurements of carbon emissions to the atmosphere, terrestrial and marine carbon stocks and other GHGs such as methane to assess their likely impact on global climate. Carbon and GHG emissions can also be combined with other remotely-sensed data products, such as landcover, vegetation indices, crop yields and habitat degradation for a variety of research applications including identifying and measuring sustainable agriculture and forestry (indicator 12 and 13). Agricultural monitoring has long been conducted with EO-based terrestrial vegetation products in order to estimate crop yields. The JRC MARS initiative generates monthly bulletins on crop growth conditions across Europe using near-real time forecasting methods which use satellite remotely-sensed information combined with meteorological forecasts (indicator 13). However linking such agro-meteorological information and other resource production information with biodiversity conservation presents a new twist on this application.
Limitations	Even amongst the existing sensors (GOSAT, Terra/Aqua and SeaWiFS) not all data products are currently available. With the exception of Terra and Aqua's MODIS instrument, many of the carbon measuring sensors focus on atmospheric monitoring rather than Earth observation. Therefore, their utility for helping to evaluate sustainable landuse in relation to biodiversity protection is yet to be proven. Indicators 11 and 14 are currently limited by the temporal, spatial and spectral resolution of current operational EO-based products.
Upcoming EO-based approaches	At least one new sensor focused on obtaining carbon transmission and related vegetation parameters is scheduled for launch in 2014 (e.g. Orbiting Carbon Observatory) and one experimental vegetation-specific sensor was launched in 2013 (Proba-V).

Target	<ol> <li>Habitat loss fragmentation and degradation</li> <li>By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced.</li> </ol>
•	EO-based information can make a significant contributions to monitoring this Target and is already widely in use in assessing changes in forest cover
<b>Operational</b> <b>Indicators</b> that can be (partly) derived from remotely- sensed data	<ul> <li>17. Trends in extent of selected biomes, ecosystems and habitats (A)</li> <li>18. Trends in proportion of degraded/threatened habitats (B)</li> <li>19. Trends in fragmentation of natural habitats (B)</li> <li>20. Trends in condition and vulnerability of ecosystems (C)</li> <li>21. Trends in the proportion of natural habitats converted (C)</li> <li>22. Trends in primary productivity (C)</li> <li>23. Trends in proportion of land affected by desertification (C)</li> <li>24. Population trends of habitat dependent species in each major habitat type (A)</li> </ul>
Relevant Operational EO products	Land cover is useful for terrestrial habitat loss and fragmentation while NDVI, LAI and FAPAR are potentially useful in assessing vegetation condition hence habitat degradation.Fire represents a major habitat disturbance. Global Forest Watch (GFW) 2.0 of the World Resources Institute is a near-real time deforestation monitoring tool incorporating Landsat satellite imagery and ground-based observations. Marine products such as ocean chlorophyll-a concentration, ocean primary productivity, suspended sediment, sea surface wind speed, sea surface temperature, sea surface salinity and sea surface state define the physical and biological state of the marine environment. Synthesising these products offers the potential to assess the overall condition of marine habitats and identify where degradation is occuring.
Current EO- based approach	Using remote sensing to monitor habitats is routinely performed in terrestrial environments (Lengyel <i>et al.</i> , 2008), and habitat distribution represents one of the most common pieces of information reported by Parties to the CBD. Optical sensors are the primary choices for landuse and landcover mapping as surrogates for habitat because the optical sensors products are most widely available and easy to use. Radar and thermal imagery are technically more advanced requiring specialist knowledge. Medium resolution imagery such as Landsat, SPOT, ASTER and IRS are often sufficient for the purpose of habitat mapping over large areas, even in complex fine- scale habitat mosaics (Lucas <i>et al.</i> , 2011). Recent very high resolution (VHR) satellites such as WorldView-2 are beginning to open up the possibility of combining high spatial and spectral resolution in the same platform (Nagendra and Rocchini, 2008).
Limitations	Satellite imagery which provides sufficient spatial ( $\leq$ 30m) and spectral ( $\geq$ 3 spectral bands) information to retrieve habitat-type information, e.g. Landsat Thematic Mapper (TM)/Earth Thematic Mapper (ETM) +, is characterised by small spatial footprints per scene, requiring numerous scenes and high data processing capacity when used for habitat monitoring at broad scales. This makes it difficult to monitor global habitat comprehensively and seamlessly for Target 5. Indeed, there is currently no long-term landcover change product which provides habitat

	information down to a level which is ecologically meaningful and global in application. This is partly due to the constraint on computing resources and expertise to conduct such a comprehensive and large-scale analysis. Validating such a product would also present considerable challenges. VHR satellite, airborne or unmanned aerial vehicle (UAV)-based imagery are frequently mentioned as being the ideal option for fine scale mapping of habitats with high spatial heterogeneity (Szantoi <i>et al.</i> ,2013). However, while such data are useful for habitat mapping at the local scale, imagery can suffer from problems of shadowing, cloud cover and high heterogeneity especially in complex landscapes. VHR can also be expensive and time consuming to procure and process.
	Airborne sensors such as the NASA Airborne Visible/Infrared Imaging Spectrometer (AVIRIS) provide hyperspectral imagery with the potential to improve monitoring of habitats by identifying the percentage coverage per image pixel of certain species of plants at a local scale. This affords fine-scale successional change and species diversity information. However, AVIRIS-derived products, like most hyperspectral imagery, are not operationally-produced, expensive to procure and are technically challenging and labour-intensive to process. The different intra- and international definitions of various types of habitats such as 'Forest', 'Wetland' and 'Marine' environments also pose a limitation to monitor habitats which affects any efforts to use remote sensing to track progress toward achieving Target 5 (GEO BON, 2011). This inconsistency of definitions may undermine the effectiveness of the monitoring of the extent of ecological regions, habitat loss, fragmentation and degradation.
	Key gaps in data on habitat extent, fragmentation and degradation include: the condition of temperate coastal marine habitats, offshore marine breeding and spawning grounds, kelp forests, intertidal and sub-tidal ecosystems, vulnerable shelf habitats, seamounts, hot-and cold seeps, ocean surface, benthic and deep sea habitats; inland wetland and non-forested terrestrial habitats and polar habitats. Better information is also needed on small-scale habitat degradation in all habitats (GEO BON, 2011).
Upcoming EO-based products	Active remote sensing through Synthetic Aperture Radar (SAR) and Light Detection and Ranging also holds great potential for the mapping and identification of structurally complex habitats, especially in tropical areas where there is high and/or frequent cloud cover.

Target	6. Sustainable exploitation of marine resources By 2020 all fish and invertebrate stocks and aquatic plants are managed and harvested sustainably, legally and applying ecosystem based approaches, so that overfishing is avoided, recovery plans and measures are in place for all depleted species, fisheries have no significant adverse impacts on threatened species and vulnerable ecosystems and the impacts of fisheries on stocks, species and ecosystems are within safe ecological limits.
•	EO-based products can contribute to this Target but must be combined with other sources of data for a more comprehensive overview. Economic information on fisheries would be particularly beneficial in this regard.
Operational Indicators that can be (partly) derived from remotely- sensed data	26. Trends in population of target and bycatch aquatic species (A) 29. Trends in fishing effort capacity (C)
Relevant Operational EO products	ocean chlorophyll-a concentration, ocean primary productivity, suspended sediment, sea surface wind speed, sea surface temperature, sea surface salinity and sea surface state
Current EO- based approach	Optical and radar sensors can be used to detect marine vessels and monitor vessel movement for tracking illegal fishing (Corbane <i>et al.</i> , 2010) (indicator 29). As with terrestrial species, direct observation with satellite remote sensing is not usually possible. In place of direct monitoring, biological and physical parameters that determine species ranges can be derived from remotely-sensed data. Kachelreiss <i>et al</i> (2013) noted that in the marine environment, primary productivity has been linked with benthic community patterns. They also review the range of current EO-based oceanographic products which structure marine biodiversity, e.g. sea surface temperature.
Limitations	Most remote sensing methods can only derive information from the upper layer of the ocean. Space-borne optical sensors are naturally limited at shallow ocean depths (up to 27 meters) due to the light absorption properties of sea water (Rohmann and Monaco 2005, as cited in Kachelreiss <i>et al.</i> , 2013. The best available sensors at airbone ranges (i.e.LiDAR) can potentially only reach up to depths of 70 meters, but more commonly penetrate in a range from 35-50m (McNair 2010, Guildford and Palmer 2008). This focus on shallow water monitoring impedes the monitoring of many marine species, with the exception of some marine mammals and phytoplankton.

Target	7. Biodiversity-friendly agriculture, forestry and aquaculture
7	By 2020 areas under agriculture, aquaculture and forestry are managed sustainably, ensuring conservation of biodiversity
•	EO-based products can contribute to this Target but must be combined with other sources of data for a comprehensive overview of status of the Target. Land use information would be particularly beneficial in this regard.
<b>Operatioanl</b> <b>Indicators</b> that can be (partly) derived from remotely- sensed data	<ul> <li>32. Trends in population of forest and agriculture dependent species in production systems (B)</li> <li>33. Trends in production per input (B)</li> <li>34. Trends in proportion of products derived from sustainable sources (C)</li> <li>35. Trends in area of forest, agricultural and aquaculture ecosystems under sustainable management (B)</li> </ul>
Relevant Operational EO products	Land cover
Current EO- based approach	Land use change is the premier driver of biodiversity loss in terrestrial habitats that can be measured by remote sensing. While there are a plethora of studies that show how remote sensing can be used to map land cover, monitor habitat and predict species distribution and species richness there are no studies that link agriculture to biodiversity through remote sensing, in an attempt to ascertain if the practices are 'biodiversity-friendly'. However, using existing land cover mapping methods, it should be feasible to combine a land cover map with non-EO spatial data layers on land use,e.g. on the type of agriculture, forestry and aquaculture being practiced, to create a 'biodiversity-friendly' land use layer. Such hybrid approaches which incorporate EO-based landcover data with non-EO data on land use could be useful for this Target.
Limitations	More work is needed to identify and define sustainable agriculture, forest and aquaculture practices that enable biodiversity conservation. Indicators of 'biodiversity friendly' practices will need to be identified and the feasibility to measure those indicators by remote sensing either directly or indirectly will need to be ascertained. For example, it would be useful to determine the species mixes of agricultural and forestry plots in order to estimate whether they are likely to support high biodiversity. Monocultures for example can feasibly be mapped by EO since they are homogenous in composition and should have a consistent spectral signature but are unlikely to be biodiversity friendly. Aquaculture may be more challenging however since the spectral information alone may not be sufficient to characterise aquaculture from spaceborne sensors. The availability of appropriate spatial datasets on land use may also be an issue since these are not routinely available or operationally produced at a global level. The availability of such datasets is likely to be intermittent and highly dependent on national scale needs. However, a land use dataset is essential for an EO-based approach to measuring and monitoring this target so may only be feasible for countries or regions with a well developed spatial data infrastructure.

Target	8. Pollution Reduction
anget 8	By 2020, pollution, including from excess nutrients, has been brought to levels that are not detrimental to ecosystem function and biodiversity.
•	EO-based products can contribute to this Target but must be combined with other sources of data for a comprehensive overview of status of the Target. Information on sources and sinks of pollutants would be particularly beneficial in this regard.
Operational Indicators that can be (partly) derived from remotely- sensed data	<ul> <li>36. Trends in incidence of hypoxic zones and algal blooms (A)</li> <li>37. Trends in water quality in aquatic ecosystems (A)</li> <li>39. Trends in pollution deposition rate (B)</li> <li>41. Trend in emission to the environment of pollutants relevant for biodiversity (C)</li> <li>44. Trends in ozone levels in natural ecosystems (C)</li> <li>46. Trends in UV-radiation levels (C)</li> </ul>
Relevant Operational EO products	ocean chlorophyll-a concentration, main greenhouse gases, i.e., ozone, methane, nitrous oxide and carbon monoxide, atmospheric $NO_2$
Current EO- based approach	<ul> <li>Atmospheric monitoring of haze, smoke and smog occupy a large proportion of remote sensing studies on pollution monitoring. However remote sensing for tracking aerosols, ozone and GHGs is less well-developed.</li> <li>Land use in the form of agriculture and development can have negative effects on marine biodiversity due to run-offs (Boersma and Parrish, 1999). The main parameters for monitoring pollution in coastal waters include suspended particulate matter (SPM) and coloured dissolved organic matter (CDOM). SPM, like many biophysical parameters available from remote sensing serves only as an indicator for land-based pollutants that cannot be detected by remote sensing (Kachelriess <i>et al.</i>, 2013). SPM and CDOM can also be inferred from ocean colour data but only when ground calibration data is available (Oney <i>et al.</i>, 2011).</li> <li>Remote sensing based methods have been critical in tracking oil spills through the use of synthetic aperture radar (SAR) or infrared sensors which can 'see' through clouds and hyperspectral data which are very good at discriminating hydrocarbons and minerals.</li> <li>Kachelriess <i>et al.</i> (2013) also notes the exceptional Hyperion hyperspectral sensor on board the EO-1 satellite as an answer to the limits of airborne options. Hyperion may be of use in long-term, broad scale pollution monitoring. However Hyperion is also limited by its modest 30 meter resolution and 16 day revisit period and therefore may not be of use in emergency situations where constant monitoring is desired.</li> </ul>
Limitations	The downside of hyperspectral sensors is that they require complex processing and computing capacity, are mostly only commercially available and therefore costly to procure and process. Radar-based oil-spill detection is also a complex task requiring specialist software and numerical skills and integrated observing systems.

Target	9. Control of invasive alien species
29	By 2020, invasive alien species and pathways are identified and prioritized, priority species are controlled or eradicated, and measures are in place to manage pathways to prevent their introduction and establishment.
•	EO-based information can make a significant contribution to monitoring this Target but solely to track invasive plant species. Invasive terrestrial and aquatic animals may not be easily detectable by remote sensing.
Operational Indicators that can be (partly) derived from remotely- sensed data	<ul> <li>47. Trends in the impact of invasive alien species on extinction risk trends (A)</li> <li>48. Trends in the economic impacts of selected invasive alien species (B)</li> <li>49. Trends in number of invasive alien species (B)</li> <li>52. Trends in invasive alien species pathways management (C)</li> </ul>
Relevant Operational EO products	No current operational products can be used directly but measures of disturbance can help indirectly, e.g. fire, land cover change, as these events open up pathways for invasive plant species to enter into previously intact habitats.
Current EO- based approach	With relation to invasive species, remotely-sensed datasets must always be used in conjunction with modelling and field information to predict changes in specific species of interest (e.g. Asner and Martin, 2009; He <i>et al.</i> , 2011; Nagendra <i>et al.</i> 2013). EO-based products are particularly useful when mapping physical pathways in the landscale for invasive plant species which frequently occur along disturbance routes, e.g. roads and other infrastructure in forests or drainage channels in wetlands. Standard multispectral remote sensing (e.g. Landsat) was found to be useful in this regard when combined with orthophotos (Somadi <i>et al.</i> 2012). Airborne hyperspectral imagery has been found to be useful on a number of occasions, especially when timing the acquisition of high precision spectroscopy data with critical phenological stages of flowering or leaf senescence (He <i>et al.</i> , 2011; Andrew and Ustin, 2008; Lucas <i>et al.</i> , 2008, Clark <i>et al.</i> , 2005; Ramsey <i>et al.</i> 2005). Employing measures of image texture with NDVI, derived from sub-metre resolution imagery, can greatly improve classification accuracy and overall ability to track invasive species in wetlands (Szantoi <i>et al.</i> , 2013).
Limitations	Intra-species variation, mixed pixels due to high levels of heterogeneity and shadowing in the image have been found to minimize success when using hyperspectral imagery. Accurate discrimination of all top-canopy species is therefore unlikely, particularly in high density forests where there is a substantial amount of overlap between leaves and branches of different species. This problem is unlikely to disappear even if hyperspectral image resolution and noise to signal ratios improve significantly in the future (Nagendra, 2001; Fuller, 2007). Very High Resolution imagery (e.g. Quickbird, IKONOS, GeoEye) has been found to be unsuitable for invasive species identification and monitoring because of the very small pixel sizes and lack of a short-wave infrared band, increasing the variability between different tree canopies (Nagendra 2013; Fuller 2005) in the scene.

Target	10. Coral reefs and other vulnerable ecosystems
	By 2015, the multiple anthropogenic pressures on coral reefs, and other vulnerable ecosystems impacted by climate change or ocean acidification are minimized, so as to maintain their integrity and functioning.
•	EO-based products can contribute to this Target but are mostly limited to shallow- water environments and site-specific studies
Operational Indicators that can be (partly) derived from remotely- sensed data	<ul> <li>53. Extinction risk trends of coral and reefs fish (A)</li> <li>54. Trends in climate change impacts on extinction risk (B)</li> <li>55. Trends in coral reef condition (B)</li> <li>56. Trends in extent, and rate of shifts of boundaries, of vulnerable ecosystems (B)</li> </ul>
Relevant Operational EO products	No current operational products can be used directly but measures of habitat condition can contribute indirectly, e.g. Sea Surface Temperature, VCI/VPI
Current EO- based approach	Large-scale coral mortality events known as coral bleaching have been successfully studied using remote sensing, as the occurrence of these events is found to be strongly correlated to Sea Surface Temperature (SST) (Maynard, 2008; Sheppardand Rayner, 2002). However the correlation between SST and bleaching varies by species owning to different mortality thresholds influenced by a variety of factors and therefore, global prediction of coral bleaching for a given SST anomaly is not always a consistent or straightforward measurement (Maynard, 2008). Kachelriess <i>et al.</i> (2013) recommended that when it comes to monitoring coral bleaching, SST should only be used as an indicator for threats, and not as a way to quantify bleaching. All of these studies emphasised the need for validation of remotely-sensed data with field surveys.
Limitations	The limitations of monitoring marine habitats and species due to shallow depth penetration of spaceborne (27 meters) and airborne sensors (47 meters) was discussed in Target 6 but is also relevant for Target 10 as it affects the ability to monitor coral reefs and other potentially vulnerable marine ecosystems in deeper waters (Kachelriess <i>et al.</i> , 2013). However monitoring coral reefs also suffers from the limited availability of high spatial resolution data. <i>In-situ</i> management often requires stratified sub-meter resolution to be useful. The best solution for bathymetric mapping and under-water habitat classification are proving to be those provided by LiDAR with its pin-point precision and high resolution; however even LiDAR falls short of capturing the complexity of coral reefs and other complex habitats (Kachelriess <i>et al.</i> 2013; Purkis and Klemas 2011). This means that for the foreseeable future, mapping individual colonies or reefs will remain unfeasible with airborne or satellite remote sensing. However, cold water, deep-sea coral reefs have been successfully mapped using side-scan sonar and underwater remotely operated vehicles (Dorschel <i>et al.</i> 2009). Airborne and spaceborne sensors are more appropriate for marine habitat mapping in pelagic ecosystems which are influenced by broader oceanographic patterns and can therefore be monitored synoptically. In terms of spectral resolution, it is difficult to discriminate between species of coral without hyperspectral sensors (Klemas, 2011;Purkis and Klemas, 2011; Wingfield <i>et al.</i> , 2011) but as previously indicated, the majority of hyperspectral data options are not freely available and require a great deal of skill and resource to utilise.

Target	11. Protected areas
11	By 2020, at least 17 per cent of terrestrial and inland water, and 10 per cent of coastal and marine areas, especially areas of particular importance for biodiversity and ecosystem services, are conserved through effectively and equitably managed, ecologically representative and well connected systems of protected areas and other effective area-based conservation measures, and integrated into the wider landscapes and seascapes.
•	EO-based information can make a significant contribution to monitoring this Target when combined with non-EO data on protected area distribution and can be complemented by field-based information to assess protected area effectiveness
<b>Operational</b> <b>Indicators</b> that can be (partly) derived from remotely- sensed data	<ul> <li>59. Trends in coverage of protected areas (A)</li> <li>60. Trends in extent of marine protected areas, coverage of key biodiversity areas and management effectiveness (A)</li> <li>61. Trends in protected area condition and / or management effectiveness including more equitable management (A)</li> <li>62. Trends in representative coverage of protected areas and other area based approaches, including sites of particular importance for biodiversity, and of terrestrial, marine and inland water systems (A)</li> <li>63. Trends in the connectivity of protected areas and other area based approaches integrated into landscapes and seascapes (B)</li> <li>64. Trends in the delivery of ecosystem services and equitable benefits from protected areas (C)</li> </ul>
Relevant Operational EO products	Land cover and land cover change, NDVI, NDVI-derived anomaliles such as the Vegetation Condition Index or the Vegetation Productivity Index, LAI, FAPAR, fire
Current EO- based approach	Hyperspectral, hyperspatial, optical, radar and LiDAR remote sensing can all be beneficial to monitoring biodiversity within and around protected areas. Informatics tools such as the JRC Digital Observatory for Protected Areas (DOPA) deliver up to date EO-based information on protected areas via web-based technologies.
Limitations	Remotely-sensed habitat change is not always a suitable indicator of protected area effectiveness (Geldmann <i>et al.</i> , 2013). More subtle variation in habitat condition, such as reduction in forest megafauna, cannot be inferred from remotely-sensed measures of deforestation (Redford, 1992). This problem is compounded by the fact that not all forest dwellers are correlated with the area of forest cover (Wilkie <i>et al.</i> , 2011). Therefore estimating deforestation by remote sensing alone may not give a realistic interpretation of habitat condition and thus protected area effectiveness. For a realistic implementation of remote sensing to support PA management, financial and human resources will need to be taken into account.

Target	12. Prevent extinction of threatened species
	By 2020 the extinction of known threatened species has been prevented and their conservation status, particularly of those most in decline, has been improved and sustained.
•	EO-based information can make a significant contributing to monitoring this Target but only for certain species and in specific habitats. Ground observations of species could be particularly beneficial when combined with EO-based information on habitat status.
Operational Indicators that can be (partly) derived from remotely- sensed data	<ul><li>65. Trends in abundance of selected species (A)</li><li>66. Trends in extinction risk of species (A)</li><li>67. Trends in distribution of selected species (B)</li></ul>
Relevant Operational EO products	Operational parameters such as NDVI, FAPAR, LAI can potentially be used to characterise the vegetation state hence habitat condition of threatened terrestrial species of animals and plants Landcover can be used in modelling habitat suitability and landcover change in assessing whether a habitat of known threatened species is in danger from land use change
Current EO- based approach	It is important to keep in mind that in relation to monitoring species, the direct observation of individual species is usually not possible using remotely-sensed information, with exceptions only among mega-fauna where the animals or their habitats can be easily detected. Examples where this kind of monitoring has been successful include blue shark (Queiroz <i>et al.</i> , 2012); bluefin tuna (Druon, 2010); whale sharks (Sequeira <i>et al.</i> , 2012); seabirds (Petersen <i>et al.</i> , 2008), elephants, wildebeest and zebra (Yang, 2012); marmots (Velasco 2009), and penguins (Fretwell <i>et al.</i> , 2012). Nonetheless, biophysical parameters that are reported to structure biodiversity patterns can be derived from remotely-sensed data.
Limitations	The challenge of mapping individual species or species richness is also variable across ecological regions. In tropical forests where there is high taxonomic diversity within plant functional groups, optical remote sensing is met with many challenges. Atmospheric influences and a wide variety of determinants of spectral variation such as sun angle, camera viewing angle, topography, and canopy three-dimensional structure persist (Kennedy <i>et al.</i> 1997; Sandmeier <i>et al.</i> 1998; Diner <i>et al.</i> 1999). Though there are ongoing studies and technological advances to overcome these challenges they have yet to come to fruition. Very high-precision, plant canopy-level measurements of foliar chemistry are produced from airborne High-Fidelity Imaging Spectrometers (HiFIS) at spatial resolutions that can resolve individual tree crownsbut represent only the firststeps toward species-level measurements (Asner and Martin 2009). LiDAR also needs to progress in the usability of its intensity data – a concentrated measure of spectral reflectance. Intensity is an opportunistic by-product of LiDAR, secondary in importance
	to height and location data but has nevertheless been the focus of many new tree species differentiation studies. Utilising intensity successfully still requires

	sophisticated post-capture calibration algorithms due to a lack of sensor calibration. Additionally airborne data capture is still prohibitively expensive. For these reasons airborne remote sensing, especially that of HiFIS and LiDAR are an impossibility for many practical monitoring procedures.
Upcoming EO- approaches	Asner and Martin (2009) suggest that there is a sufficient theoretical basis to characterise the taxonomic diversity of tropical tree species from airborne LiDAR my measuring their foliar chemistry and structural characteristics in a way that is generic and scalable. For example, HiFIS data, which can measure a range of plant chemicals, are thought to be linked with species diversity. However, rarely has the biochemical information, which seemingly sets HiFIS apart from other airborne optical sensors, been used to estimate the taxonomic composition of plant canopies. This is primarily due to the interference caused by the aforementioned factors having little to do with canopy chemistry but a lot to do with other determinants of spectral variation. In their 2009 study, Asner and Martin promote using a combination of HiFIS and LiDAR which can precisely measure canopy height and structure in 3D in a new form of remote sensing called "spectranomics". However, this fusion of technology is as yet untested and will at first be costly to pull-together but could be applied to the study of threatened tree species and potentially the primates and mammals which dpend on them to survive.

Target	13. Genetic diversity of socio-economic and culturally valuable species
13	By 2020, the genetic diversity of cultivated plants and farmed and domesticated animals and of wild relatives, including other socio- economically as well as culturally valuable species, is maintained, and strategies have been developed and implemented for minimizing genetic erosion and safeguarding their genetic diversity.
	Currently not directly measurable by an EO-based approach
Operational Indicators that can be (partly) derived from remotely- sensed data	None
Relevant Operational EO products	Genetic material contained in an individual animal or plant can not be measured directly by remote sensing, based methods or current operational EO products. However, EO-based methods of monitoring populations of species directly, e.g. by counting individuals or estimating their coverage, could potentially contribute to this Target. Monitoring isolated populations of the same species over time could be used to assess the level of exchange of genetic material and whether genetic diversity is being safeguarded. The benefit of an EO-based approach is the ability to measure the spatial distribution of different populations over large areas using image interpretation techniques. The extent to which these populations mix could be reasonably estimated in this way.
Limitations	In order to understand the exchange of genetic material between isolated populations long time series are needed, spanning several decades ideally while remotely-sensed imagery has only been available for the last few decades at most,
Upcoming EO- approaches	Studies have incorporated EO-based information on contemporary species ranges with their modelled distributions in the past to assess how genetic changes have occurred over time among isolated populations of species. This is largely an experimental application of EO data to map spatial variation in genetic diversity.

Target	<ul> <li>14. Ecosystems and essential services safeguarded</li> <li>By 2020, ecosystems that provide essential services, including services related to water, and contribute to health, livelihoods and well-being, are restored and safeguarded, taking into account the needs of women, indigenous and local communities, and the poor and vulnerable</li> </ul>
•	EO-based information can make a significant contribution to monitoring this Target by providing inputs to ecosystem service models
<b>Operational</b> <b>Indicators</b> that can be (partly) derived from remotely- sensed data	<ul> <li>73. Trends in benefits that humans derive from selected ecosystem services (A)</li> <li>75. Trends in delivery of multiple ecosystem services (B)</li> <li>76. Trends in economic and non-economic values of selected ecosystem services (B)</li> <li>78. Trends in human and economic losses due to water or natural resource related disasters (B)</li> <li>79. Trends in nutritional contribution of biodiversity: Food composition (B)</li> <li>80. Trends in incidence of emerging zoonotic diseases (C)</li> <li>81. Trends in nutritional contribution of biodiversity: Food consumption (C)</li> <li>84. Trends in natural resource conflicts (C)</li> <li>85. Trends in the condition of selected ecosystem services (C)</li> <li>87. Trends in area of degraded ecosystems restored or being restored (B)</li> </ul>
Relevant Operational EO products	Precipitation, water bodies, carbon /biomass, landcover
Current EO- based approach	Ecosystems provide ecological functions that directly or indirectly translate to a variety of beneficial contributions to society, referred to as ecosystem services. The capacity of an ecosystem to deliver services depends on the status of the biodiversity it harbours. Habitat mapping is key to assess the health of a particular ecosystem and habitats in favourable conservation status tend to supply more and better ecosystem services.
	Monitoring of vulnerable ecosystems, such as coral reefs, using remote sensing is limited due to the limited availability of high spatial resolution data. The longest running, most widely tested remote sensing products, such as that available from the Landsat and AVHRR series are at best limited to ecosystem monitoring capacity, where landcover can be used as a surrogate for ecosystems and must be combined with other data to say something about ecosystem services. Without clearly defined indicators of ecosystem services and maps of ecosystem services in relation to identified beneficiaries, measuring progress toward Target 14 will be constrained.
	<ul> <li>Carbon and water-based ecosystem services are the most readily observable by EO-based technologies. These include:</li> <li>Above-ground woody carbon terrestrial biomass measurements derived from a combination of field measurements, LiDAR and MODIS imagery. A number of authors have used this method to estimate regional and global biomass while publishing biomass carbon datasets (Baccini <i>et al.</i> 2008; Baccini <i>et al.</i> 2011; Ruesch and Gibbs 2008; Saatchi <i>et al.</i> 2007; Saatchi <i>et al.</i> 2011)</li> </ul>

	<ul> <li>Models of water-based ecosystem services         <ul> <li>Precipitation inputs can be derived from the NASA/JAXA Tropical Rainfall Measuring Mission (TRMM)</li> <li>Land surface temperature data derived from satellite sensors such as Landsat, AVHRR, MODIS and ASTER</li> <li>Groundwater provision can be measured indirectly from temporal variation in Earth's gravity field as measured by the Gravity Recovery and Climate Experiment (GRACE) mission</li> <li>Landcover and/or vegetation cover, e.g. VCF, is central to ecosystem models</li> </ul> </li> <li>Global mapping of carbon, stored in terrestrial vegetation, is not straightforward as</li> </ul>
Limitations	datasets from remotely-sensed and ground-based sources are frequently amalgamated with different approaches taken to integrating these various datasets in order to produce seamless carbon (biomass) maps. As a result, a comparision of published datasets shows that there are major differences, not only in terms of the estimates for quantity of biomass (carbon), but also in terms of the distribution pattern of carbon they provide. For example, the Baccini <i>et al.</i> (2012) dataset has higher above-ground biomass values than the Saatchi <i>et al</i> (2011) datasets in both African and the Amazonian rainforests, whereas in the Guyana shield and in west- Central Africa (Cameroon/Gabon), the above-ground biomass values in the Saatchi <i>et al</i> (2011) datasets are higher. Minor geographic discrepancies exist elsewhere for tropical regions.

Target	15. Ecosystem resilience
	By 2020, ecosystem resilience and the contribution of biodiversity to carbon stocks has been enhanced, through conservation and restoration, including restoration of at least 15 per cent of degraded ecosystems, thereby contributing to climate change mitigation and adaptation and to combating desertification.
•	EO-based products can contribute to this Target but must be combined with other sources of data for a more comprehensive overview of progress towards achieving this Target.
Operational Indicators that can be (partly) derived from remotely- sensed data	<ul> <li>88. Status and trends in extent and condition of habitats that provide carbon storage (A)</li> <li>89. Population trends of forest-dependent species in forests under restoration (C)</li> </ul>
Relevant Operational EO products	Time series of NDVI and FAPAR, e.g. to derive measures of primary productivity and vegetation phenology which can be related to the rate and timing of carbon sequestration in terrestrial vegetation. Land cover and land cover change can be used to assess conservation and restoration of habitats, especially those of high carbon stock such as mangroves and tropical forest, if applied at an appropriate scale. EO-based carbon estimates are essential in this regard but are not operationally produced or globally available.
Current EO- based approach	Remotely-sensed information on the parameters required for measuring progress toward target 15, such as NDVI and FAPAR, are globally available but would be more appropriately derived over specific habitats, e.g. coastal habitats such as saltmarshes or mangroves or terrestrial habitats such as tropical forests or peatlands, as these are essential ecosystems for climate change mitigation as well as harbouring important biodiversity. Initiatives such as the ESA GlobWetland II and the WRI GFW 2.0 have recognised the importance of these ecosystems and promoted EO-based approaches to their conservation and management. However, regardless of the EO-based tool or product adopted for ecosystem monitoring, it would be prudent to use only those data products for which change detection analyses can be conducted to ascertain resilience to climate change. The timing of EO-based information is also important asutilising seasonal data timed with peak phenological and physiological changes can be useful for early identification of climate change impacts.
Limitations	Monitoring ecosystem resilience necessities multi-decadal time series of EO data which rules out many sensors except for Landsat and NOAA-AVHRR. Mission continuity must be assured by space agencies if consistent time series of EO data are to be maintained and usable for tracking progress towards this target. Current operational EO products which are typically ≥1km in spatial resolution are not appropriate for the ecosystem-level information that is required to monitor this target comprehensively.

Target	16. Access and benefit sharing (ABS) By 2015, the Nagoya Protocol on Access to Genetic Resources and the Fair and Equitable Sharing of Benefits Arising from their Utilization is in force and operational, consistent with national legislation.
Operational Indicators that can be (partly) derived from remotely- sensed data	Currently not measurable by an EO-based approach None

Target	17. National Biodiversity Strategies and Action Plans (NBSAPs) By 2015 each Party has developed, adopted as a policy instrument, and has commenced implementing an effective, participatory and updated national biodiversity strategy and action plan.
•	Currently not directly measurable by an EO-based approach but EO data can be utilised in NBSAP planning, e.g. for identifying priority habitats from land cover data or pressures from land cover change or pollution measures
Operational Indicators that can be (partly) derived from remotely- sensed data	None
Current EO- approach	Indirectly, the achievable monitoring of other Aichi Targets over time and within national contexts could potentially indicate whether a country is succeeding at implementing its NBSAPs.

Target	18. Traditional knowledge and customary use By 2020, the traditional knowledge, innovations and practices of indigenous and local communities relevant for the conservation and sustainable use of biodiversity, and their customary use of biological resources, are respected, subject to national legislation and relevant international obligations, and fully integrated and reflected in the implementation of the Convention with the full and effective participation of indigenous and local communities, at all relevant levels.
•	Currently not directly measurable by an EO-based approach. However, EO-based products could contribute to this Target if combined with other sources of data. Existing EO-based landcover information could enhance existing socio-economic information on land tenure and landuse for a more comprehensive overview of status of the Target.
Operational Indicators that can be (partly) derived from remotely- sensed data	None

Target	19. Biodiversity knowledge improvement and transfer By 2020, knowledge, the science base and technologies relating to biodiversity, its values, functioning, status and trends, and the consequences of its loss, are improved, widely shared and transferred, and applied.
	Currently not directly measurable by an EO-based approach. However if knowledge and technology in the use of remote sensing to monitor other measurable Aichi Targets is improved as suggested herein, it would contribute toward meeting this target.
<b>Operational</b> <b>Indicators</b> that can be (partly) derived from remotely- sensed data	None

Target	20. Resource mobilisation By 2020, at the latest, the mobilisation of financial resources for effectively implementing the Strategic Plan for Biodiversity 2011-2020 from all sources, and in accordance with the consolidated and agreed process in the Strategy for Resource Mobilisation, should increase substantially from the current levels. This target will be subject to changes contingent to resource needs assessments to be developed and reported by Parties.
	Currently not directly measurable by an EO-based approach
Operational Indicators that can be (partly) derived from remotely- sensed data	None

# 3. Lessons learnt from national experiences

Over the last years, countries have adopted different approaches to the use of remote sensing to monitor biodiversity at a national level, according to their particular needs, capacities and circumstances. The following case studies provide an insight into the application of different methods and products at national and subnational level, and their impact on decision-making and policy implementation. They also offer examples of how particular limitations and challenges have been overcome. The value of open access data, application in near real time monitoring of threats and inputs to strategic conservation planning are all illustrated, as are the resource and capacity constraints often faced by governments in attempting to utilize remotely sensed data to develop national data products and indicators.

## 3.1 Remote sensing as a surveillance tool: fire monitoring in Australia

Due to the low population base and large size of Australia's land-mass (7.5 million km<sup>2</sup>), remote sensing technologies have been used for wildfire ("bushfire") monitoring, fire-scar mapping and general environmental monitoring ever since the first earth observation satellites were launched in the 1970's. For Australia, satellite technologies have proven to be one of the most appropriate technologies for use in wide-area fire detection and tracking, as well as general environmental monitoring and fuel dryness monitoring.

In 2003, the CSIRO (Commonwealth Scientific and Industrial Research Organisation), together with the department of Defense and Geoscience Australia, developed the "Sentinel Hotspots" bushfire tracking system and associated webGIS portal, which used the Moderate Resolution Imaging Spectrometer Sensor (MODIS) onboard NASA's Aqua and Terra satellites. Through the use of these two satellites, a full continental coverage is achieved up to four times every 24 hours, at a spatial resolution of about 1 km, and a time-latency from satellite overpass to visualization of the hotspot location on the webGIS system of approximately 45 minutes, making this a suitable synoptic near real-time fire monitoring system. Today, the Sentinel system is housed at Geoscience Australia (http://sentinel.ga.gov.au/), and continues to be used on a 24/7 basis by federal and state fire management agencies, natural resource managers, ecologists and the general public as fire conditions develop across the country. Other state-based or regional systems such as "FireWatch" in Western Australia and the NAFI (Northern Australia Fire Information) system in the Northern Territory, use similar approaches.

This operational concept was also adopted in 2006 by the Asia Pacific Regional Space Agencies Forum (APRSAF), as it established the "Sentinel Asia" disaster monitoring system, which now has over 15 regional member governments and relevant agencies supplying and using the information, to help countries in the Asia Pacific monitor the progression of impending disasters, and asses the impacts of floods, rainfall, landslides, earthquakes and other natural disasters.

In parallel, these remote sensing technologies have also been used in Australia to map the burnt area and burn-scars, grass-curing and other fire-related variables associated to bushfires around Australia. The "AusCover" remote sensing data facility (www.auscover.org.au) of the Terrestrial Ecosystem Research Network (TERN – www.tern.org.au) of Australia, has since 2009 been providing free and open satellite-derived information, at regional and continental scales, for use in fire ecology studies, assessment of fire impacts on protected areas and for estimation of greenhouse gas emissions, to name a few uses. A key satellite-derived product called the "fire-severity index", developed and produced for AusCover by Dr. Stefan Maier at the Charles Darwin University in Darwin, allows local land managers and ecologists to monitor the effect of often unplanned fires and strategically implement controlled burns during less damaging times of year. Similarly the "grass curing index" produced by another partner, the Bureau of Meteorology, provides a way to evaluate the dynamics of grass drying and fire-risk, as dry seasons and summers progress across the continent. Such derived datasets provide ecosystem researchers and conservation managers with greater information about the effects of fires on ecological communities, and improve estimates of carbon emissions resulting from fires in different types of ecosystems.

#### 3.2 The effectiveness of free open access data. The Brazilian example

As Brazil is large geographically—more than 8.5 million km<sup>2</sup>—and has high biodiversity, special ecosystems such as the Amazonian and Pantanal regions, an ever-growing agriculture, a fast-changing land use and land cover, and a long coastline, it is especially suited for space-based remote sensing technologies. Therefore, Brazil has been at the forefront of remote sensing research and application since 1973 when was among the first countries to build and operate its own ground station to receive Landsat-1 data.

At the end of the 1980's, Brazil began the development of a civilian remote sensing satellite program with China called China-Brazil Earth Resources Satellite (CBRES), becoming part of one of the first programs in the world involving two developing countries collaborating to develop and launch remote sensing satellites. To date, a constellation of three satellites has been launched (CBERS-1 in 1999, CBERS-2 in 2003, CBERS-2B in 2007 and CBERS-3 in 2012), and two more satellites are on the way (CBERS-3 planned for 2013, and CBERS-4 planned for 2104).

One of the main aspects of the CBERS Program is the data policy adopted after the CBERS-2 launch. Brazil adopted the free-of-charge CBERS data distribution policy when data are requested in electronic format, opening the field of remote sensing to new users, applications and business. Initially adopted for Brazilian users, it was extended for neighboring countries, and then to the world. Currently, all CBERS data gathered at Cuiaba, the Brazilian ground station, is distributed free of charge to everyone www.dgi.inpe.br/CDSR).

Since the adoption of this open-access data policy, more than 100,000 scenes have been distributed each year inside Brazil to thousands of users and institutions. The processing system is very fast and it takes only a few minutes for the user to have his request for a full-resolution scene fulfilled. This kind of data policy and easy distribution system promoted a strong increase in the number of users and applications. As a result, there is no organization related to agriculture, environment, geology, or hydrology in the country that is not a CBERS user. Hundreds of businesses in remote sensing were opened after the adoption of the current data policy. Significantly, environmental control by society has also increased.

Brazilian legislation requires that each farmer identify and notify the environmental agency about areas to be protected on each farm. This procedure is called environmental licensing and has been adopted in many states around the country. Currently, most of this procedure is done based on CBERS images and has opened hundreds of small businesses specializing in this kind of service. An interesting application of CBERS images is in tax enforcement. Some states use CBERS to help them to monitor farms to assure that all declarations made by farmers are in accordance with the tax law.

Another important environmental application of the fast and free access to CBERS data is to map and measure deforested areas. It is often the case that governmental institutions have difficulty in acquiring up-to-date remote sensing data, especially in developing countries. In Brazil the deforestation in the Amazon region is a major environmental problem. Actions from the governmental environmental protection agency depend on monitoring. Monitoring in the Amazon region on an annual basis used to be based on NASA owned Landsat data, but with the launch of CBERS, the Brazilian capacity to monitor the Amazonia experienced a major increase. In addition, CBERS data is also used, together with MODIS data, in a permanent monitoring system for the Amazonia under a project called Detection of Deforestation in Near Real Time (DETER). It allows detecting early signs of deforestation, and alerting the environmental agency in time to take action.

#### 3.3 Using remote sensing for Protected Area planning in Canada

Canada is the second largest country in the world by land area, at nearly 10 million km<sup>2</sup> in size. Monitoring biodiversity and associated ecosystems for a nation the size of Canada requires approaches that enable broad scale national assessments. Over the past five years the Universities of British Columbia (UBC) and Victoria (UVic) with the Canadian Forest Service (CFS) of Natural Resources Canada (NRCan), have investigated the role remote sensing can play in the assessment of biodiversity across Canada.

This research includes the national level application of indices which capture different aspects of species habitats, and the production of regionalizations or environmental domains which allows for the assessment of, for example, the representation of park networks which can be used to inform national biosiversity planning.

#### Application of a Dynamic Habitat Index (DHI) across Canada

Vegetation productivity is the most widely supported predictor of broad scale biodiversity patterns. In general, regions with higher productivity support higher levels of species richness. Productivity is easily amenable to rapid, repeatable monitoring with remote sensing data. A dynamic habitat index (DHI) has been applied across Canada, a tripartite measure of vegetative productivity, to monitor habitat condition repeatedly and over large extents. The DHI is computed from satellite estimates of the fraction of Photosynetheically Active Radiation (fPAR), an index which provides an indicator of vegetation growth capacity. The three components are:

- 1. Annual average landscape greenness which integrates the productive capacity of a landscape across a year and has long been recognized as a strong predictor of species richness.
- 2. Annual minimum greenness which relates the potential of a given landscape to support permanent resident species throughout the year. Locations without significant snow cover at the end of the summer will often maintain greenness into winter, and vegetation fPAR remaining above 0. In areas where snow covers the vegetation, fPAR approaches 0.
- 3. Seasonal variation in greenness is an integrated measure of climate, topography, and land use. For example, forests and grasslands in the mountainous and interior regions of continents display a much shorter growing season than those in the more maritime ecoregions. High seasonality values signify seasonal extremes in climatic conditions or

limited periods with agricultural production. Sites with low values typically represent irrigated pasture, barren land, or evergreen forests.

These three components of the DHI make it a prime candidate to test hypotheses related to diversityproductivity relationships. I Its dynamic nature, which is tailored to ecological theory, makes it more informative than single remote-sensing metrics (Figure 3.1).

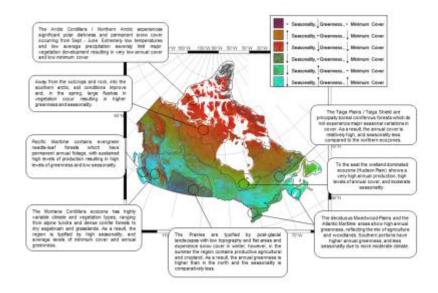


Figure 3.1. The Dynamic Habitat Index of Canada. Different ecological zones throughout the country exhibit different DHI components of productivity, seasonality and minimum cover. As a result spatial differences across the country are apparent as changes in color

The DHI has been derived from MODIS (NASA 2000 onwards) or AVHRR (Advanced Very High Resolution Radiometer (1986 onwards) and is freely available to researchers. The DHI has also been applied across North America and a global DHI product is underway.

#### **Environmental Domains and Conservation Representativeness**

Another approach for the use of remotely sensed derived indicators of biodiversity is to provide information for the characterization of the landbase. The DHI has been used together with other remotely sensed datasets, such as information on land cover, fragmentation, disturbance, snow cover to develop clusters (pixels) into environmental domains, or areas sharing common environmental conditions. Such domains are analogous to traditional ecoregions, however unlike ecoregions, which are forced to include atypical areas by the requirement of spatial contiguity, environmental domains are not spatially discrete and, therefore, allow a more consistent classification of homogenous units. These environmental domains can then be used to assess, for example, representativeness in Canada's network of parks and protected areas and systematic conservation planning of future reserves.

Work in Canada has focused on its Boreal forest where currently, ~8.1 % (448 178 km<sup>2</sup>) is under some form of protection, with many of these areas in low productivity environments located in the far north or at higher elevations. However, because of its remoteness and inaccessibility, ~80% of the boreal already functions as though protected; thus, there exists a vast potential for conservation investment

in the region. Methods which utilized 15 remotely sensed clusters and species at risk data to assess a variety of hypothetical reserve network scenarios were applied, with (i) varied levels of conservation targets and reserve compactness and (ii) the preferential prioritization of remote or intact wilderness areas (Figure 3.2).

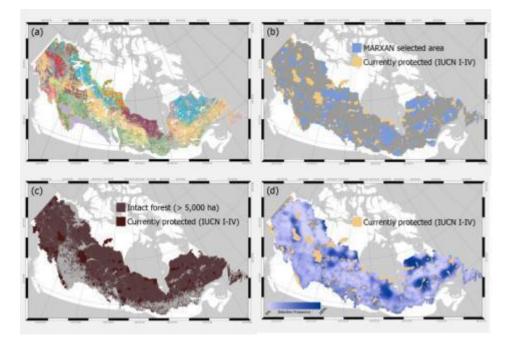


Figure 3.2. (a) Spatial distribution of 15 environmental domains (Powers *et al.*, 2013). (b) A best or near optimal MARXAN reserve design solution for a scenario that preferentially prioritizes remote areas away from human presence using an accessibility cost layer. (c) Global Forest Watch Canada (GFWC) intact forest landscape and current protected areas (IUCN I-IV). (d) The sum of all MARXAN solutions for 500 runs (iterations) of the same scenario. It is used to determine the selection frequency of each planning unit (0-100%), and provides an indication of how important the planning unit is for an efficient reserve design.

Results suggest that reserve compactness greatly influences the reserve area and cost and that restricting conservation to only intact wilderness areas also reduces flexibility and reserve cost efficiency. However, preferentially prioritizing remote portions of the boreal or areas with low human accessibility was able to provide the reserve design flexibility needed to meet all scenario targets and demonstrates that this approach is useful for aiding in biodiversity conservation efforts. Results show that the indirect indicators of biodiversity, which are available from remote sensing, are effective tools for modeling and monitoring biodiversity at national and continental scales and provide valuable insights into basic and applied ecological research.

In order to ensure the preservation of species and habitat diversity and current and anticipated future conditions, all environmental domains should be adequately represented in a comprehensive conservation network. The clustering analysis used to identify domains has also identified environmental conditions that are unique, and thus may be the most deserving of conservation attention. Spatial conservation planning tools such as MARXAN can be used to help determine where (spatially) conservation investment should be prioritized. This method works by finding cost-effective solutions to conservation problems by achieving conservation targets for the least cost, which can include a variety of factors such as area or economic costs associated with land acquisition, management, human accessibility and forgone activity.

# 3.4 Use of remote sensing in data creation for use in biodiversity indicators in South Africa

Remotely sensed data has formed the basis of many indicators used by the South African National Biodiversity Institute (SANBI) in both the National Spatial Biodiversity Assessment (NSBA), 2004 and the National Biodiversity Assessment (NBA), 2011. A total of 16 indicators have been derived (totally or partially) from remotely-sensed data.

Although the remotely sensed data is widely used in indicators, there are only two core data layers that have been created from a direct analysis of remotely sensed data, the National land cover datasets dated 1994 and 2000. The next national land cover dataset is expected to be finalized in 2017 (Parker, 2013). In the interim SANBI has updated the National land cover 2000 dataset with updated provincial land cover data and various other vector data sources (SANBI, 2009). This has provided the base data for the NBA 2011 indicators. The following biodiversity indicators have made use of the land cover as a base data set: Terrestrial ecosystem threat status; Climate change stability in Biomes; and, Biodiversity priority areas.

The following indicators in the NBA 2011 were created using either satellite or aerial photography: River ecosystem threat status; River ecosystem protection levels; Freshwater ecosystem protection areas; and, Flagship free flowing rivers; Wetland ecosystem threat status; and Wetland ecosystem protection levels; Estuarine ecosystem threat status; Estuarine ecosystem protection levels; and, Priority estuaries; Marine and coastal ecosystem threat status; and, Marine and coastal ecosystem protection levels; Species of special concern (specifically medicinal plants and threatened freshwater fish); Invasive alien species (specifically woody invasives).

#### 3.2.1 Limitations

The following limitations have been experienced in using remotely sensed data. In most cases these limitations have resulted in the decision not to use remotely sensed data for indicator generation.

#### Raw data cost vs. spatial resolution

The South African National Space Agency (SANSA) provide Level 3A and 3B SPOT 5 imagery (with a spatial resolution of 2.5m and 10m) to the provinces, the Presidency, government departments and government agencies such as SANBI (SANSA, 2012). The first Spot 5 mosaic of the country was compiled in 2006 (Campbell, 2012). Cape Nature used SPOT 2005 imagery in the CAPE Fine scale analysis (SANBI, 2007); SANBI does not currently pay to access this imagery. Landsat imagery has been obtained via download from United States Geological Survey (USGS) (U.S. Geological Survey, 2012) and Landsat 5 imagery was used in the SANBI vegetation (Mucina & Rutherford, 2006, p. 19).

However certain biodiversity features, such as wetlands, bush encroachment, streams, etc. cannot be identified on Landsat or SPOT. Unfortunately imagery generated by GeoEye and QuickBird are not available to SANBI free of charge and the cost of purchasing all the tiles for South Africa are excessive. This limits the use of remotely sensed data to areas where there are biodiversity features that cover areas in excess of 2.5 m<sup>2</sup>.

#### Analysis of various vegetation types

The differing Biomes in South Africa require different remote sensing approaches to identify the vegetation types within them. In the Fynbos biome it is problematic to identify vegetation using

remote sensing, because veld age seems to be an overriding signature in the vegetation and skews the interpretation (Mucina & Rutherford, 2006, p. 22). This limitation has been mitigated by making use of vector vegetation distribution data. Certain invasive species such as Acacia are also misidentified as Fynbos. This limitation cannot be mitigated due to a lack of invasive distribution data.

In the Grassland Biome remote sensing faces other challenges. Fallow agricultural fields are identified as natural grassland, whereas in reality they contain only a small number of the grass species that pristine Grasslands should contain. This limitation is mitigated through the introduction of a vector layer of cultivated fields (SANBI, 2009).

## Differing mandates and the cost of going commercial

In South Africa there are very limited numbers of remote sensing experts. National Geo-spatial Information, a component of the national Department of rural development and land reform, is responsible for creating and maintaining the National land cover and land use datasets. Unfortunately the process has not yielded a complete dataset since 2000 (released in 2005) and plans to complete the classification and change detection for the entire country only in 2017 (images captured in 2012 – 2014), with a pixel size of 10 m and a minimum mapping unit of 1 hectare (Parker, 2013). To mitigate this limitation the provinces have turned to commercial experts to provide land cover data at a high cost. Three provinces out of a total of nine have developed their own provincial land covers (SANBI, 2008), while a further three provinces have partial land covers. SANBI has mitigated this issue by generating an updated land cover of sorts through the intersection of provincial land covers and various other updated vector layers. This updated national land cover has been generated for 2009 (SANBI, 2009) and will now be updated again for 2013, this layer is primarily used for the generation of other data layers and biodiversity indicators (Driver, *et al.*, 2011).

#### Ground truthing

The ground truthing of land cover data is a limitation for remote sensing in South Africa, since the country is vast and diverse in its land cover, commercial entities have mitigated this by making use of aerial or high resolution satellite imagery to undertake random ground truthing (SANBI, 2008). The fine scale planning project made use of expert workshops (SANBI, 2007) to review the newly generated land cover and determine if it was accurate.

#### Lack of experience

SANBI has as yet not been able to create a full national land cover due to all the limitations mentioned above along with an additional limitation of a lack of skilled staff, software and hardware. Recently SANBI has had one staff member trained in the use of ENVI and has acquired licenses for both ENVI and ERDAS, however the staff required to advise on the science underlying this work are still lacking.

# 3.2.2. Spatial and termporal resolution

National monitoring requires the highest spatial and radiometric resolution possible, so that mapping and analysis can occur at regional as well as national scale. The ideal model of data capture and analysis for monitoring in South Africa is that much of the work happens at the regional (municipal and provincial) scale, this data is merged and gaps are filled to produce the national scale data. However in undertaking this approach it is imperative that the results reflected in the national and regional analyses do not differ, it is thus impossible to make use of SPOT imagery regionally and then Landsat imagery nationally. The requirements for temporal resolution vary between one and five years. Although five years is an acceptable time lapse between land cover data sets, it is also desirable to be able to monitor large land cover changes that happen in much shorter time spans. Considering that it takes approximately one year to collect, classify, check and create a land cover change map, it would be prudent to suggest that the temporal resolution be a minimum of two years and a maximum of four years. In addition when mapping biodiversity features it is imperative to obtain imagery for the wet and dry seasons, in South Africa this would mean a minimum of a December and a June image.

# 3.2.3 Complementary information to develop an indicator

Two key data types are used to complement remote sensing data.

- Existing non-remotely sensed vector and raster data: This data informs the data creation by revealing what is known to be in that location already, for example a portion of land cannot revert back to a natural classification if it has been cultivated, it is most likely fallow instead.
- Expert opinion: Expert opinion in vegetation mapping is crucial. The group of experts, constituting the South African Vegetation Map committee, still meets on a regular basis to discuss changes to the National vegetation map (Mucina & Rutherford, 2006). These changes may be as a result of new species classifications or new field work.

# 3.2.4 Priorities for the future

South Africa is urgently in need of a series of regularly updated land cover datasets that allow for the assessment of the condition of terrestrial ecosystems, rivers, wetlands and estuaries (Driver, *et al.*, 2011). This task would benefit from well-defined leadership and international exposure to best practices in land cover creation, specifically in a biodiversity context.

# 4. Limitations and challenges

### 4.1 What has limited the use of remote sensing in developing indicators?

The selection of an EO product for indicator development requires a trade-off between available data, spatial resolution and coverage, spectral characteristics of the sensor, timing of image acquisition, degree of cloud cover, practicality of ground validation and subsequent analysis, combined with the overall cost of the imagery and analytical effort. Any of these criteria can potentially limit the use of RS data for developing indicators.

## 4.1.1 Type of available data

More user-friendly and intuitive data portals for accessing EO-based data are a requirement for the biodiversity community (Leidner *et al.*, 2012). The type of data that can be accessed through these portals can limit the level of indicator development. For example, pre-processing steps, i.e. georeferencing, topographic correction, orthorectification and atmospheric correction, should be done centrally and systematically, so as to produce a consistent set of EO products which are ready to use. More standardisation of approaches can be achieved under initiatives such as the GMES fast-track service, making EO-based analysis more cost effective and efficient to the end-user community (Infoterra, 2007). The Joint Research Centre (JRC) Digital Observatory for Protected Areas (DOPA) web service has automated the collection and pre-processing of remotely-sensed imagery in order to provide protected-area level biodiversity information (Dubois et al., 2011). The GFW 2.0 monitoring system also incorporates a consistent set of pre-processing steps to generate consistent deforestation information from Landsat imageryalthough this is also in development and has not been officially launched at the time of writing. Therefore initiaves are under way to adress the need for a centralised system of digital image collection and processing.

The lack of suitable product documentation and metadata has also been cited as a limitation associated with EO-based products. Operational products provided through Copernicus or NASA are accompanied by technical documentation which can assist users in understanding the content of a product, its limitations and strengths and its application. These are commonly in the form of an Algorithm Theoretical Basis Document (ATBD).

Finally, the level of product development from unprocessed satellite imagery is also an important concern. Frequently, derived geophysical fields, such as vegetation indices, are more useful than raw remote sensing data to non-specialists (Leidner *et al.*, 2012). The Copernicus Global Land service and similar systems in use by NASA, e.g. the Distributed Active Archive Centers (DAACs), enhance end-user capabilities by providing ready to use and free geophysical and biophysical products from satellite imagery. However, limitations on bandwidth and internet access speed in developing countries can be a constraint on data access and limit the use of EO data (Roy *et al.*, 2010).

#### 4.1.2 Cost of data acquisition and data access policy

Access to EO data is frequently highlighted as a key limitation by many biodiversity stakeholders. Many space agencies and some countries are now offering free and open data access to their satellite data. Thus, some Earth Observation data products are freely available to the community but some are not, especially high and very high spatial resolution imagery (Leidner *et al.*, 2012). To date, this has limited the development of EO-based products in the biodiversity community to Landsat and MODIS which are typically free and suited for regional scale applications. The launch of NASA Landsat 8 in February 2013

and the upcoming ESA/EC Copernicus Sentinels offer more access to high resolution data. For more detailed information on data production and acquisition, please refer to Annex 5.

However, open access to remote sensing data is sometimes conditional on the type of user, whether it is a research organization, private sector or academic department. More barrier-free approaches with no organizational or user access limitation, such as NASA's access policy to its USGS archive and Landsat data would be extremely useful. However, a full and open access data policy does not necessarily mean easy and fast data access. For example, ESA/EC Copernicus Sentinels data policy will allow a free and open data access but it is not yet clear how easy the data will be accessible especially outside ESA Member States.

Larger scale mapping is now possible with the advent of private sector, airborne and spaceborne sensors with spatial resolutions appropriate for local to site-level land cover mapping (Infoterra, 2007). However, the financial cost is proving a challenge to most biodiversity researchers and conservation practitioners as very high resolution data are expensive to acquire (Leidner *et al.*, 2012).

One possibility to overcome this limitation is the involvement of government agencies in public-private sector partnerships to enable researchers and analysts to access high resolution data at low cost. For example, several federal agencies of the U.S. government have established data purchase programs with commercial image providers in order to access new commercial remote sensing products which meet research and operational requirements (Birk *et al.*, 2003). This requires initiative on the part of government bodies to recognise the duty that central Government plays in providing mapping and monitoring information to meet the needs of its citizens. An agreement between NASA Earth Science Enterprise (ESE) and the Space Imaging IKONOS system has been a good example of cooperation between industry, government and end users (Goward *et al.*, 2003). However, the organisational and legal aspect of the partnership is more of an important determinant of success than any technical factors (Goward *et al.*, 2003).

#### 4.1.3 Internet access and data access

Linked to the above limitations is the issue of internet access in certain regions. For example, access to the USGS Landsat archive is considerably constrained by a limited bandwidth in many African countries (Roy *et al.*, 2010). However, while the situation is improving, with new fibre-optic cables opening up access to broadband connectivity, there are still problems of establishing networks within countries. Government regulation may also continue to restrict Internet access across the continent (Roy *et al.*, 2010).

#### 4.1.4 Capacity to use EO-based data in indicator development

A lack of capacity among biodiversity experts is frequently cited as a limitation in using remote sensing for monitoring biodiversity and developing indicators (Leidner *et al.*, 2012). A greater understanding of how to use remotely-sensed information is often sought in preference over more computing power or more advanced EO products. For example, there have been calls for more access to open-source software and more online resources and guidebooks for the conservation community (Leidner *et al.*, 2012).

Generally, indicator development from raw remote sensing data requires capacity and expertise in numerical data processing and statistical analysis. This is a common limitation to both developed and

developing nations. More information on data analysis and process costs can be found in Annex 5. Centres of expertise for remote sensing to address user needs at a regional or national level may be beneficial, as has been done with the Canada Centre for Remote Sensing (CCRS) for example.

#### 4.1.5 Effective data validation strategy

The lack of sufficient validation has limited the use of remote sensing data by biodiversity practitioners. More in-situ measurements are required for the calibration and validation of terrestrial EO products if they are to be used with confidence by biodiversity practitioners (Infoterra, 2007). Space agencies should also be concerned with in situ data for validating EO products, without which EO-based products are less likely to be used with confidence (Green *et al.*, 2011). However, there are efforts to address this issue. For example, the CEOS Land Product Validation (LPV) subgroup has eight thematic areas where it is actively pushing efforts to globally validate EO-based products using in-situ measures. The themes are diverse and vary from validation of phenology products to snow cover, fire/burn area and land cover products (CEOS LPV, 2013). The U.K. Department of Environment, Food and Rural Affairs (Defra) Science Directorate has already addressed some of the limitations in the use of EO data for biodiversity monitoring in the UK. In China, significant investment in land cover classification and validation is likely to yield global land cover change products in the near future.

Land cover is a thematic area that needs advanced ground validation strategies especially if land cover change is to be monitored with reliability (Green *et al.*, 2011; Hansen and Loveland, 2012). The most frequent reason for the absence of accuracy assessment is the lack of contemporary ground data with sufficient spatial coverage (Infoterra, 2007). Field campaigns are generally costly, labour intensive and sometimes difficult to synchronise with satellite image acquisition. However, an effective validation strategy is critical if the EO-based approach to landcover and habitat mapping is to be proposed as a cost-effective alternative to field-based methods (Vanden Borre *et al.*, 2011). Online tools such as DOPA will provide capacity for the validation of uploaded products by end users using Google Earth.

#### 4.1.6 Insufficient spatial resolution and spatial scale

The issue of spatial scale is often cited as a limitation to indicator development as operational remote sensing products are provided at spatial resolutions which are often coarser than needed for operational monitoring. For example, tackling conservation issues, such as loss of habitat, at the level of protected area, requires an indicator which is sensitive to that scale of change. Land cover, for example, is a particularly scale-sensitive parameter. A global or continental scale landcover product such as GLC 2000 or Globcover might meet nationa level needs but not be appropriate to address change at the protected area level. However, a product developed to meet the needs of protected area level monitoring is unlikely to be generated globally, on a routine basis, due to sensor limitations.

There is a demand among the biodiversity community for land cover products at the Landsat spatial scale ( $\leq$ 30m) and MODIS/AVHRR scales (250-1000m) (Leidner *et al.*, 2012). However, high resolution land cover ( $\leq$ 5m) information can also be very beneficial for monitoring site -specific variation at the plant community level or to map surface objects such as tree crowns and hedgerows. Two European GMES projects, Biodiversity Multi-Source Monitoring System: From Space to Species (BIOSOS) and MS MONINA, are researching EO-based tools and models for monitoring NATURA 2000 sites and their surroundings incorporating high or very high resolution satellite imagery. Indicator development at the

local level, using airborne or higher resolution satellite sensors, can be a potential solution to address site-specific conservation needs but is not yet operational.

# 4.1.7 Long temporal repeat cycle and short time series for trend analysis

The temporal rate of change in surface processes is inconsistent with the repeat cycle of some EO satellites and therefore may limit the sensitivity of the product to detect certain surface changes. For example, the 16-day repeat cycle of Landsat is further limited by seasonality and cloud cover, especially in tropical areas; reducing the effectiveness of annual land cover updates (Hansen and Loveland, 2012). However, the INPE in Brazil have developed the DETER product (see section 3.2 for further details), which uses daily MODIS data to provide a near-real time alert system to relevant authorities to monitor Amazon deforestation (Hansen and Loveland, 2012).

The low revisit time can limit the applicability of Landsat to indicator development, especially where surface change is on a daily to weekly time scale. Furthermore, time composited satellite products, e.g. 8-day MODIS, are insensitive to some natural phenomena, e.g. phenological changes in terrestrial vegetation, which occur on finer time scales (Cleland *et al.*, 2007). A high revisit time is required for optimal change monitoring, as provided by the Landsat 8 and Sentinel 2 satellites, with a revisit time of 4-5 days for example. However, there is always a careful balanced between the spatial resolution, spatial coverage and repeat visit time of the satellite sensors.

The length of remote sensing time series can be limiting on efforts to monitor long-term change in ecosystems. However, there is a need to characterise decadal-scale land cover change and at a global level with landcover classes which can be related to the ecosystem level (Leidner *et al.*, 2012). Decadal-scale time series are only available for certain sensors, e.g. Landsat and AVHRR while MODIS and MERIS time series are limited to a decade approximately. This is a particular problem for land cover products which tend to be a static representation of one point in time with only a few periodic updates, e.g. CORINE 1990, 2000 and 2006 for Europe.

# 4.1.8 Harmonisation of methodologies and data collection at national and international level

Greater coordination of methods in data collection and processing is required for harmonised EO products. This is one of the aims of the GMES initiative (Infoterra, 2007). For example, there are calls for a consistent pan-European habitat typology to reduce the uncertainty surrounding the inter comparison of national-level habitat classification systems (Vanden Borre *et al.*, 2011). However, the kind of habitat parameters which can be retrieved is highly dependent on pixel size and sensitive to scale (Nagendra, 2001). Therefore, any harmonisation of efforts across national systems must take into account the availability of appropriate imagery. The Group on Earth Observations Biodiversity Observation Network (GEO BON) has been set up to focus efforts among different agencies in linking observing system for an integrated biodiversity monitoring system (Scholes *et al.*, 2012).

# 4.1.9 Cloud clover

Cloud cover is a significant limitation to optical remote sensing. This has forced end users to accept a 'use what you can get' approach that has made it difficult to streamline EO-based working procedures (Infoterra, 2007). However, there has been progress in automating the process of cloud removal and atmospheric correction through a harmonised approach to pre-processing methodologies. For example, the Landsat Ecosystem Disturbance Adaptive Processing (LEDAPS) system has applied cloud

and cloud shadow removal, as well as automatic atmospheric correction, to a collection of Landsat 5 and Landsat 7 scenes. This harmonisation of cloud screening and atmospheric correction methods results in a consistent set of pre-processed Landsat imagery. These scenes are available through the USGS Earth Explorer site under the Landsat CDR option in the Datasets list. On demand pre-processing of any Landsat scene is now possible through the LEDAPS system.

In addition to the above, due to their specific characteristics, terrestrial, marine and intertidal environments possess unique limitations to indicator development using remotely-sensed data.

#### 4.1.10 Specific limitations of remote sensing in terrestrial ecosystems

The terrestrial domain has not yet developed a joined up approach, involving multiple disciplines, to gain a greater understanding of the global terrestrial system, as has been done in the marine environment (Infoterra, 2007). For example, The World Meteorological Organization (WMO) and the Intergovernmental Oceanographic Commission (IOC) of UNESCO have developed a joint working group for a global met-ocean observing network in which remote sensing observations play a crucial role (JCOMM, 2013). This has hindered the development of terrestrial simulation/prediction models which have been more widespread in the marine and atmospheric domains (Infoterra, 2007). Terrestrial ecosystem variables derived from remote sensing can play a key role in model development.

Typical terrestrial habitat variables include tree, shrub or grass species composition, canopy cover, tree size distribution, density of dead trees, three-dimensional forest structure, understory characteristics, vegetation architecture and the timing and duration snow and ice cover (Green *et al*, 2011). The benefits of UAVs in mapping and monitoring these variables at close range are discussed in detail in Annex 3. However, their use in terrestrial environmental applications to date has been limited by restrictions imposed by civil aviation authorities. UAV technology is easier to apply to marine applications, whereas airspace management over land is more complex (Infoterra, 2007).

A challenging area for EO is to supply adaptable landcover products which can answer specific biodiversity and conservation research questions at a suitable spatial resolution, with sufficient spatial coverage, accuracy that can be updated when and where change occurs. Global land cover mapping at low resolution is challenging and has not always produced comparable results. For example, there are inconsistent cover estimates between GLC-2000, MODIS and GlobCover, especially for cropland, which introduces uncertainty in end user applications. Ways to overcome these challenges in future global landcover products include increasing data sharing efforts and the provision of more in situ data for training, calibration and validation (Fritz *et al.*, 2011).

It is also challenging to translate landcover to habitat type, though it is often used as a proxy for habitat, the assumption that they are equivalent is questionable. However, mapping habitat directly from remote sensing imagery has been achieved using medium resolution satellite imagery, in the Phase 1, national-scale habitat map of Wales for example (Lucas *et al.*, 2011). The method was based on object-oriented, rule-based classification coupled with multi-temporal, multi-sensor imagery and shows considerable promise in providing habitat-specific change updates. Such continual monitoring of habitat change, at the national scale, is not possible with current static landcover maps.

Landcover is not the only EO variable in use to infer habitat characteristics. Habitat variables such as species diversity and species richness can be estimated from spectral information alone (Rocchini *et al.* 

2010, 2004). Variables such as VCF and fCover (see Annex 2) offer an alternative approach to global landcover mapping. Instead of considering discrete borders between landcover types, the VCF product estimates a continuous field of woody vegetation cover. This is a more realistic interpretation of gradients in spatial landcover variability (DeFries *et al.*, 1999). Products such as fCover and VCF could potentially be one of several layers in an adaptable landcover map that could be routinely updated. However, understanding how EO products translate across different scales has been noted as a limitation in the terrestrial system (Infoterra, 2007). For example, LAI, FAPAR and fCover all demonstrate variable sensitivity to scale (Weiss *et al.*, 2000), and LAI is scale dependent, while fCover is not (Baret *et al.*, 2011). In addition, generating continuous-field land cover datasets at Landsatresolution and on a global level is challenged by the difficulty of acquiring suitable reference data for validation. Local LiDAR measurements of tree height could be a potential solution to bolstering ground-based validation efforts (Sexton *et al.*, 2013).

## 4.1.11 Specific limitations of remote sensing in aquatic ecosystems

Remote sensing and spatial analysis techniques used to study aquatic ecosystems differ from those used in terrestrial systems (Strand *et al.*, 2007). This is largely due to the nature of reflectance from water bodies which reflect sunlight in different wavelengths to those from terrestrial surfaces, e.g. water bodies appear very dark in satellite images due to almost total absorption of near infrared radiation (Campbell, 2006).

The typical satellite sensor used in marine environments is therefore different in design and instrumentation to that used in terrestrial areas. For example, Synthetic Aperture Radar (SAR) systems such as Radarsat-1, Envisat ASAR and ALOS PALSAR, are mainly intended for marine applications such as oil-spill monitoring, ship detection, shallow-water bathymetry mapping, sea-ice monitoring and sea surface state (Infoterra, 2007, Kerbaol and Collard, 2005). Other satellite sensors such as the NOAA AVHRR and METEOSAT are dedicated to marine meteorology and tracking extreme events such as hurricanes.

The two great benefit of EO-based monitoring of oceans and water bodies is the synoptic view of the spaceborne sensors and their regular repeat cycles which allow dynamic processes to be monitored on a regular and repeatable basis (Campbell, 2006). The aquatic environment and the wider hydrological cycle demonstrate unique challenges to EO-based monitoring however. For example, ocean colour monitoring sensors such as SeaWiFS and Envisat MERIS measure slight changes in colour which are easily attenuated by atmospheric interference. Highly dynamic surface features such as ocean currents and the movement of suspended sediment can occur at a rate not measurable by polar orbiting sensors. The recently launched Geostationary Ocean Color Imager (GOCI) has been designed to monitor short-term and regional oceanic phenomena in order to address this problem (He *et al.*, 2013).

Within the marine community, the use of EO data for monitoring biodiversity is relatively widespread and there is a core set of global and regional products to serve user needs (Infoterra, 2007). Such products are underpinned by a good scientific understanding of many of the processes in the marine environment. This has led to well established fields of research such as remote sensing for monitoring individual marine species, using telemetry (e.g., Blumenthal *et al.* 2006), or factors controlling their distribution, such as algal blooms (e.g., Burtenshaw *et al.* 2004). However, it is worth noticing that remote sensing is more typically used in mapping tropical rather than temperate marine areas as the visibility through the water column is generally better due to lower a lower volume of suspended sediment (Strand *et al.*, 2007).

For aquatic environments, key environmental parameters required by the conservation community have been listed as "biological productivity of marine areas (critical for all marine spatial distribution models), sea surface temperature, frequency of marine and freshwater algal blooms, plankton density, seasonality of extent of sea ice cover, including polynas, sediment type of intertidal zones, bathymetry of intertidal zones (and hence the duration of tidal coverage), the mobility of intertidal mud and sand flats, volume and seasonal pattern of river flows and species identity of emergent marsh vegetation" (Green *et al.*, 2011).

However, not all of these variables are routinely monitored by satellite sensors. For example, more data are needed on carbon storage and sequestration value in oceans – similar to those which are used to generate maps of terrestrial carbon (Green *et al*, 2011). However, there are currently large discrepancies between satellite-based and model-based estimates. Furthermore, satellite-based estimates tend to suffer from wide error margins. For example, the Southern Ocean  $CO_2$  sink in 1997/1998 was estimated at –0.08 GtC yr–1 with an error of 0.03 GtC yr–1 (Rangama *et al*. 2005) which was approximately 38% smaller than that based on in-situ measurements and climatological data of the same area (Takahashi *et al.*, 2002). Some of this uncertainty can be explained by the weak correlation between in-situ and RS-derived measures of the same surface variable, e.g. chlorophyll-a, which are used in the estimation of CO2 flux (Chen *et al.*, 2011).

There is less understood on habitat fragmentation and connectivity in marine habitats than for terrestrial ecosystems (Strand *et al.*, 2007). High-resolution measurements based on LiDAR can offer spatial, structural as well as thematic information on localised coastal habitats (Collin *et al.*, 2012), while offshore benthic habitat mapping can be achieved with a combination of ship-based sonar devices and LiDAR (Costa *et al.*, 2009). However, it is challenging to acquire the same level of information on a broader scale due to logistical constraints and financial cost. Therefore, mapping the connectivity of the marine habitat is not straight forward as different remote sensing platforms are employed and are not always compatible in producing seamless habitat maps.

#### 4.1.12 Specific limitations of remote sensing in the intertidal zone

Intertidal habitats such as mangroves, sea grasses and salt marshes exhibit both terrestrial and marine characteristics. However, satellite and airborne mapping methods for these habitats are less developed than those for purely terrestrial or marine (Green *et al.*, 2011) and the selection of appropriate imagery is constrained by tidal regime where the surface cover is frequently inundated by water. Spatiotemporal variation in substrate, i.e. sand, mud and gravel and dynamic processes such as coastal currents and tides also make the intertidal zone difficult for ground validation work.

Therefore, for satellite image selection or for planning an airborne survey, a balance must be achieved between tidal regime, cloud cover, vegetation seasonality, timing with field visits and the need for very high spatial resolution imagery (Murphy *et al.*, 2008). Furthermore, airborne surveys tend to be expensive and logistically challenging and therefore not suitable for operational monitoring. Field-based methods such as diver survey, underwater videography and acoustic techniques such as sonar can be used in a complimentary fashion in mapping shallow coastal habitats but suffer from error in

interpolation of mostly point measurements (Dekker *et al.*, 2005). A nested approach, employing observations at multiple scales, combining in-situ and airborne mapping methods, appears to be the future for high resolution mapping of intertidal zones.

# 4.2 Key challenges in the use of remote sensing for indicator development

# 4.2.1 Knowledge transfer and capacity building

Knowledge transfer in remote sensing education is a particular challenge for the developing world as traditional expertise in the topic is located in western institutions. Despite some access limitations, the benefits of internet access for knowledge exchange in the field of remote sensing are numerous. Firstly, access to geospatial data is almost on demand, secondly, access to a network of scientists and practitioners who can assist each other remotely, and thirdly, development of EO-based data sets that are coordinated locally, e.g. in citizen science initiatives (Global Marketing Insights, 2009).

In addition, a lack of capacity building is of particular importance in developing countries where there is rarely access to commercial software, appropriate educational material or university - based education in remote sensing. North-South knowledge transfer is been promoted with approaches such as that adopted by ESA, whose EO projects have a strong capacity building component, covering both basic education on remote sensing theory and training courses on particular EO products. South-South cooperation will also be key to improving capacity at national level. In this regard, Brazil, through the National Institute for Space Research (INPE), has led the way in making remote sensing courses available to professionals in Latin America since the mid 1980s (Sausen, 2000).

# 4.2.2 Product accuracy

Accuracy of EO data is an issue in several themes of the discipline, e.g. in landcover mapping and land cover change detection, and in recording position-accurate geospatial data in the field and accurate EO-derived inputs for modeling work (Infoterra, 2007). As EO data are prone to error, uncorrected data are limited in their utility for ecological applications (Kerr and Ostrovsky, 2003). In a survey of nature agencies involved in management and monitoring of NATURA 2000 sites, it was found that thematic accuracy of EO-based habitat maps is seen as the most important measure of quality (Vanden Borre *et al.*, 2011). According to the Committee on Earth Observation Satellites (CEOS) Societal Benefit Area on Biodiversity, a critical drawback of EO data is spatial accuracy and alignment (Leidner *et al.*, 2012). Therefore, an EO-based approach to indicator development will be hindered by issues of reliability unless steps are taken to address error and uncertainty in input data.

The abstraction of remote sensing data in geographical information systems from lower to higher levels tends to propagate error and accumulate uncertainty (Gahegan and Ehlers, 2000). The challenge of product accuracy might be addressed on two fronts, firstly by promoting methods which produce the least error (harmonization of methodologies will play a key role in this) and by limiting the number of processing steps performed on raw EO data (quantifying error at every transformation step can help calculate overall error). Thorough documentation of error and highlighting the limitations of EO-based products must become mandatory if EO-based biodiversity indicators are to be used with confidence.

# 4.2.3 Uncertainty in long-term continuity

Ensured long-term (decadal) continuity of earth observations is a key requirement for user organizations interested in biodiversity change. Therefore, uncertainty in long-term continuity is a key

challenge to increasing the use of remote sensing in monitoring biodiversity as it restrains some organizations to invest in EO projects and development. Initiatives such as ESA/EC Copernicus Sentinel missions that are envisaged to guarantee a long term continuity of earth observations for future decades (+25 years) will be very beneficial.

# 4.2.4 Dialogue between EO community, biodiversity practitioners and decision makers

Greater dialogue between the remote sensing community, biodiversity practitioners and decision makers has often been called for. Within the scientific community, dialogue between earth observation and biodiversity experts has significantly improved over the last years, as demonstrated by the substantial increase in biodiversity related EO publications. The major gap seems to be insufficient dialogue with decision makers. Improved dialogue can have many positive results. For example, clearer user requirements can be expressed, data and options for image processing can be thoroughly evaluated, unrealistic expectations can be moderated or refined, and the cost effectiveness of different options discussed take place (Kennedy *et al.*, 2009).

The CEOS Group on Remote Sensing for Biodiversity and Conservation is an example of such an initiative as well as the Land Product Validation (LPV) sub-group of the CEOS Working Group on Calibration and Validation. The latter initiative is particularly important as it requires validation of the spatial and temporal consistency of EO products using in-situ data gathered by field experts.

# **5. Conclusions**

- Remotely sensed data and derived-measures, combined with appropriate validation and modeling, has improved insights into the ecological processes and anthropogenic disturbances that influence biological diversity, and have shown potential to fill gaps in the suite of indicators that could be used to track the implementation of the Strategic Plan for Biodiversity 2011-2020 and the achievement of the Aichi Biodiversity Targets. With a large number of examples to demonstrate this potential, remote sensing and biodiversity experts are beginning to explore these opportunities. However, caution should be taken not to oversell the promise of remote sensing for monitoring biodiversity. It is **not a fit-for all solution**, and despite the important contribution it has the potential to provide to any biodiversity monitoring system, validating the remotely sensed data with ground truth data and traditional methods of inventorying and assessing biodiversity will still be required.
- As explored throughout this review, there are potentially many areas for future development • of remote sensing products experts could focus on. However, human and financial resources are limited and therefore priorities must be established. As part as an enhanced dialogue between the different stakeholders, priorities should be driven by end users needs. A significant requirement of the conservation community is for long-term Land Cover Change (LCC) products. Current global landcover products are too coarse in resolution, single-date or infrequently updated. Consistent and repeatable land cover products over time, adopting a standardised hierarchal classification scheme, e.g. the Land Cover Classification System (LCCS), can address this need. As landcover changes such as agricultural expansion have been identified as major drivers of biodiversity loss, monitoring landcover change over time can identify where the pressures are occurring and how likely they are to impact the current status and future trends in global biodiversity. The success of conservation interventions can also be measured by assessing landcover change in and around protected areas. However, it is vital that the spatial resolution of such products are commensurate with the scale of conservation units e.g. ecoregions and units smaller than these.
- Monitoring forest cover change has been the area of most intense research in global analyses of land cover change to date. There are numerous reasons for this. Firstly, forests are most easily distinguished in satellite imagery than other vegetation cover types, such as croplands or urban areas. Forest reserves are important conservation areas and are global in distribution. Monitoring forest cover change has important implications for carbon accounting, biodiversity monitoring, and other issues such as illicit logging. However, there is a need to address this bias in land cover monitoring. Other terrestrial ecosystems such as open grasslands, savannah, peatlands and wetlands also need to be considered in land cover change studies. They provide ecosystem services such as carbon storage, clean drinking water, fuel and shelter and are important habitat. Although marine ecosystems are not as readily monitored as terrestrial ecosystems for biodiversity purposes, inshore and intertidal ecosystems are also important landcover types. However, these are considerably challenging

landcover types to monitor as their discminination is difficult, and therefore require further research and development of routine and robust monitoring methods.

- Remote sensing products are a useful tool to assess the effectiveness of conservation interventions. However, most of the work done to date has focused on forested protected areas. Further habitats types and broarder sets of data need to be included in future studies to expand the use of remote sensing in monitoring implementation of the Strategic Plan for Biodiversity 2011-2020.
- To date, dialogue between data providers and end users has been limited. There is a disconnection on the awareness of what is available, what can be done and what is expected. A closer relationship between the earth observation community and potential users in the biodiversity policy and management communities would help to enhance understanding, align priorities, identify opportunities and overcome challenges, ensuring data products more effectively meet user needs.
- Developing indicators to monitor biodiversity in general, and the Aichi Biodiversity Targets in
  particular can be challenging and heavily data consuming. Most biodiversity indicators need a
  variety of data streams, from several sensors and often including non remotely-sensed
  sources. It can become a challenge to have all of them available at the required time, spatial
  coverage and temporal resolution. It only takes a blockage in one of the data streams to
  prevent execution and development of the indicator. This complexity makes it even more
  necessary to nurture a productive dialogue among all data providers and end users in order to
  facilitate and align priorities.
- The link between remotely-sensed derived measures and the development of indicators for high-level policy making is still poorly developed. There is a lack of common standards regarding the measures required by the biodiversity community and the spectral information collected by the remote sensing community. In addition, a full harmonization of methodologies and data collection at national and international level and a delivery approach that works across different landscapes is still not in place. An agreed set of minimum requirements and common standards from biodiversity monitoring practitioners would help focus the efforts of the Earth Observations' experts. Initiatives such as the development of EBVs led by GEO BON could offer the necessary conceptual framework to bridge the gap between both communities and map the pathway from primary remote sensing observations to the delivery of high-level indicators. Closer collaboration between the GEO BON community on the establishment of EBVs and the BIP work on biodiversity indicators could contribute to this.
- Bottlenecks in data access are a key limitation for the expansion of remote sensing for biodiversity monitoring. Free open access data policies have been adopted and implemented by various space agencies and national institutions to date, proving effectively for increasing the use of remote sensing in biodiversity monitoring, as well as enhancing policy implementation and law enforcement in some cases. Free open data access schemes should

continue to be the international trend among data providers to support the democratization of access to remotely-sensed data. **Free and open access to all taxpayer-funded satellite remote sensing imagery** will address this significant constraint.

- However, free open access data policy does not necessarily translate into easy and fast data access. This might be due to limited bandwidth and internet constrains, or related to a hierarchical approach to prioritizing data dissemination among different user groups. A concerted international action to secure easy access to remotely-sensed data should be implemented, especially to ease access from developing countries.
- Enhanced access to data will only be effective if Parties have the sufficient technical and human capacity to make use of it. The international trend of including a major capacity building component in Space Agencies' Earth Observations projects will play an important role. In addition, better mechanisms should be established to support the participation of Parties in Space Agencies' projects.
- Uncertainty in the long-term (decadal) continuity of Earth Observations from satellites and other remote sensing missions is a key challenge for the funding of projects as it restrains funders from invesing in Earth Observation projects, affecting further research and development on remote sensing. More initiatives to guarantee a long term continuity of Earth Observations are needed.
- Accessing comprehensive information on Earth Observations is often difficult for Parties since
  it is still very scattered, hosted by different organizations, space agencies and national
  agencies, and across a wide range of projects. Therefore, what is missing for Parties to the
  CBD and other international Conventions and MEAs is to have a unique reference point they
  can consult on Earth Observation matters in relation to biodiversity (much as the BIP
  represents for information on biodiversity indicators). Such a reference entity that would act
  as a hub to concentrate and coordinate existing information and is easily accessible globally
  could be a key component to facilitate greater use of remotely-sensed data and products in
  biodiversity monitoring. This hub would require significant work to constantly offer the most
  updated information due to the fast pace of development of the EO field.

# 6. References

Achleitner, D., Gassner, H. & Luger, M., 2012. Comparison of three standardised fish sampling methods in 14 alpine lakes in Austria. *Fisheries Management and Ecology*, **19**(4), pp.352–361.

Allouche, O., et al., (20129. Area-heterogeneity tradeoff and the diversity of ecological communities. Proceedings of the National Academy of Sciences of the United States of America **109** (43), pp. 17495-17500.

Amit Chawla, P.K.Y., 2012. Long-term ecological and biodiversity monitoring in the western Himalaya using satellite remote sensing. *Current Science*, **102**(8), pp.1143 – 1156.

Andrew, M. & Ustin, S., 2008. The role of environmental context in mapping invasive plants with hyperspectral image data. *Remote Sensing of Environment*, **112**(12), pp.4301–4317.

Antoine, D., André, J.-M. & Morel, A., 1996. Oceanic primary production: Estimation at global scale from satellite (Coastal Zone Color Scanner) chlorophyll. *Global Biogeochemical Cycles*, **10**(1), pp.57–69.

Arino, O., Plummer, S. & Defrenne, D., 2005. Fire Disturbance: The Ten Years Time Series of the ATSR World Fire Atlas. *Proceedings of the MERIS (A)ATSR Workshop 2005 (ESA SP-597). 26 - 30 September 2005 ESRIN*.

Arizaga, J. *et al.*, 2011. Monitoring communities of small birds: a comparison between mist-netting and counting. *Bird Study*, **58**(3), pp.291–301.

Asner, G.P. & Martin, R.E., 2009. Airborne spectranomics: mapping canopy chemical and taxonomic diversity in tropical forests. *Frontiers in Ecology and the Environment*, **7**(5), pp.269–276.

Asner, G.P. *et al.*, 2010. High-resolution forest carbon stocks and emissions in the Amazon. Proceedings of the National Academy of Sciences Of the United States of America **107** (38): 16738-16742.

Baccini, A., Goetz, S.J., Walker, W.S., Laporte, N.T., Sun, M., Sulla-Menashe, D., Hackler, J., Beck, P.S.A., Dubayah, R., Friedl, M.A., Samanta, S., Houghton, R.A. 2012. Estimated carbon dioxide emissions from tropical deforestation improved by carbon-density maps. *Nature Clim.Change* 2, 182-185

Baccini, A., Goetz, S.J., Laporte, N., Sun, M., Dong, H. 2011. Reply to Comment on 'A first map of tropical Africa's above-ground biomass derived from satellite imagery'. *Environmental Research Letters* 6, 049002.

Baccini, A. *et al.*, 2008. A first map of tropical Africa's above-ground biomass derived from satellite imagery. *Environmental Research Letters* 3, 045011.

Bailey, S.-A., Haines-Young, R.. & Watkins, C., 2002. Species presence in fragmented landscapes: modelling of species requirements at the national level. *Biological Conservation*, **108**(3), pp.307–316.

Baker, G.H., Tann, C.R. & Fitt, G.P., 2011. A tale of two trapping methods: Helicoverpa spp. (Lepidoptera, Noctuidae) in pheromone and light traps in Australian cotton production systems. *Bulletin of entomological research*, **101**(1), pp.9–23.

Balch, W.M., 2005. Calcium carbonate measurements in the surface global ocean based on Moderate-Resolution Imaging Spectroradiometer data. *Journal of Geophysical Research*, **110**(C7)

Banks, A.C. *et al.*, 2012. A satellite ocean color observation operator system for eutrophication assessment in coastal waters. *Journal of Marine Systems*, **94**, pp.S2–S15.

Baret, F. *et al.*, 2013. GEOV1: LAI and FAPAR essential climate variables and FCOVER global time series capitalizing over existing products. Part1: Principles of development and production. *Remote Sensing of Environment*, **137**, pp.299–309.

Baret, F. *et al.*, 2007. LAI, fAPAR and fCover CYCLOPES global products derived from VEGETATION. *Remote Sensing of Environment*, **110**(3), pp.275–286.

Baret, F. et al., 2011. Towards an Operational GMES Land Monitoring Core Service, BioPar Methods Compendium, LAI, FAPAR and FCOVER from LTDR AVHRR series.

Berni, J.A.J. *et al.*, 2009. Mapping canopy conductance and CWSI in olive orchards using high resolution thermal remote sensing imagery. *Remote Sensing of Environment*, **113**(11), pp.2380–2388.

Birk, R.J. *et al.*, 2003. Government programs for research and operational uses of commercial remote sensing data. *Remote Sensing of Environment*, **88**(1-2), pp.3–16.

Blumenthal, J. *et al.*, 2006. Satellite tracking highlights the need for international cooperation in marine turtle management. *Endangered Species Research*, **2**, pp.51–61

Boersma, P.D. & Parrish, J.K., 1999. Limiting abuse: marine protected areas, a limited solution. *Ecological Economics*, **31**(2), pp.287–304.

Vanden Borre, J. *et al.*, 2011. Integrating remote sensing in Natura 2000 habitat monitoring: Prospects on the way forward. *Journal for Nature Conservation*, **19**(2), pp.116–125.

Baldeck, C. A., *et al.* (2013). Soil resources and topography shape local tree community structure in tropical forests. *Proceedings of the Royal Society B-Biological Sciences* **280**(1753)

Braga-Neto, R., Magnusson, W. & Pezzini, F., 2013. *Biodiversity and Integrated Environmental Monitoring*, Santo André, SP, Brasil: Instituto Nacional de Pesquisas da Amazônia (INPA). Áttema Editorial.

Brewin, R.J.W. *et al.*, 2011. An intercomparison of bio-optical techniques for detecting dominant phytoplankton size class from satellite remote sensing. *Remote Sensing of Environment*, **115**(2), pp.325–339.

Burrage, D.M., 2002. Evolution and dynamics of tropical river plumes in the Great Barrier Reef: An integrated remote sensing and in situ study. *Journal of Geophysical Research*, **107**(C12), p.8016.

Burtenshaw, J.C. *et al.*, 2004. Acoustic and satellite remote sensing of blue whale seasonality and habitat in the Northeast Pacific. *Deep Sea Research Part II: Topical Studies in Oceanography*, **51**(10-11), pp.967–986.

Butterfield, H.S. & Malmström, C.M., 2009. The effects of phenology on indirect measures of aboveground biomass in annual grasses. *International Journal of Remote Sensing*, **30**(12), pp.3133–3146.

Cairns, M.A. *et al.*, 1997. Root biomass allocation in the world's upland forests. *Ecologia*, **111(**1), pp.1–11.

Cardillo, M., Macdonald, D.W. & Rushton, S.P., 1999. Predicting mammal species richness and distributions: testing the effectiveness of satellite-derived land cover data. *Landscape Ecology*, **14**(5), pp.423–435.

Chen, L. *et al.*, 2011. Estimation of monthly air-sea CO2 flux in the southern Atlantic and Indian Ocean using in-situ and remotely sensed data. *Remote Sensing of Environment*, **115**(8), pp.1935–1941

Clark, M., Roberts, D. & Clark, D., 2005. Hyperspectral discrimination of tropical rain forest tree species at leaf to crown scales. *Remote Sensing of Environment*, **96**(3-4), pp.375–398.

Collen, B. et al. eds., 2013a. Biodiversity Monitoring and Conservation, Oxford, UK: Wiley-Blackwell.

Collen, B. et al., 2013b. Biodiversity Monitoring and Conservation: Bridging the Gap Between Global Commitment and Local Action (Conservation Science and Practice), Wiley-Blackwell.

Collin, A., Long, B. & Archambault, P., 2012. Merging land-marine realms: Spatial patterns of seamless coastal habitats using a multispectral LiDAR. *Remote Sensing of Environment*, **123**, pp.390–399.

Coops, N.C., *et al*. 2008. The development of a Canadian dynamic habitat index using multi-temporal satellite estimates of canopy light absorbance. Ecological Indicators. **8**: 754-766

Coops, N.C., Wulder, M.A., and Iwanicka, D., 2009. Development of A Satellite-Based Methodology To Monitor Habitat at a Continental-Scale. Ecological Indicators: **9**: 948-958

Corbane, C. *et al.*, 2010. A complete processing chain for ship detection using optical satellite imagery. *International Journal of Remote Sensing*, **31**(22), pp.5837–5854.

Costa, B.M., Battista, T.A. & Pittman, S.J., 2009. Comparative evaluation of airborne LiDAR and shipbased multibeam SoNAR bathymetry and intensity for mapping coral reef ecosystems. *Remote Sensing of Environment*, **113**(5), pp.1082–1100.

DeFries, R. *et al.*, 2005. Increasing isolation of protected areas in tropical forests over the past twenty years. *Ecological Applications*, **15**(1), pp.19–26.

DeFries, R.S., Townshend, J.R.G. & Hansen, M.C., 1999. Continuous fields of vegetation characteristics at the global scale at 1-km resolution. *Journal of Geophysical Research*, **104**(D14), p.16911-16923.

Dekker, A.G., Brando, V.E. & Anstee, J.M., 2005. Retrospective seagrass change detection in a shallow coastal tidal Australian lake. *Remote Sensing of Environment*, **97**(4), pp.415–433.

Diner, D.J. *et al.*, 1999. New Directions in Earth Observing: Scientific Applications of Multiangle Remote Sensing. *Bulletin of the American Meteorological Society*, **80**(11), pp.2209–2228.

Doney, S.C. *et al.*, 2009. Ocean acidification : a critical emerging problem for the ocean sciences. *Oceanography*, **22**(4), pp. 16-25

Dong, J. *et al.*, 2003. Remote sensing estimates of boreal and temperate forest woody biomass: carbon pools, sources, and sinks. USDA Forest Service. UNL Faculty Publications. Paper 43.

Dorschel, B., *et al.*, 2009. Cold-water coral mounds in an erosive environmental setting: TOBI side-scan sonar data and ROV video footage from the northwest Porcupine Bank, NE Atlantic, *Marine Geology* **264**(3–4): 218-229

Dozier, J., 1981. A method for satellite identification of surface temperature fields of subpixel resolution. *Remote Sensing of Environment*, 11, pp.221–229.

Driver, A., *et al.* 2011. An assessment of South Africa's biodiversity and ecosystems. National Biodiversity Assessment. South African National Biodiversity Institute and Department of Environmental Affairs, Pretoria

Druon, J.-N., 2010. Habitat mapping of the Atlantic bluefin tuna derived from satellite data: Its potential as a tool for the sustainable management of pelagic fisheries. *Marine Policy*, **34**(2), pp.293–297.

Dubois, G., *et al.*, 2011. On the contribution of remote sensing to DOPA, a Digital Observatory for Protected Areas. In: "Proceedings of the 34th International Symposium on Remote Sensing of Environment", April 10-15, 2011, Sydney, Australia

Dubuis, A. *et al.*, 2011. Predicting spatial patterns of plant species richness: a comparison of direct macroecological and species stacking modelling approaches. *Diversity and Distributions*, **17**(6), pp.1122–1131.

Dunford, R. *et al.*, 2009. Potential and constraints of Unmanned Aerial Vehicle technology for the characterization of Mediterranean riparian forest. *International Journal of Remote Sensing*, **30**(19), pp.4915–4935.

Dunn, E. & Ralph, C. J., 2002. Use of mist nets as a tool for bird population monitoring, Studies in Avian Biology, 29:1–6.

Duro, D.C. *et al.*, 2007. Development of a large area biodiversity monitoring system driven by remote sensing. *Progress in Physical Geography*, **31**(3), pp.235–260.

Eiss, M.W. *et al.*, 2000. Investigation of a model inversion technique to estimate canopy biophysical variables from spectral and directional reflectance data, 20, pp.3–22.

Elvidge, C.D. *et al.*, Mapping city lights with nighttime data from the DMSP Operational Linescan System. *Photogrammetric engineering and remote sensing*, **63**(6), pp.727–734.

Engelhardt, F.R., 1999. Remote Sensing for Oil Spill Detection and Response. *Pure and Applied Chemistry*, **71**(1), pp.103–111.

Fabry, V.J. *et al.*, 2008. Impacts of ocean acidification on marine fauna and ecosystem processes. *ICES Journal of Marine Science*, **65**(3), pp.414–432.

Fritz, S. *et al.*, 2011. Highlighting continued uncertainty in global land cover maps for the user community. *Environmental Research Letters*, **6**(4), pp.6. doi: 10.1088/1748-9326/6/4/044005

Fretwell, P. T., & Trathan, P. N. (2009). Penguins from space: faecal stains reveal the location of emperor penguin colonies. *Global Ecology and Biogeography*, **18**(5), 543–552. doi:10.1111/j.1466-8238.2009.00467.x

Fretwell, P. T. *et al.*, 2012. An Emperor Penguin Population Estimate: the first global, synoptic survey of a species from space. PLoS ONE 7 (4):e33751. doi: 10.1371/journal.pone.0033751

Fuller, D.O., 2005. Remote detection of invasive Melaleuca trees (Melaleuca quinquenervia) in South Florida with multispectral IKONOS imagery. *International Journal of Remote Sensing*, **26**(5), pp.1057–1063.

Fuller, D.O., 2006. Tropical forest monitoring and remote sensing: A new era of transparency in forest governance? *Singapore Journal of Tropical Geography*, **27**(1), pp.15–29.

Gahegan, M. & Ehlers, M., 2000. A framework for the modelling of uncertainty between remote sensing and geographic information systems. *ISPRS Journal of Photogrammetry and Remote Sensing*, **55**(3), pp.176–188.

GEO BON (2011). Adequacy of Biodiversity Observation Systems to support the CBD 2020 Targets. Pretoria, South Africa.

Geldmann, J. *et al.*, 2013. Effectiveness of terrestrial protected areas in reducing habitat loss and population declines. *Biological Conservation* 161: 230–238

Gledhill, D.K., Wanninkhof, R. & Eakin, C.M., 2009. Observing Ocean Acidification from Space. *Oceanography*, **22**(4), pp.48-59

Gobron, N. *et al.*, 2006. Monitoring the photosynthetic activity of vegetation from remote sensing data. *Advances in Space Research*, **38**(10), pp.2196–2202.

Goward, S.N. *et al.*, 2003. Acquisition of Earth Science Remote Sensing Observations from Commercial Sources: Lessons Learned from the Space Imaging IKONOS Example. Remote sensing of environment, **88**(2), pp.209-219

Green, R.E. *et al.* (2011) What do conservation practitioners want from remote sensing? Cambridge Conservation Initiative Report, Cambridge, UK.

Groombridge, B., 2002. World Atlas of Biodiversity: Earth's Living Resources in the 21st Century, University of California Press.

Guildford, J. and Palmer, M. 2008. *Mulitple Applications of Bathymmetric Lidar*. Proceedings of the Canadian Hydrographic Conference and National Surveyors Conference 2008.

Haines-Young, R. et al., Modelling natural capital: The case of landscape restoration on the South Downs, England [An article from: Landscape and Urban Planning], Elsevier.

Hansen, M.C. *et al.*, 2003. Global Percent Tree Cover at a Spatial Resolution of 500 Meters: First Results of the MODIS Vegetation Continuous Fields Algorithm. *Earth Interactions*, **7**(10), pp.1–15.

Hansen, M.C. & Loveland, T.R., 2012. A review of large area monitoring of land cover change using Landsat data. *Remote Sensing of Environment*, **122**, pp.66–74.

He, K.S. *et al.*, 2011. Benefits of hyperspectral remote sensing for tracking plant invasions. *Diversity and Distributions*, **17**(3), pp.381–392.

He, X. *et al.*, 2013. Using geostationary satellite ocean color data to map the diurnal dynamics of suspended particulate matter in coastal waters. *Remote Sensing of Environment*, **133**, pp.225–239.

Hestir, E.L. *et al.*, 2008. Identification of invasive vegetation using hyperspectral remote sensing in the California Delta ecosystem. *Remote Sensing of Environment*, **112**(11), pp.4034–4047.

Hill, D. *et al.*, 2005. *Handbook of Biodiversity Methods: Survey, Evaluation and Monitoring*, Cambridge University Press.

Holmes, K.R., *et al.*, 2013. Biodiversity indicators show climate change will alter vegetation in parks and protected areas. Diversity **5**(2): 352-353. doi: http://dx.doi.org/10.3390/d5020352

Huete, A. *et al.*, 2002. Overview of the radiometric and biophysical performance of the MODIS vegetation indices. *Remote Sensing of Environment*, **83**(1-2), pp.195–213.

Hyrenbach, K.D. *et al.*, 2007. Community structure across a large-scale ocean productivity gradient: Marine bird assemblages of the Southern Indian Ocean. *Deep Sea Research Part I: Oceanographic Research Papers*, **54**(7), pp.1129–1145.

Infoterra (2007). Development of 'Land' Earth Observation Requirements as Input to the Defra Earth Observation Strategy - Final Report. October 2007.

Integrated Ocean Observing System (IOOS), 2013. Available at: <u>http://www.ioos.noaa.gov/observing/animal\_telemetry/</u> [Accessed August 23, 2013].

IPCC, 2007: Summary for Policymakers. In: *Climate Change 2007: The Physical Science Basis. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.

Joint WMO-IOC Technical Commission for Oceanography and Marine Meteorology (JCOMM). Available at: <u>http://www.wmo.int/pages/prog/amp/mmop/jcomm\_partnership\_en.html</u> [Accessed August 13, 2013].

Johnson, C.R. *et al.*, 2011. Climate change cascades: Shifts in oceanography, species' ranges and subtidal marine community dynamics in eastern Tasmania. *Journal of Experimental Marine Biology and Ecology*, **400**(1-2), pp.17–32.

Jones, D.A. *et al.*, 2009. Monitoring land use and cover around parks: A conceptual approach. *Remote Sensing of Environment*, **113**(7), pp.1346–1356.

Jürgens, N. *et al.*, 2012. The BIOTA Biodiversity Observatories in Africa--a standardized framework for large-scale environmental monitoring. *Environmental monitoring and assessment*, **184**(2), pp.655–78.

Kachelriess, D. *et al.*, 2014. The application of remote sensing for marine protected area management. *Ecological Indicators*, **36**, pp.169–177.

Kalko, E.K. V *et al.*, 2008. Flying high--assessing the use of the aerosphere by bats. *Integrative and comparative biology*, **48**(1), pp.60–73.

Kalma, J.D., McVicar, T.R., McCabe, M.F. 2008. Estimating Land Surface Evaporation: A Review of Methods Using Remotely Sensed Surface Temperature Data. *Surveys in Geophysics* **29**:421–469

Kennedy, R.E. *et al.*, 2009. Remote sensing change detection tools for natural resource managers: Understanding concepts and tradeoffs in the design of landscape monitoring projects. *Remote Sensing of Environment*, **113**(7), pp.1382–1396.

Kennedy, R.E., Cohen, W.B. & Takao, G., 1997. Empirical methods to compensate for a view-angledependent brightness gradient in AVIRIS imagery. *Remote Sensing of Environment*, **62**(3), pp.277–291.

Kerr, J.T. & Ostrovsky, M., 2003. From space to species: ecological applications for remote sensing. *Trends in Ecology & Evolution*, **18**(6), pp.299–305.

Klemas, V., 2011. Remote Sensing Techniques for Studying Coastal Ecosystems: An Overview. *Journal of Coastal Research*, **27**, pp.2–17.

Kokaly, R.F. *et al.*, 2007. Characterization of post-fire surface cover, soils, and burn severity at the Cerro Grande Fire, New Mexico, using hyperspectral and multispectral remote sensing. *Remote Sensing of Environment*, **106** (3), pp.305–325.

Krishnaswamy, J. *et al.*, 2009. Quantifying and mapping biodiversity and ecosystem services: Utility of a multi-season NDVI based Mahalanobis distance surrogate. *Remote Sensing of Environment*, **113**(4), pp.857–867.

Kunkel, K.E., 2004. Temporal variations in frost-free season in the United States: 1895–2000. *Geophysical Research Letters*, **31**(3), p.L03201.

Langdon, C. *et al.*, 2000. Effect of calcium carbonate saturation state on the calcification rate of an experimental coral reef. *Global Biogeochemical Cycles*, **14**(2), pp.639–654.

Lapointe, N.W.R., Corkum, L.D. & Mandrak, N.E., 2006. A Comparison of Methods for Sampling Fish Diversity in Shallow Offshore Waters of Large Rivers. *North American Journal of Fisheries Management*, **26**(3), pp.503–513.

Larsen, R.J. *et al.*, 2007. Mist netting bias, species accumulation curves, and the rediscovery of two bats on Montserrat (Lesser Antilles). *Acta Chiropterologica*, **9**(2), pp.423-435.

Laurance, W.F. *et al.*, 2012. Averting biodiversity collapse in tropical forest protected areas. *Nature*, **489**(7415), pp.290–4.

Le Toan, T. *et al.*, 2011. The BIOMASS mission: Mapping global forest biomass to better understand the terrestrial carbon cycle. *Remote Sensing of Environment*, **115**(11), pp.2850–2860.

Leidner, A.. *et al.*, 2012. Satellite Remote Sensing for Biodiversity Research and Conservation Applications: A Committee on Earth Observation Satellites (CEOS) Workshop. German Aerospace Center (DLR-EOC).

Leifer, I. *et al.*, 2012. State of the art satellite and airborne marine oil spill remote sensing: Application to the BP Deepwater Horizon oil spill. *Remote Sensing of Environment*, **124**, pp.185–209.

Lengyel, S.,*et al.*, 2008. Habitat monitoring in Europe: a description of current practices. Biodiversity and Conservation **17**, 3327–3339.

Lillesand, T., Kiefer, R.W. & Chipman, J., 2008. *Remote Sensing and Image Interpretation*. John Wiley and Sons, New York, 763 p.

Liu, S., Liu, R. & Liu, Y., 2010. Spatial and temporal variation of global LAI during 1981–2006. *Journal of Geographical Sciences*, **20**(3), pp.323–332.

Löffler, E., Margules, C. 1980. Wombats detected from Space. Remote sensing of environment 9: 47-56

Lucas, R. *et al.*, 2008. Classification of Australian forest communities using aerial photography, CASI and HyMap data. *Remote Sensing of Environment*, **112**(5), pp.2088–2103.

Lucas, R. *et al.*, 2011. Updating the Phase 1 habitat map of Wales, UK, using satellite sensor data. *ISPRS Journal of Photogrammetry and Remote Sensing*, **66**(1), pp.81–102.

Malthus, T.J. & Mumby, P.J., 2003. Remote sensing of the coastal zone: An overview and priorities for future research. *International Journal of Remote Sensing*, **24**(13), pp.2805–2815.

Margaret Kalacska, G. Arturo Sanchez-Azofeifa (2008). Hyperspectral rmote sensing of tropical and sub-tropical forests. CRC Press. 352p.

Maynard, J.A. *et al.*, 2008. ReefTemp: An interactive monitoring system for coral bleaching using highresolution SST and improved stress predictors. *Geophysical Research Letters*, **35**(5), L05603

Mazerolle, M.J. *et al.*, Making Great Leaps Forward : Accounting for Detectability in Herpetological Field Studies. *Journal of herpetology*, **41**(4), pp.672–689.

McCallum, I. *et al.*, 2009. Satellite-based terrestrial production efficiency modeling. *Carbon balance and management*, **4**:8.

McNair, G. 2010. *Coastal Zone Mapping with Airborne Lidar Bathymetry*. Masters Thesis. Department of Mathematical Sciences and Technology, Norwegian University of Life Sciences, Ås, Norway.

Metzger, M.J. *et al.*, 2006. The vulnerability of ecosystem services to land use change. *Agriculture, Ecosystems & Environment*, **114**(1), pp.69–85.

Migliavacca, M. *et al.*, 2011. Using digital repeat photography and eddy covariance data to model grassland phenology and photosynthetic CO2 uptake. *Agricultural and Forest Meteorology*, **151**(10), pp.1325–1337.

Movebank, 2013. Movebank for Animal Tracking data. Available at: <u>https://www.movebank.org/</u> [Accessed August 9, 2013].

Mucina, L., Rutherford, M.C. & Powrie, L.W., 2006. Strelitzia 19: Vegetation of South Africa, Lesotho & Swaziland (2 CD set). SANBI. Available at: <u>http://www.sanbi.org/documents/strelitzia-19-vegetation-south-africa-lesotho-swaziland-2-cd-set</u>.

Mulligan, M (2006) Global Gridded 1km TRMM Rainfall Climatology and Derivatives.Version 1.0. Database: http://www.ambiotek.com/1kmrainfall

Murphy, R.J. *et al.*, 2008. Field-based remote-sensing for experimental intertidal ecology: Case studies using hyperspatial and hyperspectral data for New South Wales (Australia). *Remote Sensing of Environment*, **112**(8), pp.3353–3365.

Myeong, S., Nowak, D.J. & Duggin, M.J., 2008. A temporal analysis of urban forest carbon storage using remote sensing. Remote Sensing of Environment **101** (2006), pp. 277-282.

Nagai, S. *et al.*, 2013. Utility of information in photographs taken upwards from the floor of closed-canopy deciduous broadleaved and closed-canopy evergreen coniferous forests for continuous observation of canopy phenology. *Ecological Informatics*, **18**, pp.10–19.

Nagendra, H. *et al.*, 2013. Remote sensing for conservation monitoring: Assessing protected areas, habitat extent, habitat condition, species diversity, and threats. *Ecological Indicators*, **33**, pp.45–59.

Nagendra, H., 2001. Using remote sensing to assess biodiversity. *International Journal of Remote Sensing*, **22**(12), pp.2377–2400.

Nagendra, H. & Rocchini, D., 2008. High resolution satellite imagery for tropical biodiversity studies: the devil is in the detail. *Biodiversity and Conservation*, **17**(14), pp.3431–3442.

OAPS, 2013. Ocean Acidification Product Suite (OAPS) (Version 0.6) (Experimental Product, Monthly Update). Available at: <u>http://coralreefwatch.noaa.gov/satellite/oa/index.php</u> [Accessed August 23, 2013e]

Oindo, B.O., Skidmore, A.K. (2002) Interannual variability of NDVI and species richness in Kenya. *International Journal of Remote Sensing* **23**:285-298

Oney, B., Shapiro, A. & Wegmann, M., 2011. Evolution of water quality around the Island of Borneo during the last 8-years. *Procedia Environmental Sciences*, **7**, pp.200–205.

Paarmann, W. & Stork, N.E., 1987. Canopy fogging, a method of collecting living insects for investigations of life history strategies. *Journal of Natural History*, **21**(3), pp.563–566.

Patenaude, G. *et al.*, 2004. Quantifying forest above ground carbon content using LiDAR remote sensing. *Remote Sensing of Environment*, **93**(3), pp.368–380.

Pereira, H.M. *et al.*, 2013. Ecology. Essential biodiversity variables. *Science (New York, N.Y.)*, **339**(6117), pp.277–8.

Petersen, S.L. *et al.*, 2008. Albatross overlap with fisheries in the Benguela Upwelling System: implications for conservation and management. *Endangered Species Research*, **5**, pp. 117-127.

Campbell, J.B. & Wynne, R.H., 2006. Introduction to Remote Sensing. 4<sup>th</sup> edition, The Guilford Press.

PHOENIX, G.K. *et al.*, 2006. Atmospheric nitrogen deposition in world biodiversity hotspots: the need for a greater global perspective in assessing N deposition impacts. *Global Change Biology*, **12**(3), pp.470–476.

Pinty, B. & Verstraete, M.M., 1992. GEMI: a non-linear index to monitor global vegetation from satellites. *Vegetatio*, **101**(1), pp.15–20.

Powers, R.P., *et al.*, 2013. Integrating accessibility and intactness into large-area conservation planning in the Canadian boreal. Biological Conservation (in press)

Purves, D. et al., 2013. Ecosystems: Time to model all life on Earth. Nature, 493(7432), pp.295–7.

Queiroz, N. *et al.*, 2012. Spatial dynamics and expanded vertical niche of blue sharks in oceanographic fronts reveal habitat targets for conservation. Y. Ropert-Coudert, ed. *PloS one*, **7**(2): e32374. doi: 10.1371/journal.pone.0032374

Myneni, R.B, *et al.* 2002. Global Products Of Vegetation Leaf Area And Fraction Absorbed Par From Year One Of Modis Data. NASA *Publications*. Paper 39.

Raes, N. *et al.* 2009. Botanical Richness and endemicity patterns of Borneo derived from species distribution models. Ecography **32**(1), pp.180-192.

Ramsey III, E. *et al.*, 2005. Mapping the invasive species, Chinese tallow, with EO1 satellite Hyperion hyperspectral image data and relating tallow occurrences to a classified Landsat Thematic Mapper land cover map. *International Journal of Remote Sensing*, **26**(8), pp.1637–1657.

Rangama, Y., 2005. Variability of the net air–sea CO 2 flux inferred from shipboard and satellite measurements in the Southern Ocean south of Tasmania and New Zealand. *Journal of Geophysical Research*, **110**(C9): C09005. doi: 10.1029/2004JC002619

Redford, K.H. 1992. The Empty Forest. BioScience 42 (6): 412-422

Ribeiro-Júnior, M.A., Gardner, T.A. & Ávila-Pires, T.C.., 2008. Evaluating the Effectiveness of Herpetofaunal Sampling Techniques across a Gradient of Habitat Change in a Tropical Forest Landscape. *Journal of Herpetology*, **42**(4), p.733.

Rocchini, D. *et al.*, 2010. Remotely sensed spectral heterogeneity as a proxy of species diversity: Recent advances and open challenges. *Ecological Informatics*, **5**(5), pp.318–329.

Rocchini, D., Chiarucci, A. & Loiselle, S.A., 2004. Testing the spectral variation hypothesis by using satellite multispectral images. *Acta Oecologica*, **26**(2), pp.117–120.

Rodell, M., Velicogna, I. & Famiglietti, J.S., 2009. Satellite-based estimates of groundwater depletion in India. *Nature*, **460**(7258), pp.999–1002. Available at: http://dx.doi.org/10.1038/nature08238.

Rowlands, G. *et al.*, 2012. Satellite imaging coral reef resilience at regional scale. A case-study from Saudi Arabia. *Marine pollution bulletin*, **64**(6), pp.1222–37.

Roy, D.P. et al., 2010. Accessing free Landsat data via the Internet: Africa's challenge. Remote Sensing Letters, 1(2), pp.111–117.

Roy, D.P. *et al.*, 2005. Prototyping a global algorithm for systematic fire-affected area mapping using MODIS time series data. *Remote Sensing of Environment*, **97**(2), pp.137–162.

Roy, P.S. & Saran, S., 2004. Biodiversity Information System for North East India. *Geocarto International*, **19**(3), pp.73–80.

Ruesch, A.S. and Gibbs, H. 2008. New IPCC Tier-1 Global Biomass Carbon Map for the Year 2000. Available online from the Carbon Dioxide Information Analysis Center [<u>http://cdiac.ornl.gov/]</u>, Oak Ridge National Laboratory's Carbon Dioxide Information Analysis Center, Tennessee, USA.

Saatchi, S.S. *et al.*, 2011. Benchmark map of forest carbon stocks in tropical regions across three continents. *Proceedings of the National Academy of Sciences of the United States of America*, **108**(24), pp.899–904.

Saatchi, S.S. *et al.*, 2007. Distribution of aboveground live biomass in the Amazon basin. *Global Change Biology*, **13**(4), pp.816–837.

Sam J. Purkis, Victor V. Klemas, 2011. Remote Sensing and Global Environmental Change. 384p. Wiley-Blackwell

Sandmeier, S. *et al.*, 1998. Physical Mechanisms in Hyperspectral BRDF Data of Grass and Watercress. *Remote Sensing of Environment*, **66**(2), pp.222–233.

Sausen, T.M., 2000. Space education in developing countries in the information era, regional reality and new educational material tendencies: example, South America. *ISPRS Journal of Photogrammetry and Remote Sensing*, **55**(2), pp.129–135.

Scales, K.L. *et al.*, 2011. Insights into habitat utilisation of the hawksbill turtle, Eretmochelys imbricata (Linnaeus, 1766), using acoustic telemetry. *Journal of Experimental Marine Biology and Ecology*, **407**(1), pp.122–129.

Scholes, R.J. *et al.*, 2012. Building a global observing system for biodiversity. *Current Opinion in Environmental Sustainability*, **4**(1), pp.139–146.

Schubert, P. et al., 2010. Estimating northern peatland CO2 exchange from MODIS time series data. *Remote Sensing of Environment*, **114**(6), pp.1178–1189.

Scott, J.M. & Jennings, M.D., Large-Area Mapping Of Biodiversity . *Annals of the Missouri Botanical Garden*, **85**(1), pp.34–47.

Sequeira, A. *et al.*, 2012. Ocean-scale prediction of whale shark distribution. *Diversity and Distributions*, **18**(5), pp.504–518.

Sewell, D. *et al.*, 2012. When is a species declining? Optimizing survey effort to detect population changes in reptiles. B. Fenton, ed. *PloS one*, **7**(8), p.e43387

Sexton, J.O. *et al.*, 2013. Global, 30-m resolution continuous fields of tree cover: Landsat-based rescaling of MODIS vegetation continuous fields with lidar-based estimates of error. *International Journal of Digital Earth*, **6**(5), pp.1–22.

Sheppard, C. & Rayner, N.A., 2002. Utility of the Hadley Centre sea ice and sea surface temperature data set (HadISST1) in two widely contrasting coral reef areas. *Marine pollution bulletin*, **44**(4), pp.303–308.

Siliang,L., Ronggao L., and Yang, L. 2010. Spatial and temporal variation of global LAI during 1981–2006. *Journal of Geographical Sciences* **20** (3): 323-332.

Sonnentag, O. *et al.*, 2012. Digital repeat photography for phenological research in forest ecosystems. *Agricultural and Forest Meteorology*, **152**, pp.159–177.

Stagakis, S. *et al.*, 2012. Monitoring water stress and fruit quality in an orange orchard under regulated deficit irrigation using narrow-band structural and physiological remote sensing indices. *ISPRS Journal of Photogrammetry and Remote Sensing*, **71**, pp.47–61.

Strand, H. et al., 2007. Sourcebook on Remote Sensing and Biodiversity Indicators. Convention on Biological Diversity Technical Series 32, CBD. Montreal, Canada

Sung, Y.-H., Karraker, N.E. & Hau, C.. H., 2011. Evaluation Of The Effectiveness Of Three Survey Mehtods For Sampling Terrestrial Herpetofauna In South China., **6**(7), pp.479–489.

Sutton, P.C. & Costanza, R., 2002. Global estimates of market and non-market values derived from nighttime satellite imagery, land cover, and ecosystem service valuation. *Ecological Economics*, **41**(3), pp.509–527.

Swetnam, R.D. *et al.*, Lewis,2011. Mapping socio-economic scenarios of land cover change: A GIS method to enable ecosystem service modelling, *Journal of Environmental Management* **92** (3):563-574.

Szantoi, Z., *et al.*, 2013. Wetland Composition Analysis Using Very High Resolution Images and Texture Features. International Journal of Applied Earth Observation and Geoinformation **23**(8):204-212

T.I. Bern, T.W., 1993. Oil spill detection using satellite based SAR - Experience from a field experiment. *Photogrammetric Engineering and Remote Sensing*, **2**, 59:3.

Takahashi, T. *et al.*, 2002. Global sea–air CO2 flux based on climatological surface ocean pCO2, and seasonal biological and temperature effects. *Deep Sea Research Part II: Topical Studies in Oceanography*, **49**(9-10), pp.1601–1622..

Thackeray, S.J. *et al.*, 2010. Trophic level asynchrony in rates of phenological change for marine, freshwater and terrestrial environments. *Global Change Biology*, **16**(12), pp.3304–3313.

Tropical Rainfall Measuring MIssion (TRMM), 2013. Available at http://trmm.gsfc.nasa.gov/ [Accessed September 5, 2013]. Last Updated September 3, 2013.

Tynan, C.T. *et al.*, 2005. Cetacean distributions relative to ocean processes in the northern California Current System. *Deep Sea Research Part II: Topical Studies in Oceanography*, **52**(1-2), pp.145–167.

USGS, 2008. 2008-2018 USGS Africa Remote Sensing Study, Aerial and Spaceborne Ten-Year Trends. Available at: <u>http://www.globalinsights.com/USGS2008AfricaRSS.pdf</u> [Accessed August 19, 2013].

Velasco, M. (2009) A quickbird's-eye view on marmot. MSc Thesis, ITC, Enschede, The Netherlands. 51 pp

Watts, A. C., et al., 2010. Small unmanned aircraft systems for low-altitude aerial surveys. Journal of Wildlife Management **74**(7):1614–1619

Wilkie, D.S., Bennett, E.L., Peres, C.A., Cunningham, A.A., 2011. The empty forest revisited. *Ann. N.Y. Acad. Sci.* **1223**: 120–128.

Wingfield, D.K. *et al.*, 2011. The making of a productivity hotspot in the coastal ocean. A. Chiaradia, ed. *PloS one*, **6**(11), p.e27874.

Yamano, H. & Tamura, M., 2004. Detection limits of coral reef bleaching by satellite remote sensing: Simulation and data analysis. *Remote Sensing of Environment*, **90**(1), pp.86–103.

Yang, Z. Evaluating high resolution GeoEye-1 Satellite imagery for mapping wildlife in open savannahs. MSc Thesis, ITC, Enschede, The Netherlands. 61 pp.

Yanoviak, S.P., Nadkarni, N.M. & Gering, J.C., 2003. Arthropods in epiphytes: a diversity component that is not effectively sampled by canopy fogging. *Biodiversity & Conservation*, **12**(4), pp.731–741.

Zarco-Tejada, P.J., González-Dugo, V. & Berni, J.A.J., 2012. Fluorescence, temperature and narrowband indices acquired from a UAV platform for water stress detection using a micro-hyperspectral imager and a thermal camera. *Remote Sensing of Environment*, **117**, pp.322–337.

Zhang, J. *et al.*, 2006. Intra- and inter-class spectral variability of tropical tree species at La Selva, Costa Rica: Implications for species identification using HYDICE imagery. *Remote Sensing of Environment*, **105**(2), pp.129–141.

Zhang, Y.-H. *et al.*, 2003. Monthly burned area and forest fire carbon emission estimates for the Russian Federation from SPOT VGT. *Remote Sensing of Environment*, **87**(1), pp.1–15.

# Annex 1. The basics of remote sensing in biodiversity monitoring

# 1.1 What is remote sensing?

There are many possible definitions of the term Remote Sensing. Remote means away from or at a distance and sensing means detecting a property or characteristics. Therefore, Remote Sensing could be very broadly defined as the science of collecting and interpreting information about the Earth's surface without actually being in contact with it.

Remote sensing can be classified according to the vehicle or carrier (called platform) by which remotes sensors are borne. According to the height of platforms, remote sensing can be classified into three levels:

Level	Operational range	Height	Pros
Ground	Short range	50-100 m	-Panoramic mapping -Millimeter accuracies -High definition surveying
	Medium range	150-250m	
	Long range	Up to 1km	-
Airborne	Aircraft	Up to 20km	<ul> <li>Last minutes timing changes can be made to adjust for illumination from the sun, the location of the area to be visited and additional revisits to that location.</li> <li>Sensor maintenance, repair and configuration changes are easily made to aircraft platforms. Aircraft flight paths know no boundaries except political boundaries</li> <li>Quantitative measurement of ground features using radiometrically calibrated sensors</li> <li>Semi-automated computerized processing and analysis</li> <li>Unique way of covering a broad range of altitudes for in-situ or remote sensing measurements in the stratosphere</li> <li>Opportunity for additional, correlative data for satellite based measurements, including both validation and complementary data</li> <li>Important and inexpensive venue for testing instruments under development.</li> </ul>
	Balloon based	Up to 40 km	
Spaceborne	Space shuttle	250-300km	<ul> <li>Large area coverage</li> <li>Frequent and repetitive coverage of an area of interest</li> <li>Quantitative measurement of ground features using radiometrically calibrated sensors</li> <li>Semi-automated computerized processing and analysis</li> </ul>
	Space stations	300-400 km	
	Low level satellites	700-1500 km	
	Geostationary satellites	36000 km	

Table 1.1. Remote sensing classification according to the height of sensor-borne platforms

Aircraft based airborne remote sensing can be further categorized to manned aerial vehicle remote sensing and UAV remote sensing according to the platform. The name UAV covers all vehicles which are flying in the air with no person onboard with the capability of controlling the aircraft. Thanks to GPS and communication technology, UAVs can be remotely controlled or flown autonomously based on pre-programmed flight plans or more complex dynamic automation systems. The benefits of UAVs mainly lie in the ease, rapidity and cost of flexibility of deployment that lends itself to many land surface measurement and monitoring applications, especially those requiring access to higher altitudes and longer times on station (i.e., longer flight times). Although conventional airborne remote sensing has some drawbacks, such as altitude, endurance, attitude control, all-weather operations, and monitoring of the dynamics, it is still an important technique of studying and exploring the Earth's resources and environment.

#### 1.2 An overview of remote sensing sources and applicability for monitoring biodiversity

Remote sensing systems can be classified in two major groups: passive and active sensors. The following pages contain a brief and simple description for each system, which is adopted throughout this review. The more technical aspects, as well as detailed discussion of advantages and drawbacks of each sensor have not been included since it is not the nature of this report to provide this level of technical information, which can be easily found in the available literature.

#### **1.2.1** Passive remote sensing

Remote sensing systems which measure energy that is naturally available are called passive sensors. The way to use passive sensors to examine, measure and analyse an object is called **passive remote sensing or optical remote sensing**. Measurable energy takes the form of electromagnetic radiation from a surface, either as a reflection (reflected light) or as an emission (radiation emitted from the surface itself). For all reflected energy, this can only take place during the time when the sun is illuminating the Earth as there is no reflected energy available from the sun at night. Energy that is naturally emitted (such as thermal infrared) can be detected day or night.

Optical remote sensing is based on different areas of light's spectrum. For example, theVisible spectrum (VIS) is the portion of the electromagnetic spectrum from about 0.39 to 0.7  $\mu$ m that is visible to the human eye. The VIS is often displayed through the use of three spectral bands:: Blue band (0.45-0.515  $\mu$ m) is used for atmospheric and deep water imaging, and can reach up to 50m deep in clear water; green band (0.515-0.6  $\mu$ m) is used for imaging of vegetation and deep water structures, up to 30m in clear water; and red band (0.6-0.69  $\mu$ m) is used for imaging of man-made objects, in water up to 9m deep, soil, and vegetation, and it is sensitive to clorophyll. Infrared light occurs at longer wavelengths just below red light, hence the name, infra- (below) red. Near-infrared spectrum (NIR) ranges from about 0.7 to 1.1  $\mu$ m that lies just out of the human vision, which is used primarily for imaging of vegetation. The NIR can be used to discriminate plant species. Short-wave infrared (SWIR) light is typically defined as light in the 1.1 – 3.0  $\mu$ m wavelength range. One major benefit of SWIR imaging is the ability to image through haze, fog and glass. The SWIR are known to be very sensitive to leaf water content (Tucker, 1980), which therefore can enhance plant species identification. Mid-wave infrared spectrum (MWIR) ranges from about 3.0 to 5.5  $\mu$ m and thermal infrared (TIR) ranges from 8 to 14  $\mu$ m. Both MWIR and TIR imaging can capture the intrinsic heat radiated by objects (i.e., the objects'

thermal emission): warm objects stand out well against cooler backgrounds. Warm-blooded animals become easily visible against the environment atnight..

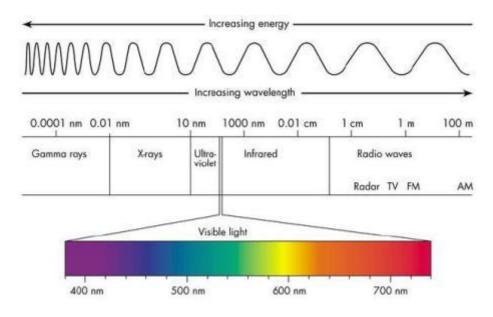


Figure 1.1 Diagram of the light's electromagnetic spectrum, showing the different wavelengths

There are two methods to collect data using passive sensors:

#### **Multispectral**

Multispectral remote sensing collects data in a few relatively wide and noncontiguous spectral bands, typically measured in micrometers or nanometers (1 micrometer = 1000 nanometers) These spectral bands are selected to collect radiation in specifically defined parts of the spectrum and optimized for certain categories of information most evident in those bands.. Different spectral behavior allows detailed classification of specific types of land surfaces (depending on the spatial, spectral and radiometric resolution of the used sensor). The remotely sensed spectral heterogeneity information provides a crucial baseline for rapid estimation or prediction of biodiversity attributes and hotspots in space and time.

#### **Hyperspectral**

Hyperspectral sensors or imaging spectrometers measure energy in many narrow, contiguous bands-often as many as 200 or more.. A reasonable criterion, to be considered in a rather flexible way, is that the hyperspectral remote sensing collects at least 100 spectral bands of 10-20 nm width. The numerous narrow bands of hyperspectral sensors provide a continuous spectral measurement a portion of the electromagnetic spectrum and therefore are more sensitive to subtle variations in reflected energy. Hyperspectral sensors generally contain much more information than images and have a greater potential to detect differences among land and water features. For example, multispectral imagery can be used to map forested areas, while hyperspectral imagery can be used to map tree species within the forest, contingent upon appropriate spatial resolution.

#### **1.2.2** Active remote sensing

Active remote sensing sensors provide their own energy source for illumination. The active sensor emits radiation which is directed toward the target to be investigated. The radiation reflected from that target is detected and measured by the sensor. Using active sensors to examine, measure, and analyze an object is called active remote sensing.. Active sensors can be used for examining wavelengths with insufficient energy provided by the sun, such as microwaves, or to better control the way a target is sensed. Advantages of active microwave sensors include the ability to obtain measurements anytime, regardless of the time of day or season. However, active systems require the generation of a fairly large amount of energy to adequately sense targets.

#### Radar

Radar is an acronym for "Radio detection and ranging", which essentially characterizes the function and operation of a Radar sensor. Radar works by sending out microwave (radio) signals towards the target and detects the backscattered portion of the signal. By measuring the amount of time it takes for the signals to return, it is possible to detect the location, speed, direction and altitude of an object.

For example, the ground-based Radar technology allows us to track bird migration at night. It also serves as a useful tool for the study of bird migration patterns and behaviors, as well as alerting us to any changes in those patterns and behaviors (Liechti *et al.* 1995; Hilgerioh 2001; Ruth *et al.* 2005; Ruth 2007; Gudmundsson 2008). An important advantage to using airborne and spaceborne Radar systems is that they can penetrate thick clouds and moisture, which would not be possible using optical remote sensing. This allows scientists to accurately map areas such as rain forests that are otherwise too obscured by clouds and rain. The high resolution Radar monitoring system is perfectly suitable in support of mapping and monitoring wildlife habitat. The system can provide regular information on the location of changes, such as changes in the forest canopy through logging or landslides, (illegal) clearing of areas (for agriculture, mining, oil palm plantation) and encroachment patterns, expansion of road networks, fire impacts and vegetation development (Bergen *et al.* 2009; Swatantran *et al.* 2012).

#### Lidar

LiDAR stands for "Light Detection And Ranging" and is very similar to the better known Radar. Basically, a laser pulse is sent out of a transmitter and the light particles (photons) are scattered back to the receiver. The photons that come back to the receiver are collected with a photodetectorand counted as a function of time. Using the speed of light we can then calculate how far the photons have traveled round trip.

LiDAR is a remote sensing technology that is now becoming more widespread in ecological research. The metrics derived from airborne or spaceborne LiDAR measurements can be used to infer forest canopy height and/or canopy structure complexity. Its ability to accurately characterize vertical structure makes LiDAR a valuable and cost-effective approach for estimating forest attributes that are related to important ecological characteristics. In this regard, an attribute of particular interest is 3-dimensional habitat heterogeneity, which reflects the variability in both horizontal and vertical forest structure (e.g. stem, branch and foliage density and distribution). This structural variability may be

correlated with species richness and other biodiversity metrics, which are central components to understanding, modeling and mapping patterns of biodiversity (Vierling *et al.* 2008; Bergen *et al.* 2009; Goetz *et al.* 2010).

#### Sonar

Sonar – short for "Sound navigation and ranging" - is a technique that uses sound propagation (usually underwater, as in submarine navigation) to navigate, communicate with or detect objects on or under the surface of the water. Sonar works in a similar manner as Radar. However, instead of sending out radio waves, Sonar sensors send out sound waves. By measuring the time it takes for these sound waves to travel towards an object, bounce off of it, and then return, it is possible to calculate distances.

Two types of technology share the name "Sonar": passive Sonar is essentially listening for the sound made by vessels; and, active Sonar which emitts pulses of sounds and listening for echoes. Sonar sensing may be used as a means of acoustic location and of measurement of the echo characteristics of *targets* in the water. Active Sonar allows scientists to accurately map the two thirds of the Earth that is under water. In addition, Active Sonar has been used to investigate the population dynamics of both deep and shallow water fish populations. Passive Sonar sensors that receive underwater sounds help overcome many of the limitations experienced with visual surveys.

Both passive and active Sonar have been incorporated into survey methods to improve animal abundance estimates, especially for cetacean surveys. For example, passive Sonar sensors have successfully been used in abundance estimates for several cetacean species including right whales, beaked whales, sperm whales, humpback dolphins, and finless porpoises (Akamatsu *et al.* 2001; Van Parijs *et al.* 2002; Barlow *et al.* 2005; Wade *et al.* 2006; Mellinger *et al.* 2007; Clark *et al.* 2010). The use of passive Sonar sensors may allow for more animal detections across larger ranges than would be obtained from visual methods alone, and facilitate the detection of animals that spend a large amount of time under water.

#### 1.3 How to use remote sensing to monitor biodiversity?

There are several approaches possible to use remote sensing to monitor biodiversity. Which approach is most suitable depends on the environment in which biodiversity is to be monitored; the characteristics of relevant species that occur in these ecosystems and the availability of remote sensing data. Twomajor approaches can be distinguished:

#### 1.3.1 Direct measurements of individuals and populations

Direct measurements of individuals and populations are possible when high to very high resolution imagery is available, such as RapidEye (5m), WorldView ( $\leq 2m$ ), GeoEye (< 2m), Pleiades (< 1m) or Ikonos (3.2m). A key feature of very high resolution imagery is the ability to detect and classify individual tree canopies. Direct measurement of animal populations is constrained to situations where the animals or their traces (such as burrows) can be easily detected. This means a limited vegetation cover, or a vegetation cover that is less high than the species involved. Examples where this kind of monitoring has been successfully implemented include elephants, wildebeest and zebra in the Serengeti (Yang 2012), marmots in Mongolia (Velasco 2009) or emperor penguins in Antartica (Fretwell *et al.*, 2012).

Already in the 1980's Wombat burrows were identified from medium resolution Landsat MSS imagery (Löffler and Margules 1980). The breeding distribution of the Emperor penguin in Antarctica has been mapped by spectral characterisation of breeding colonies on snow in Landsat imagery (Fretwell & Trathan, 2009).

#### **1.3.2 Indirect proxies of biodiversity**

Indirect proxies involve approaches where derived information from the reflectance values that are recorded by satellite sensors is used to infer information about biodiversity on the surface that was monitored. Such proxies can be based on variability along three potential axes, a spatial, a temporal and a spectral axis. The sensor at hand determines to great extent which proxies can be generated. Sensors with high spatial resolution offer a possibility to look at variability in the reflectance in neighborhoods of small size, i.e. with great detail. But satellite borne sensors of this kind are normally limited in their spectral and temporal dimensions. Likewise, sensors with high temporal resolutions (e.g. NOAA AVHRR or MODIS) are limited in their spectral and spatial resolution. Which combination offers the best solution to monitor biodiversity depends heavily on the ecosystem and target species to be monitored. Recent literature suggest that spectral resolution would be preferred over spatial resolution (Rocchini *et al.* 2010 and references therein). The minimal size of homogeneous units within the system determines to a large extent which pixel size is acceptable. Likewise, the difference in phenology of key species in the system determines whether variation over the temporal axes can help in identifying changes in biodiversity (Oindo and Skidmore 2002).

Indirect proxiescan often be derived from satellite data that have direct biophysical meanings, such as altitude from digital elevation models, green biomass from Normalized Difference Vegetation Index (NDVI) products, vegetation cover, or surface temperature. These data sometimes can have a direct link to diversity (Baldeck *et al.* 2013) and be used as a proxy value. In addition they are often used as explanatory variables in species distribution modeling (SDM), which in turn can be used for species diversity assessments, as described below. Nevertheless, diversity in ancillary data, such as altitude also provides information about species diversity at intermediate scales, because it can represent heterogeneity in available niches (Allouche *et al.* 2012).

### 1.3.2.1 Inputs to Models

Remotely sensed data can also be used as an essential input to several kinds of models that predict diversity, such as Species Distribution Models (SDMs) where empiricial relationships between observed occurrences of species and remotely-sensed environmental conditions are used to extrapolate potential species distributions. These models are often implemented to map the distribution of single species, but they can be also be aggregated to map areas with high probabilities of many species (i.e. hot spots) and few species (i.e. cold spots). Often this does not involve raw satellite reflectance signals, but further refined products such as indirect proxies (see above) that have a logical relationship with species survival such as surface temperature, rainfall data, NDVI or seasonality of NDVI. These are often important parameters for most species that try to find an optimum in a multidimensional optimization of environmental conditions.

Another type of model worth mention in the context of this review is the bottom-up models that describe ecosystem dynamics, from which biodiversity can be inferred. These models, called Dynamic Global Vegetation Models (DGVMs), stimulate changes in potential vegetation and their impacts on hydrological and biochemical cycles, often using satellite based climate data as input.

#### 1.4 Developing biodiversity indicators from remotely-sensed data

The development of biodiversity indicators involves a two stage process. Firstly it needs to be determined which biodiversity variables are needed to capture the status of the system. Secondly, a suitable remote sensing product has to be selected that can be linked to this variable. Many methods exist to derive information from remote sensing data, but depending on the system under monitoring and the required level of detail, a choice has to be made. In Annex 2 a summary of existing operational EO products and their applications in biodiversity monitoring can be found.

It is worth noting that satellite-derived information is not in a format which can be readily used as a biodiversity indicator but requires some modification in order to become an indicator (Strand *et al.*, 2007). GIS-based analysis of remotely-sensed information, supported by ground validation, is usually required before the data can become a usable indicator. This process of refining remote sensing information to the level of a biodiversity indicator is not straightforward and there are sometimes limits to the type and complexity of the indicators which can be developed. This applies to both terrestrial and marine environments which demonstrate unique challenges to indicator development (see sections 4.1 and 4.2 for further details).

#### 1.5 Why use remote sensing to monitor biodiversity?

#### 1.5.1 Traditional in situ methods

A variety of traditional *in situ* methods exist to survey (and then monitor) biodiversity. Their adequacy strongly depends on the target taxon. Common methods for sessile organisms (plants, fungi) are quadrant and transect sampling, where a square frame or rope, respectively, delineates the plot horizontally. Scientific methods to collect mobile species include canopy fogging (insects; e.g. Paarman & Stork 1987, Yanoviak *et al.* 2003), netting (birds: e.g. Dunn & Ralph 2004, Arizaga *et al.* 2011); bats: e.g. Larsen *et al.* 2007, Kalko *et al.* 2008; and fish: e.g. Lapointe *et al.* 2006, Achleitner *et al.* 2012, ), pitfalls (e.g. herpetofauna: Ribeiro-Júnior *et al.* 2008, Sung *et al.* 2011), pheromones or light (insects: e.g. Baker *et al.* 2011) and camera traps (e.g. O'Brian & Kinnaird 2013). Occasionally artifacts (e.g. pellets, dung, larval pupae) serve as evidence too (Hill *et al.* 2005), and for some species, other measurements may suffice for identification (e.g. acoustic monitoring of bats and birds Jones *et al.* 2013).

To obtain a representative sample of the examined habitat, a number of plots are typically required. To optimally allocate sampling effort in this respect, plots may be (systematically or randomly) stratified and/or clustered. In addition, often only a (random) subset of a quadrant is sampled, and observations along transects are recorded at predefined intervals only. Temporal variability of the target habitat may be as important to survey planning as spatial heterogeneity, because seasonality, daytime, weather and irregular disturbances (e.g. fires) co-determines the presence and / or detectability of an organism. In such situations plots may require multiple sampling visits to avoid/reduce temporal bias.

Species accumulation curves (which plot sampling effort unit versus. species found) are used to assess the sufficiency of sampling effort in a given plot. Inventory results are typically summarized into various diversity indices (e.g. Simpson or Shannon-Wiener), which are calculated from the observed number of different species (richness) and their relative abundance per sample unit (evenness).

Monitoring biodiversity with traditional *in situ* methods often requires as much effort as compiling the initial inventory (see above), because repeat measurements should be based on (nearly) the same sampling design and methods to accurately detect changes. Some optimization is possible though using occupancy modeling and power analysis (e.g. Sewell *et al.* 2012).

Especially in case of sparsely distributed organisms, as well as difficult to detect individuals (discussed e.g. in Mazerolle *et al.* 2007), traditional *in situ* sampling efforts may also become prohibitively expensive before a sample size is reached with sufficient statistical power to allow for estimates of (changes in) abundance.

Inaccessibility of some habitats within a study region (e.g. steep slopes, thick mangrove) but also practical considerations (e.g. proximity to roads or observer populations) may affect the comprehensiveness of results obtained with traditional in situ methods.

All sample site allocation schemes require a priori knowledge of the spatial (habitat) heterogeneity, which may be insufficient – especially at finer scales. Consequently some biodiversity values within the study region may remain undetected.

Insufficiently standardized sampling protocols may reduce the reproducibility of the initial inventory and thus inflate uncertainty of subsequent monitoring results (e.g. Braga-Neto *et al.* 2013).

Results cannot be extrapolated to the surrounding landscape or different temporal periods. At most, using expert knowledge and some generalized habitat maps, observed species-habitat relationships can be used to infer biodiversity in similar settings. The common practice however is to depict results of traditional in situ methods either as atlas grid cells or homogeneously for an entire examined area or strata.

### 1.5.2 Remote sensing

**Remote sensing** cannot replace traditional in situ methods for compiling initial inventories of species, except in case of very large species identifiable on airborne images, and very high resolution imagery collected by UAVs. However, remote sensing is a valuable large scale biodiversity monitoring tool at the level above species if coupled with quality ground data and likely to grow in value if embedded in a global, harmonized observation network (Pereira *et al.* 2013).

Remote sensing can be very useful for both planning surveys (and delineating strata in which initial surveys take place) as well as most importantly monitoring biodiversity changes thereafter. For example, remotely sensed imagery allows delineation of (spatial-temporal) habitat classes and strata within a study area, which is crucial for optimal sample site allocation. Remote sensing can also be used to identify habitat in space and time, which has not been examined yet with traditional in situ methods, and may harbor overlooked or yet unknown species. To meet the requirement of carrying out repeat measurements under spatiotemporal conditions similar to the initial inventory, remote sensing is extremely useful in identifying when and where to monitor.

If a robust relationship between ground truth observations and multivariate remote sensing data can be established, biodiversity conditions may be estimated for similar settings outside the study area – at species level by means of aggregated Species Distribution Models (SDMs) (e.g. Raes *et al.* 2009, Dubuis *et al.* 2011) or at ecosystem level (e.g. Duro *et al.* 2007, Roccini *et al.* 2010). Using SDM techniques, remote sensing represents an efficient and cost-effective monitoring tool. To identify and calibrate reliable biodiversity proxies and indicators, permanent monitoring plots and standardized survey protocols are essential (e.g. Jürgens *et al.* 2012, Chawla *et al.* 2012, and Braga-Neto *et al.* 2013).

Table 1.2. List of some key advantages and disadvantages of airborne and spaceborne remote sensing compared to traditional *in situ* methods

Advantages	Disadvantages
Provide a continuous, repetitive, large-scale synoptic view relative to traditional point-based field measurements	Remote sensing instruments are expensive to build and operate
Practical way to obtain data from dangerous or inaccessible areas	Remote sensing data are not direct samples of the phenomenon and it must be calibrated against reality. The measurement uncertainty can be large
Relatively cheap and rapid method of acquiring up-to-date information over a large geographical area	Remote sensing data must be corrected geometrically and georeferenced in order to be useful as maps, not only as pictures. This can be easy or complicated
Easy to manipulate with the computer, and combine with other geographic coverage in the GIS.	Remote sensing data interpretation can be difficult, which usually need to understand theoretically how the instruments are making the measurements, need to understand measurement uncertainties, and need to have some knowledge of the phenomena you are sampling.

## Annex 2. Overview of available remote sensing / Earth Observation products

### 2.1 Operational Earth Observation products used to monitor biodiversity

On the following pages existing operational EO products are summarized according to their applications in biodiversity monitoring and their potential to support the Convention. To this purpose they have been mapped against the key Aichi Targets they have the potential to help tracking progress towards and the CBD operational indicators. In addition, candidate EBVs they could contribute to have been identified. Databases mentioned can be found in Annex 4, Tables 4.1 and 4.2. In addition, a more detailed mapping including secondary Aichi Biodiversity Targets these products could support, key features, summary of key features and available datasets can be found in Annex 4, Table 4.3.

### 2.1.1. Operational land-based EO products

### Land cover and Land cover change

Land cover is the visible features of the Earth surface including vegetation cover as well as natural and manmade features which cover the surface of the Earth (Campbell, 2006). These are physical features of the Earth surface in contrast to land use which is an implied use of the feature, e.g. a field for agriculture. Physical features of the Earth's surface reflect solar radiation in different ways and therefore demonstrate unique spectral characteristics. The spectral characterization of different land cover types allows land cover to be mapped over broad areas from EO satellite sensors. Land cover can be mapped at a range of spatial scales. At the local-scale ground surveys are often employed while aerial and satellite images are more commonly employed from regional to national scales.

Land cover maps are frequently used as a means of visually assessing broad-scale patterns in land cover across regions, countries or continents and relating these with species distributions or species richness (Cardillo *et al.*, 1999) and identifying likely biodiversity hotspots through 'gap analysis' (Scott and Jennings, 1998). Such maps can also be useful to identify land cover change in and around protected areas and can contribute to improved management of existing protected areas (Jones *et al.*, 2009). Land cover can be used as a variable to parameterise land use, agro-meteorological, habitat and climate models and as inputs to more complex EO-based products such as the MODIS LAI and FAPAR (Myneni *et al.*, 2002).

Examples of operational land cover maps and some land cover data distributing centers are listed in the Annex 4. While these are open-access land cover maps, they have been created using different methodologies and classification systems which have been designed to satisfy different end user requirements and institutional needs. This makes integration of land cover maps very difficult. Furthermore, these tend to be static maps giving a snapshot of land cover in time although some have periodic updates, e.g. CORINE Land Cover (CLC) 1990, 2000 and 2006. The biodiversity community could benefit from an assessment of needs in relation to land cover mapping. This could help to focus efforts to produce a set of land-cover/use products that meet the needs of the biodiversity community.

Land cover and land cover change is most relevant to:

- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Trends in extent of selected biomes, ecosystems and habitats (decisions VII/30 and VIII/15)
  - $\checkmark$  Trends in the proportion of natural habitats converted

#### • GEO BON EBVs

- ✓ Ecosystem extent and fragmentation
- ✓ Habitat disturbance

#### Fire

The thermal radiation emitted by surface fires is detectable from EO sensors (Dozier, 1981). For example, the Along Track Scanning Radiometer (ATSR) sensor produces monthly fire maps based on land surface temperature data. The ATSR World Fire Atlas shows the spatial extent of burnt areas and the locations of active fire fronts (Arino *et al.*, 2005). However, spectral information in range of wavelengths, from the visible to infrared, can be potentially be used to detect active fires and separate them from non-burned areas, as has been done with MODIS (Roy *et al.*, 2007). Forest fire can rapidly alter ecosystem structure and change the nature of surface materials from living vegetation to charred organic matter and ash (Kokaly *et al.*, 2007).

Regularly-acquired fire data can contribute to understanding the temporal cycle of fire activity on a seasonal and annual basis and its impact on greenhouse gas emissions, in particular carbon dioxide (Zhang *et al.*, 2003). Operational fire products are produced at continental to global scales and updated in near real-time. The International Strategy for Disaster Reduction provides a comprehensive list of EO-based fire products. Fire products from 1999 to present are open access from the Global Land Service portal using SPOT/VGT data and MODIS products from the Land Processes Distributed Active Archive Centre (LP-DAACs). The MODIS Rapid Response System provides near real-time fire monitoring from a variety of EO sensors. The European Space Agency ATSR World Fire Atlas has monthly global fire maps from 1995 to present. While these data sources provide information on the spatial distribution of fires and their timing, understanding the cause of fires is important for conservation planning.

Fire products are most relevant to:

- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
   ✓ Trends in extent of selected biomes, ecosystems and habitats (decisions VII/30 and VIII/15)
- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
- GEO BON EBVs
  - ✓ Disturbance regime

#### **Biophysical vegetation parameters**

There are two operationally-produced biophysical vegetation parameters, Leaf Area index (LAI) and the Fraction of Absorbed Photosynthetically Active Radiation (FAPAR) which are important in several surface processes, including photosynthesis, respiration and transpiration (Baret *et al.*, 2013).

LAI is defined as the area of leaf surface per unit area of soil surface (Campbell, 2006) and is an important variable for surface-atmosphere interactions such as water interception, photosynthesis and evapotranspiration and respiration. FAPAR acts like a battery for the plant photosynthetic process measuring the plants ability to assimilate Photosynthetically Active Radiation (PAR) and generate green leaf biomass (Gobron *et al.*, 2006). Both of these parameters are related as LAI is the biomass equivalent of FAPAR and both play a role in driving ecosystem process models. For example, FAPAR is an essential variable in light use efficiency models (McCallum *et al.*, 2009).

LAI can be measured in-situ by measuring leaf area directly or through hemispherical photography while FAPAR can be inferred from measurements of incoming and outgoing solar radiation. However, both of these methods are labour intensive. Remotely-sensed LAI and FAPAR products are generated at regional and global scale and produced operationally form sensors such as Envisat EMRIS (non-operational since 2012) and Terra MODIS. However, gaps due to cloud cover necessitate compositing daily data into regular intervals typically from 8 to 16 days. Time series of LAI and FAPAR can be used to monitor seasonal vegetation dynamics such as crop cycles and land surface phenology. For example, a slight global greening trend has been detected using a multi-decadal time series of LAI (Siliang *et al.*, 2010).

The biophysical vegetation parameters are most relevant to:

- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
  - ✓ Target 10. By 2015, the multiple anthropogenic pressures on coral reefs, and other vulnerable ecosystems impacted by climate change or ocean acidification are minimized, so as to maintain their integrity and functioning.
  - ✓ Target 14. By 2020, ecosystems that provide essential services, including services related to water, and contribute to health, livelihoods and well-being, are restored and safeguarded, taking into account the needs of women, indigenous and local communities, and the poor and vulnerable.
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Status and Trends in extent and condition of habitats that provide carbon storage
  - ✓ Trends in primary productivity
- GEO BON EBVs
  - ✓ Net Primary Productivity (NPP)
  - ✓ Phenology

#### Vegetation Productivity Spectral Indices

A spectral index such as the Normalised Difference Vegetation Index (NDVI) is generic to any sensor recording electromagnetic radiation in the red and near infrared spectral bands. However, the shortcomings of NDVI, in relation to the influence of atmosphere and sensor-specific variation, have already been documented (Pinty and Verstraete, 1992). Other spectral indices such as the MODIS Enhanced Vegetation Index (EVI) have been designed for specific sensors however. While the NDVI solely employs spectral information, indices such as the EVI are built on spectral information parameterised for sensitivity to green biomass and are therefore less likely to saturate in areas of dense biomass such as rainforest (Huete et al., 2002). The NDVI is a general indicator of vegetation presence or absence but is less stable than the EVI, particularly in time series analysis. However, both indices can show variation in vegetation productivity and condition when mapped spatially. These spectral indices can be used at any scale from local to global, particularly the NDVI as any sensor measuring radiation in the red and near infrared spectral bands is all that is required. However, there is a need for awareness of the strengths and weakness of these indices and caution in applying them to strictly quantitative rather than qualitative analyses (Campbell, 2006). The biophysical variables are best used in quantitative analysis of vegetation variables. These indices are best used as general indicators of the vegetation state and are useful to detect relative change in vegetation condition, in particular to detect where habitat disturbances are occurring and causes a reduction in the spatial extent of vegetated areas.

The Vegetation Condition Index (VCI) and the Vegetation Productivity Index (VPI) are operational global products based on NDVI. These products compare contemporary NDVI data with historic trends to identify vegetation growth anomalies, e.g. drought, and so are useful to monitor temporal change in vegetation condition. The VCI and VPI can be obtained from the Copernicus Global Land Service.

The biophysical vegetation parameters are most relevant to:

• CBD Aichi Biodiversity Target

- ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Trends in condition and vulnerability of ecosystems
  - ✓ Trends in primary productivity

#### • GEO BON EBVs

- ✓ Ecosystem extent and fragmentation
- ✓ Habitat disturbance.

#### Vegetation Cover and Density

Vegetation Continuous Fields (VCF) and Fraction of vegetation Cover (fCover) are designed to measure the relative spatial coverage of vegetation in an image pixel. While the VCF estimate the relative proportions of vegetative cover types per pixel: woody vegetation, herbaceous vegetation, and bare ground (de Fries *et al.*, 1999, Hansen *et al.*, 2003), the fCover is a relative measure of the gap fraction in green vegetation (Baret *et al.*, 2007). However, fCover has also been used as an input to climate models in separating the contribution of soil from vegetation (Baret *et al.*, 2013).

They are also important components of land cover. For example, the continuous classification scheme of the VCF product may be more effective in characterising areas of heterogeneous land cover better than discrete classification. Regularly updating static land cover maps with measures of fCover can incorporate disturbance as a land cover variable producing more adaptable land cover products. Annual and global VCF data from Terra-MODIS (NASA) imagery are distributed by the Global Land Cover Facility (GLCF). The fCover product is accessible from the Copernicus Global Land Service.

Vegetation Continuous Field and fraction of green cover are most relevant to:

- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Trends in proportion of degraded/threatened habitats
  - ✓ Trends in fragmentation of natural habitats

• GEO BON EBVs

- ✓ Ecosystem extent and fragmentation
- ✓ Habitat disturbance.

#### **Biomass**

Biomass is quantified in terms of the overall mass of plant material (Campbell, 2006). EO-based measures of biomass are calibrated and validated using local-scale in-situ measures of above-ground biomass (Saatchi *et al.*, 2007), while below-ground biomass is a more challenging parameter for EO-based technology (Cairns *et al.*, 1997). However, the total combined above-ground and below-ground biomass has been estimated from a synthesis of EO and airborne sensor data, as well as ground

measurements, across Latin America, sub-Saharan Africa, and Southeast Asia (Saatchi *et al.*, 2011). As there is currently no EO sensor directly monitoring biomass, remotely-sensed methods of biomass estimation are indirect and inferred from estimates of vegetation canopy volume. Therefore canopy height estimation from airborne or satellite Lidar is an important first step in biomass calculations which are then extrapolated over large areas using a model based on coarser resolution satellite imagery such as MODIS (Saatchi *et al.*, 2011).

As most of the global biomass is held in woody trees (Groombridge and Jenkins, 2002), biomass is frequently used as preliminary variable to assess forest carbon stocks. Satellite-derived estimates of above-ground woody biomass provide reliable indications of terrestrial carbon pools (Dong *et al.*, 2003). Therefore, remote sensing of deforestation, land use change and global forest fires can contribute to improved models of the global carbon cycle. Changes in biomass are also likely to result in changes in biodiversity.

As biomass estimation methods are labour intensive and indirect, EO-based biomass products are not yet operational. However, Dry Matter Productivity (DMP) is produced operationally and can be accessed from the Global Land Service, GEONET Cast and DevCoCoast. DMP represents the daily growth of standing biomass (equivalent to the Net Primary Productivity) and is expressed in kilograms of dry matter per hectare per day. The European Space Agency mission, BIOMASS, due in 2020 and based on radar technology, will provide global measurements of forest biomass (Le Toan *et al.*, 2011).

#### Biomass is most relevant to

- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
  - ✓ Target 15. By 2020, ecosystem resilience and the contribution of biodiversity to carbon stocks has been enhanced, through conservation and restoration, including restoration of at least 15 per cent of degraded ecosystems, thereby contributing to climate change mitigation and adaptation and to combating desertification.
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Trends in primary productivity
    - $\checkmark$  Status and trends in extent and condition of habitats that provide carbon storage
- GEO BON EBVs
  - ✓ Habitat Structure
  - ✓ Net Primary Productivity (NPP)

#### 2.1.2. Operational marine EO products

Ocean-based EO products differ in their method of retrieval and their spatial and temporal coverage from land-based products (Campbell, 2006). This difference is predominately due to the physical reflectance characteristics of land surfaces and water bodies. Water reflectance is determined by the state of the water surface, the amount and type of suspended material in the water column and the bottom substrate in areas of shallow water (Lillesand *et al.*, 2008). Furthermore, dynamic ocean variables such as eddies and currents change at a more rapid rate than polar-orbiting sensors can sufficiently monitor (Campbell, 2006).

Nevertheless, satellite sensors (e.g. SeaWiFs, Envisat MERIS and NOAA AVHRR) have been optimised to retrieve ocean variables such as ocean colour (chlorophyll-a concentration in mg/m3) (Brewin et al., 2011), ocean Primary Productivity (Antoine et al., 1996), suspended sediment, sea surface wind speed (m/s), sea surface temperature (°C), sea surface salinity and sea surface state (Campbell, 2006). While these are important state variables of the oceans and routinely monitored to track climate change, they are also habitat parameters in themselves. For instance, oceanic variables can be correlated with sea bird density and species compositions (Hyrenbach et al., 2007), cetacean species ranges (Tynan et al., 2005), as well as the distribution of pelagic species and near shore fishes (Johnson et al., 2011). Measures of ocean colour can be related to the abundance and type of phytoplankton which has important implications for the marine food chain (Brewin et al., 2011). For climate change monitoring in the marine envrionment, satellite remote sensing has been used to track Arctic sea ice extent, sea level rise, tropical cyclone activity and sea surface temperature (IPCC, 2007). This application of satellite remote sensing is discussed further in relation to Aichi target 15 in section 3 of the review.Global ocean colour, sea surface temperature and salinity are operationally produced and available for download from the NASA Ocean Colour website or from the GMES My Ocean website. ESA have an operational data portal for Ocean colour products called Globcolour. The NOAA Ocean Surface and Current Analysis (OSCAR) provide near-real time global ocean surface currents maps derived from satellite altimeter and scatterometer data.

The marine EO products are ocean colour (chlorophyll-a concentration in mg/m<sup>3</sup>), ocean Net Primary Productivity (NPP), suspended sediment, sea surface wind speed (m/s), sea Surface temperature (°C), sea surface salinity and sea surface state. They are most relevant to:

- CBD Aichi Biodiversity Target
  - ✓ Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
  - ✓ Target 8. By 2020, pollution, including from excess nutrients, has been brought to levels that are not detrimental to ecosystem function and biodiversity.
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Trends in condition and vulnerability of ecosystems
  - ✓ Trends in sediment transfer rates storage
- GEO BON EBVs
  - ✓ Ecosystem extent and fragmentation
  - ✓ Habitat disturbance
  - ✓ Net Primary Productivity (NPP)

#### 2.1.3 EO products for pollution monitoring

Remote sensing has considerable potential in monitoring the spatial extent of polluting material both in the upper atmosphere, on the land surface and in the marine environment. Though this is a relatively new application of earth observation satellite technology, it is a promising field of development and potentially impacts on a number of EBV categories and in helping to chart the progress towards achieving the 2020 Aichi targets. The EO products related to pollution are not strictly operational in that these products are mostly in development or form part of larger data dissemination and early warning systems. Nevertheless, examples of EO-based information systems which are currently in use for monitoring and forecasting pollution events are listed below.

#### Atmospheric pollution and greenhouse gas emissions

Some atmospheric pollutants contribute to the greenhouse effect while others are directly harmful to life and can contribute to habitat degradation and biodiversity loss. The main greenhouse gases are carbon dioxide, methane and nitrous oxide (N2O). Further information on these gases and their implication for climate change can be found online (Greenhouse Gas Online, 2013).

The European Infrared Atmospheric Sounding Interferometer (IASI) measures the total column content of the main greenhouse gases, i.e., ozone, methane, nitrous oxide and carbon monoxide. These measurements contribute to an understanding of climate processes though their assimilation into global climate models. Products can be obtained from the IASI or associated sensors such as the EUMetsat Polar System (EPS). These products relate to temperature, humidity, ozone content and trace gas constituents of the atmosphere.

The NASA Microwave Limb Sounder (MLS) instrument measures passive microwave radiation from the upper atmosphere and derives estimates of atmospheric gases, temperature, pressure, and cloud ice. The MLS instrument is unique in its measurements of pollution in the upper troposphere as it can see through ice clouds that previously prevented such high altitude measurements. Such data can provide insights into the long-range transport of pollution and its possible effects on global climate. Near real time MLS products such as temperature, water vapor, ozone, carbon monoxide, water vapor, nitrous oxide, nitric acid and sulphur dioxide can be viewed online.

Nitrogen dioxide (NO2) is a mainly man-made gas which forms nitric acid when oxidised creating acid rain. Acid rain has adverse impacts on soil, vegetation and can contribute to ocean acidification. Nitrogen oxides such as NO2 are produced by emissions from power plants, heavy industry and road transport, along with biomass burning. NO2 is important in atmospheric chemistry as it is responsible for the overproduction of tropospheric ozone, i.e. in the lower part of the atmosphere. A global NO2 pollution map was produced by the ESA Envisat Sciamachy satellite in 2004 although this sensor was decommissioned in 2012. However, a variety of Sciamachy-based atmospheric products from 2002 to 2012 are available though registration with ESA on their data user portal. Upper atmosphere, stratospheric N2O is inferred from measurements by sensors on board the US AURA and European MetOp satellite series.

The atmospheric EO products that relate to  $NO_2$  and ozone are most relevant to:

- CBD Aichi Biodiversity Target
  - ✓ Target 8. By 2020, pollution, including from excess nutrients, has been brought to levels that are not detrimental to ecosystem function and biodiversity.
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - ✓ Trends in nitrogen footprint of consumption activities
  - ✓ Trends in ozone levels in natural ecosystems
- GEO BON EBVs
  - ✓ Habitat disturbance

#### **Ocean pollution**

Oil spills such as the Prestige disaster of 2002, the Exxon Valdez in 1989 or the Deepwater Horizon oil rig of 2010 are a reminder of the threat posed to the marine environment of oil spills. Fortunately, large-scale surveillance of oil spills in the marine environment can now be readily achieved by satellite and airborne remote sensing (Leifer *et al.*, 2012). Accidental, high-impact oil spills, and non-accidental incidental spills from marine vessels can be tracked in spatial extent and flow direction (Engelhardt, 1999). Remote sensing is also used to localise point sources of oil slicks and for tactical assistance in emergency remediation.

Synthetic Aperture Radar (SAR) is the most frequently used satellite-based tool since it operates at night time. It penetrates cloud cover and is sensitive to surface roughness (Bern *et al.*, 1993; Campbell, 2006). The smooth oil slick contrasts with the surrounding surface water and appears as a dark patch on the SAR image.

CleanSeaNet is an example of an operation oil spill monitoring service based on EO technology which consists of oil slick imaging systems which also provide real-time sea state and weather information. This information is essential to track the rate and direction of slick movement. CleanSeaNet, which is operationally employed by marine authorities in EU member states, is part of the Global Monitoring for Environment and Security (GMES) initiative. Pollution alerts and related information is relayed to the relevant authorities 30 minutes after image acquisition for timely response. Currently, there are no operational open access products on ocean pollution events as they are relayed to relevant users as they occur and therefore need rapid delivery through formalised systems.

The impact of spills on biodiversity can be accessed through the integration of remote sensing imagery with other geographical layers such as marine and coastal protected areas and marine species ranges (Engelhardt, 1999). For example, the NOAA Office of Rapid Response and Restoration has produced an open-access Environmental Sensitivity Index (ESI) system, based on multiple data layers on biological and human land use of shorelines, for the U.S. This index is used to rank shorelines according to their sensitivity to an oil spill. The system is useful to planners for contingency planning before an oil spill occurs and for rapid response once it has occurred in order to direct resources to where they are most needed.

The oceanic EO products that relate to oil spill detection and shoreline sensitivity are most relevant to

- CBD Aichi Biodiversity Target
  - ✓ Target 8. By 2020, pollution, including from excess nutrients, has been brought to levels that are not detrimental to ecosystem function and biodiversity.
- CBD Strategic Plan for Biodiversity 2011-2020 operational indicators
  - $\checkmark$  Trends in emission to the environment of pollutants relevant for biodiversity
- GEO BON EBVs
  - ✓ Habitat disturbance

# Annex 3. Emerging applications of remote sensing in the context of the Convention

This section summarises emerging applications of remote sensing for both marine and terrestrial environments relevant for tracking progress towards the Aichi Biodiversity Targets, setting the basis for discussing on future directions.

#### 3.1 Near real-time remote sensing for surveillance

Operational near real-time imagery has a great potential as tool for surveillance and monitoring implementation of law and policies, which has been underused to date. Satellite imagery and derived products can have a short 'shelf-life 'when it comes to such applications as crop monitoring, deforestation monitoring or disaster response. The images are made available after an event or a potential hazard has occurred limiting their utility in disaster response and hazard mitigation. Operational near real-time availability of imagery is needed in such cases.

An example of this applicability is the monitoring of illegal deforestation in the Brazilian Amazonia. The Disaster Monitoring Constellation International Imaging Ltd (DMCii) is now providing imagery to the DETER service of the INPE in Brazil which uses regularly acquired MODIS satellite images to detect forest clearance (Hansen and Loveland, 2012). The DMCii imagery will provide INPE with medium resolution monitoring capabilities to overcome the ability of illegal loggers to go undetected at the 250m spatial resolution of the MODIS pixel. Further details can be found in section 3 of the review.

Fire surveillance also adopts near real-time monitoring systems based on EO data. For example, the Geoscience Australia Sentinel system uses daily MODIS imagery to monitor fires as they occur across the Australian continent (see section 3.1 for further details). This approach has also been adopted in different African countries.

Main CBD Aichi Biodiversity Target it supports:

- ✓ Aichi Target 5. By 2020, the rate of loss of all natural habitats, including forests, is at least halved and where feasible brought close to zero, and degradation and fragmentation is significantly reduced
- ✓ Aichi Target 7. By 2020 areas under agriculture, aquaculture and forestry are managed sustainably, ensuring conservation of biodiversity

#### 3.2 Pollution and its impact on biodiversity

The role of remote sensing in monitoring atmospheric gases in the context of climate change was discussed in Annex 2. However, there are considerable negative impacts of increased atmospheric nitrogen on biodiversity, in particular floristic diversity and plant health (Phoenix, *et al.*, 2006). Although there are currently no direct ways to monitor the biodiversity impact of atmospheric nitrogen deposition using remote sensing, its impacts on plant vigour can be monitored using the vegetation products discussed in Annex 2.

Eutrophication of water bodies occurs with overload of plant nutrients, closely linked to land use/ land cover changes, and frequently result in 'algal blooms'. The reflectance of water changes with chlorophyll concentration as water with high chlorophyll concentration is usually typified by high green reflectance and absorption in the blue and red spectral regions (Lillesand *et al.*, 2008). Quantitative methods of algal bloom monitoring from aerial and spaceborne sensors use these reflectance properties to map and monitor their occurrence. Due to the spectral similarities between blue-green and green algae, narrow band sensors such as hyperspectral imagery or filtered airborne cameras are frequently used. More advanced methods relying on hydrodynamic–biogeochemical models which assimilate bio-optical measurements from ocean-observing satellites are being used for more accurate EO-based products for eutrophication assessment (Banks *et al.*, 2012).

Ocean acidification has wide-ranging implications in marine ecosystems and has stimulated studies in areas ranging from biochemistry of calcareous shell-forming processes to the socio-economic impacts on marine fisheries, aquaculture, and other ecosystem services (Doney *et al.*, 2009). Acidification happens when changes in seawater chemistry result from the oceanic uptake of anthropogenic  $CO_2$ . The change in pH levels has detrimental impacts for calcareous shell-building organisms such as foraminifera and pteropod molluscs (Fabry *et al.*, 2008). Coral reefs are also at risk as the rate of coral reef calcification is projected to decrease by 40% by 2065 based on increased abundance of oceanic  $CO_2$  (Langdon *et al.*, 2000). Satellite remote sensing can play a role in monitoring this phenomenon, e.g. by measuring reflectance from calcium carbonate, also known as Particulate Inorganic Carbon (PIC), as measured by MODIS (Balch *et al.*, 2005).

The NOAA Experimental Ocean Acidification Product Suite (OAPS) synthesises satellite and modelled environmental data sets to provide a synoptic estimate of sea surface carbonate chemistry which is updated monthly (OAPS, 2013). Satellite - based estimates of sea surface temperature based on the NOAA-AVHRR satellite are one of many parameters which contribute to the OAPS (Gledhill *et al.*, 2009). Modelling of surface-ocean carbonate chemistry, using remote sensing as a tool, allows regional to basin wide trends in ocean acidification to be explored on seasonal to interannual time scales. This is very important for monitoring ocean-wide marine biodiversity impacts since ship-based measurement are limited in spatial scope and frequency of measurement.

Main CBD Aichi Biodiversity Target it supports:

- ✓ Aichi Target 8. By 2020, pollution, including from excess nutrients, has been brought to levels that are not detrimental to ecosystem function and biodiversity.
- ✓ Aichi Target 10. By 2015, the multiple anthropogenic pressures on coral reefs, and other vulnerable ecosystems impacted by climate change or ocean acidification are minimized, so as to maintain their integrity and functioning.

#### 3.3 Monitoring the spread of invasive plant species

Spatial mapping of the spread of invasive alien plant species is a high priority for the conservation community and an area where a remote sensing-based approach could make a substantial contribution. There have been considerable advances in using remote sensing to map species that

dominate forest canopies using remote sensing imagery. However, a large proportion of invasive plants in native forests occur in the understory where they are often obscured by the canopy. In addition, plant communities are often present in the form of mixed-species mosaics which can be difficult to separate using spectral data alone (Zhang *et al.*, 2006). Indirect methods of mapping including the use of GIS data layers and modeling have been used in these cases. Besides passive sensor data, LiDAR has proved useful.

The key challenge the conservation community faces when monitoring invasive alien plant species is that species-level plant discrimination is not possible using current operational EO-based land cover or habitat products. Nevertheless, hyperspectral imagery has potential to provide species-level discrimination at the ecosystem level (Hestir *et al.*, 2008). However hyper-spectral-based products are not operational and hyperspectral remote sensing is frequently limited to local-scale studies employing airborne hyper spectral sensors, e.g. the Airborne Visible/Infrared Imaging Spectrometer (AVIRIS) operated by NASA/JPL. Spaceborne hyper spectral sensors are the Hyperion sensor onboard EO-1 spacecraft and the Compact High Resolution Imaging Spectrometer (CHRIS) of ESA's Proba-1 instrument.

Further exploration and operational development of hyperspectral-based products from these sensors is a necessity for future site-level plant species mapping which will highly benefit monitoring the spread of invasive alien plant species. Airborne imagery and sub-metre resolution satellite imagery can also make a significant contribution to invasive species mapping.

Main CBD Aichi Biodiversity Target it supports:

✓ Aichi Target 9. By 2020, invasive alien species and pathways are identified and prioritized, priority species are controlled or eradicated, and measures are in place to manage pathways to prevent their introduction and establishment.

### 3.4 Assessment of management effectiveness and establishment of ecologically effective Protected Areas networks

Land use change around protected areas has been recognised as an important determinant of forest reserve health in tropical regions (Laurance *et al.*, 2012). As observed from MODIS VCF data, up to 68% of protected areas in a wide-ranging, global sample of highly protected tropical forests had their cover reduced within a 50-km periphery of their administrative boundaries. Far fewer of those protected areas experienced loss of forest habitat within their administrative boundaries (De Fries *et al.*, 2005). Such studies demonstrate the importance of considering land use dynamics at or beyond the boundaries of protected areas for more effective protected area management strategies.

Currently, large area monitoring of land cover change at medium spatial resolution predominately uses Landsat data due to the availability of a multi-decadal time series (Hansen and Loveland, 2012). Assessing protected area effectiveness requires change analysis methods which are consistent and repeatable over time, preferably at high to very high spatial resolution. Change mapping methods are therefore set to change from analyst interactions with individual scenes to automated processing chains which harness powerful computing to process large data volumes (Hansen and Loveland, 2012). Ideally, this would be combined with near-real time alert systems which are triggered by sudden change, as proposed by Verbesselt *et al.* (2012). This approach would increase sensitivity of alert systems to natural and anthropogenic disturbance events such as illegal logging and drought. Protected area level monitoring using EO-based tools is now possible with the Digital Observatory for Protected Areas (DOPA) jointly developed by GEO BON and JRC. The DOPA has delivered a suite of informatics-based, we-enabled tools to conservation managers to monitor the state and pressures on protected areas globalle (Dubois *et al.*, 2011).

In Canada, candidate areas for protection status and existing protected area networks are being monitored through remotely-sensed indicators on land cover, fragmentation, disturbance and snow cover. Areas sharing common environmental conditions using this approach can be used to assess the effectiveness of Canada's network of parks and identify sites requiring protection. More details of this approach can be found in section 3.3 of the review.

#### Main CBD Aichi Biodiversity Target it supports:

✓ Aichi Target 11. By 2020, at least 17 per cent of terrestrial and inland water, and 10 per cent of coastal and marine areas, especially areas of particular importance for biodiversity and ecosystem services, are conserved through effectively and equitably managed, ecologically representative and well connected systems of protected areas and other effective area-based conservation measures, and integrated into the wider landscapes and seascapes.

### 3.5 The use of terrestrial and marine mammals as sensor platforms

Technological advances in the last few decades have made it possible to use animals as platforms to carry remote-sensing devices in a growing discipline known as animal telemetry. However, there has been more limited use of terrestrial animals as sensor platforms in comparison to marine ecosystems. Commonly used methods for tracking animals in the terrestrial environment using individual tags are Global positioning system (GPS), Argos Doppler tags, very high frequency radio tags, light-level geolocator and banding or rings. However, not all of these rely on satellite sensor technology as acoustic devices are based on radio signals (Movebank, 2013).

The U.S. Integrated Ocean Observing System (IOOS) is making efforts to use data from electronic tags attached to marine animals to enhance understanding of the marine environment (IOOS, 2013). For example, movement of the hawksbill turtle in the Caribbean Sea has been characterized using telemetry, showing that they are more abundant in protected areas than previously thought (Scales *et al.*, 2011). Animal-based tags are so useful because sensors can track individuals over long distances for multiple years, collecting sub-surface data from remote and difficult to reach environments. Conventional earth observation techniques are technically or economically unfeasible for monitoring movement and environmental conditions at the individual level.

Main CBD Aichi Biodiversity Target it supports:

 Aichi Target 12. By 2020 the extinction of known threatened species has been prevented and their conservation status, particularly of those most in decline, has been improved and sustained.

#### 3.6 Ecosystem services: carbon storage and climate change

Remote sensing-based assessment of carbon stocks in terrestrial habitats is a major field of research and relies heavily on remote sensing for quantitative spatial data on vegetation biomass, among other variables such as Gross Primary Production (GPP). Remotely-sensed surrogates of tree species diversity, such as the NDVI-based eco-climatic distance measure, have been related to carbon storage and sequestration in forests as well. This measurement demonstrates a strong relationship with treedensity, LAI and degree of deciduousness. Therefore continuous measurements over broad spatial scale can detect broad scale patterns of bio-diversity in forested landscapes and ecosystem services that can be used in conservation planning (Krishnaswamy *et al.*, 2009).

The relation between biomass and carbon storage has already been discussed in Annex 2. In order to quantify above ground carbon content in forests, LiDAR is a frequently used tool, but is mostly used at a local scale owing to the small footprint of LiDAR instruments. In heteregenous forests, LiDAR-has been proven to be a more effective tool than ground-based methods in quantifying above ground carbon content (Patenaude *et al.*, 2004). The forest carbon stock of areas the size of the Peruvian Amazon can be quantified at high resolution (0.1-ha) based on the integration of LiDAR, Landsat imagery and field plots (Asner *et al.*, 2010). Landsat-derived NDVI is well correlated to carbon storage in urban forestry, based on field measurements, providing the potential for cost-effective and efficient regional forest carbon mapping (Myeong *et al.*, 2006).

However, there are few studies of carbon stocks in ecosystems other than forest. Efforts to model the land-atmosphere exchange of  $CO_2$  from high latitude, northern hemisphere peat lands using satellite remote sensing inputs are already well established (Schubert *et al.*, 2010). Similar methods are employed to monitor grassland gross primary production and  $CO_2$  uptake, but using in-situ spectral measurements of vegetation phenology combined with an estimation of radiation use efficiency (Migliavacca *et al.*, 2011). The conservation community would find it especially useful to assess carbon stocks for grasslands and peat lands (Green *et al.*, 2011). This would represent a worthwhile avenue for research in future carbon assessments based on EO data.

The role of remote sensing in monitoring the impact of climate change on ecosystems can be shared between observation data on primary and secondary indicators. Primary indicators include temperature, precipitation and FAPAR. A secondary indicator, vegetation phenology, is an essential component of ecosystem functioning (Thackeray *et al.*, 2010), an important climate change indicator (van Vliet, Overeem *et al.* 2002, Butterfield and Malström, 2009), and has been widely observed for several decades.

Remote sensing of land surface phenology is now a well established field of research providing an objective and repeatable method of phenological observation that can contribute to climate change studies. However, remotely sensed phenological patterns are observed from multiple vegetation ecosystems and not a single plant or tree species and are limited in time series as compared to ground-based observations. Finer-scale ecosystem level observation are now possible using fixed-position, digital-camera based sensors, e.g. the Phenocam in selected forests in the U.S.A. (Sonnentag *et al.*, 2012) or the Phenological Eyes Network in Japan (Nagai *et al.*, 2013). Canopy-level monitoring of phenology has important implications for estimation of gross primary production of forested or

grassland ecosystems. Therefore, phenological information gathered by in-situ sensors such as digital cameras, can be used in estimating local carbon sinks and sources.

Main CBD Aichi Biodiversity Target it supports:

✓ Aichi Target 15. By 2020, ecosystem resilience and the contribution of biodiversity to carbon stocks has been enhanced, through conservation and restoration, including restoration of at least 15 per cent of degraded ecosystems, thereby contributing to climate change mitigation and adaptation and to combating desertification.

The use of UAVs for remote sensing has become more widespread due to recent technical advances in miniaturisation, communication, the strength of lightweight materials and power supplies (Campbell, 2006). They offer near-surface observations in order to record complementary environmental information such as temperature, CO<sub>2</sub> and humidity. Their rapid deployment allows greater flexibility for use in dangerous and inaccessible environments permitting rapid change analysis while flights can be planned according to local weather conditions (Watts *et al.*, 2010). As they operate below the cloud line, cloud-free observations are guaranteed and atmospheric correction of imagery is not required. UAVs can be considered as flexible sensor platforms as different sensors can be mounted giving them adaptability in different applications including aerial photography, optical, thermal and hyperspectral analysis. They are limited in spatial scope however and are frequently employed in site-level monitoring for which satellite or airborne sensors are too coarse in resolution or too infrequent in revisit time. Therefore UAVs are effective tools for modeling and monitoring biodiversity-related variables at a local scale.

UAV flights can be flown at the same time as satellite or other airborne sensors for coincident measurements (Campbell, 2006). Applications include invasive species mapping (Watts *et al.*, 2010) and precision agriculture, to detect water stress and irrigation effectiveness in orchards (Stagakis *et al.*, 2012, Zarco-Tejada *et al.*, 2012) and to measure temperature at the plant canopy level using thermal remote sensing (Berni *et al.*, 2009). UAVs are also used in the coastal zone (Malthus and Mumby, 2003) and in riparian habitats (Dunford *et al.*, 2009). However, combining multiple images from different flight lines and dates can be problematic due to variability in solar illumination and sensor movement (Dunford *et al.*, 2009).

## Annex 4. Detailed mapping of databases, remote sensing sensors, targets and indicators

Variable	Existing database	Institution	Satellite Sensors	Access
Land-based	Global Land Service	Copernicus	SPOT- VGT	Open
	Distributed Active Archive Centers (DAACs)	NASA	MODIS	Open
Land, atmosphere and water based	Giovanni <sup>2</sup>	NASA Goddard Earth Sciences	Multiple	Open
Marine	Ocean Colour website	NASA	Multiple	Open
Land, atmosphere and ocean	Office of <sup>3</sup> satellite and product operations	NOAA	Multiple	Open
Atmospheric, ocean and land	GEONETCast website	Group on Earth Observation (GEO)	Space-based, air-borne and in situ	Open
Land-based (developing countries)	DevCoCast website	Global Earth Observation System of Systems (GEOSS)	Multiple	Open
Land-based (Indian sub- continent)	Biodiversity Information System (Roy and Saran, 2004)	Indian Institute of Remote Sensing	IRS-LISS II/ SPOT/Landsat	Open

Table 4.1. Mentioned existing global databases for the main EO products used to monitor biodiversity

Table 4.2 Existing landcover databases at different spatial scales

<sup>2</sup> The Giovanni data parameter database contains over 4,000 data parameters which are catalogued by their corresponding data product or sensor but are more restricted in terms of their spatial coverage, access rights and require more processing and user input. It has in-built analytical tools and is more of a scientific analysis tool than a download portal

<sup>3</sup> Spatial coverage is sometimes restricted to the United States

Variable	Existing database	Year	Institution	Scale	Sensor
Landcover (and	National Land Cover Database (NLCD)	1992, 2001, 2006	USGS Earth Resources	U.S.A.	Landsat
associated variables)			Observation and		
			Science (EROS) Centre		
Landcover	Global Land Cover (GLC) 2000	2000	Joint Research Centre	global	SPOT-VGT
			(European		
			Commission)		
Landcover	GlobCover Portal	2006, 2009	European Space	global	MERIS
			Agency (ESA)		
Landcover (and	Africover database	Various	The Food and	National (African	Various
associated variables)			Agricultural	countries)	
			Organisation (FAO)		
Landcover	CORINE Land Cover (CLC)	1990, 2000, 2006	European Environment	Pan-European	
			Agency(EEA)		

Table 4.3. Mapping of EBVs, Aichi targets, CBD Operational indicators and relevant EO products

Operational indicator	Candidate EBV	Most relevant Aichi target	Other Aichi Target supported	EO Product	Acronym	In-situ	Key features	Variable Measured	Spatial scale	Application to conservation	Access	Existing databases	Tempor al coverag e	Level of product development
Trends in climate impacts on population trends	Phenology (vegetation)	15	8,14, 10	Leaf Area Index	LAI	Measuring leaf area directly or through	Important in surface- atmosphere	Area of leaf surface per unit area of soil	Global, 10°x10° tiles, Continental tiles	Input to Net Primary Productivity Models or as a correlate of	Open access	Global Land Service	1999- present	Operational
Status and Trends in extent and condition of habitats that						hemispherical photography	interactions such as	surface	Global, 10°x10° tiles	other environmental variables understand		Global Land Service	2009- present	
provide carbon storage							photosynthesis, evapotranspiratio n and respiration		Africa and South America continental tiles	vegetation-climate interactions		GEONET Cast DevCoCast website	Near- real time only Aug 2007- present	
Trends in primary productivity Status and Trends in extent and condition of habitats that provide carbon storage		5 15		Fraction of Absorbed Photosynthet ically Active Radiation	FAPAR	Eddy covariance measurements	Acts like a battery for the plant photosynthetic process	FAPAR absorbed by the plant canopy instantaneous with satellite overpass	Global, 10°x10° tiles, Continental tiles	Input to Net Primary Productivity Models or as a correlate of other environmental variables	Open access	Global Land Service	1999- present	Operational
Trends in condition and vulnerability of ecosystems Trends in proportion of degraded/threatened habitats				Normalised Difference Vegetation Index	NDVI	Flux towers and digital cams	Spectral band ratio to detect differential reflectance in red and near infrared bands from green vegetation	Not a biophysical variable but an estimate of the vegetation amount	Global, 10°x10° tiles Africa and South America continental tiles	Monitor vegetation state, health and disturbance	Open access	Global Land Service GEONET Cast DevCoCast website	1999- present Near- real time only Aug 2007- present	Operational
Trends in primary productivity Status and Trends in extent and condition of habitats that provide carbon storage	Net primary productivity	5		Dry Matter Productivity	DMP	Not measurable	Directly related to NPP but customised for agronomic applications	Dry matter biomass increase (growth rate) expressed in kilograms of dry matter per hectare per day	Global, 10°x10° tiles Africa and South America continental tiles	Identify anomalies in vegetation productivity and to forecast crop yields	Open access	Global Land Service GEONET Cast DevCoCast website	2009- present Near- real time only Aug 2007- present	Operational
Trends in condition and vulnerability of ecosystems	Net primary productivity	5		Ocean colour	n/a	Not measurable	Phytoplankton contain chlorophyll and	Chlorophyll-a	Regional seas, major oceans, major inland water bodies	Related to phytoplankton, primary production and marine food chain	Open access	GMES My Ocean NASA Ocean Colour	Variable	Operational
Trends in condition and vulnerability of ecosystems	Net primary productivity	5		Sea Surface Temperature	SST	Marine weather buoy network	Depends on method , e.g. optical measures 'skin' temperature, radar penetrates sub-surface	Temperature of water surface		Determines the distributions of marine plant and animal species	Open access	PO DAAC (NASA) GMES My Ocean ESA CCI SST	Variable	Operational

Operational indicator	Candidate EBV	Most relevant Aichi target	Other Aichi Target supported	EO Product	Acronym	In-situ	Key features	Variable Measured	Spatial scale	Application to conservation	Access	Existing databases	Temporal coverage	Level of product development
Trends in distribution of selected species	Migratory behavior	12	5,6,10,11	Banding/ marking/ tagging and observation of individuals	Internatio nal Cooperatio n for Animal Research Using Space (ICARUS)	Measurable	Satellite or radio tagging	Global position but also physiological characteristics	All scales	Species range and habitat, foraging behavior, migration patterns	Open access	Movebank	Variable	Operational
Trends in extent of selected biomes, ecosystems and habitats (decision VII/30 and VIII.15)	Disturbance regime	5	7,9,10,11, 14,15	Burnt Areas	n/a	Not measurable	Fire detection	Spatial extent of burnt scars	Continental, 10°x10° tiles Global	Temporal information on the fire season	Open access	Global Land Service MODIS Global Burned Area product	1999- present 2000- present	Operational
Trend in emission to the environment of pollutants relevant for biodiversity	_	8		Oil spill detection	Synthetic Aperture Radar (SAR)	Spatial extent not measurable	Tracking potential pollution events	oil slicks, vessels and installations at sea	Local to regional	Marine pollution represents a habitat disturbance	Open access for maritime administra tion in EU member states	CleanSeaNet Data Centre	2007- present	Operational
Trends in condition and vulnerability of ecosystems	_	5		Vegetation Condition Index	VCI	Not measurable	Compares the observed NDVI to the range of values in same period in previous years	Good or bad vegetation state as a percentage of normal range	Continental, 10°x10° tiles	Identify areas of poor or improving vegetation state on a qualitative basis	Open access	Global Land Service	2013- present	Operational
Trends in primary productivity		5		Vegetation Productivity Index	VPI	Not measurable	Compares the observed NDVI to NDVI value from previous years over the same 10-day period	Overall vegetation condition	Continental, 10°x10° tiles Africa and South America continental tiles	Useful to monitor growing season in – progress i.e. As an early warning system for anomalous change	Open access	Global Land Service GEONET Cast DevCoCast website	2013- present Near-real time only Aug 2007- present	Operational
Trends in condition and vulnerability of ecosystems		5		Sea Surface State	n/a	Offshore weather buoys	Radar Scatterometry (wind) Radar Altimetry, e.g. Jason-2 (wave height)	Wave height, direction, length and frequency	Regional seas and major oceans	Monitoring of extreme weather events with potential for marine habitat disturbance	Open access	ESA Globwave (satellite and <i>in-situ</i> data) Aviso (altimetry products)	Variable	Operational

Table 4.4A .Mapping of the adequacy of the use of remote sensing for the development of the indicators contained in Decision XI/3, for the strategic Goal A of the Strategic Plan for Biodiversity 2011-2020

			Magazinek			0 al al 61 a 11 a 1	Other		Global			
Target	Code	Operational Indicator	Measurable by RS	Metric/Proxy	EO product	Additional non-RS	requirements / standards	Spatial	Temporal	Sensor	Spatial	
1.	By 202	0, at the latest, people are awar	re of the values of	of biodiversity and	d the steps they can take	to conserve and	use it sustainably.					
	1	Frends in awareness and	NO									
		attitudes to biodiversity (C)										
	2	Frends in public engagement	NO									
		with biodiversity (C)										
	3	Frends in communication	NO									
		programmes and actions										
		promoting social corporate responsibility (C)										
2.		0, at the latest, biodiversity valu	ies have been in	tegrated into nati	ional and local developm	ent and poverty	reduction strategies a	nd planning pro	cesses and are beir	ng incorporate	d into national	acc
	4	Frends in number of countries										
	7	ncorporating natural	NO									
		esource, biodiversity, and										
		ecosystem service values into										
		hational accounting systems										
		В)										
	5	Frends in number of countries	NO									
		hat have assessed values of										
		piodiversity, in accordance										
		with the Convention (C)										
	6	Frends in guidelines and	NO									
		applications of economic										
	7	appraisal tools (C) Frends in integration of	NO									-
	<i>'</i>	piodiversity and ecosystem	NO									
		service values into sectoral										
		and development policies (C)										
	8	Frends in policies considering	NO									
		piodiversity and ecosystem										
		ervice in environmental										
		mpact assessment and										
		trategic environmental										
-		assessment (C)								L		
3.	-	D, at the latest, incentives, includ								positive incent	ives for the co	nser
	applied 9	, consistent and in harmony wit Frends in the number and	NO	n and other relev	ant international obligati	ions, taking into a	account national socio	economic cond	iitions.	1		
	9	value of incentives, including	NO									
		subsidies, harmful to										
		biodiversity, removed,										
		eformed or phased out (B)										
	10	Frends in identification,	NO									$\vdash$
	_	assessment and										
		establishment and										
		strengthening of incentives										
		hat reward positive										
		contribution to biodiversity										
		and ecosystem services and										
		penalize adverse impacts (C)										

Regional	-		National	
Temporal	Sensor	Spatial	Temporal	Sensor
counting, as app	propriate, and r	eporting sy	vstems.	
ervation and sus	tainable use of I	biodiversit	y are develope	ed and

11	Frends in population and	YES	intrinsic rate of	daily surface water	in situ weather			1					various	30d	microway
	extinction risk of utilized species, including species in trade (A)	TE3	increase,	inundation fraction, surface air temperature, soil moisture, and	station data								Various	50u	AMSR- Landsa
		1		microwave vegetation opacity											
12	Frends in ecological footprint and/or related concepts (C)	YES	natural capital consumption, area units	thematic classification		population model	low/medium	monthly/yearly	MODIS, Lansat, Sentinel 2	low/medium	monthly/yearly	MODIS, Landsat, Sentinel3	ow/mediu m	monthly/ yearly	MODI Landsa Sentine
13	Ecological limits assessed in terms of sustainable production and consumption (C)	YES	usd/ha	crop yield	ecosystem capacity	model - indirect				low/medium		MODIS/ Landsat/ Sentinel2	low/ medium	6months	MODI Landsa Sentine
14	Frends in biodiversity of cities C)	YES	green space - area unit, green infrastucture	classification		indirect							high/ medium	monthly/ yearly	ikono rapide Landsa sentino
	Frends in extent to which biodiversity and ecosystem service values are ncorporated into brganizational accounting and reporting (B)	NO													

Table 4.4B. Mapping of the adequacy of the use of remote sensing for the development of the indicators contained in Decision XI/3, for the strategic Goal B of the Strategic Plan for Biodiversity 2011-2020-

t (	Code	Operational indicator	Mesurable	Metrics/Proxy	EO product	Additional	Other requirements		Global			Regional			National	
		-	by RS	•	is at least halved and wh	non-RS data	/ standards	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor
	16	Extinction risk trends of habitat dependent species in each major	YES	NO												
1	17	habitat type (A) Trends in extent of selected biomes, ecosystems and habitats (A)	YES	surface circulation features	water surface vertical displacements			Large scale circulation features	weeks to months	radar altimeter	Large scale circulation features	weeks to months	radar altimeter			
1	18	Trends in proportion of degraded/threatened habitats (B)	YES	surface circulation features	Ocean Color, water surface vertical displacements					LiDAR, radar altimeter			LiDAR, radar altimeter			LiDAR, radar altimeter
1	19	Trends in fragmentation of natural habitats (B)	YES	area	classification, change detection map						medium/high	monthly/yearly	ikonos, rapideye, geoeye, landsat, sentinel2	medium/ high	monthly/yearly	ikonos, rapideye, geoeye, landsat, sentinel3
		Trends in condition and vulnerability of ecosystems (C)	YES	eco- environmental vulnerability index	spatial principle component analysis		elevation, slope, accumulated temperature, drought index, land use, vegetation, soil, water- soil erosion, and population density	low	year	modis				high	monthly	ikonos, rapideye, geoeye
		Trends in the proportion of natural habitats converted (C)	YES	area	classification, change detection map						medium/high	monthly/yearly	ikonos, rapideye, geoeye, landsat, sentinel2	medium/ high	monthly/yearly	ikonos, rapideye, geoeye, landsat, sentinel3
2		Trends in primary productivity (C)	YES	NPP	fAPAR, NDVI											
4		Trends in proportion of land affected by desertification (C)	YES	RUE	fAPAR, NDVI	precipation										
2	24	Population trends of habitat dependent species in each major habitat type (A)	YES	kg/km², mg/cu.m	echosounder echograms, fish school density, chlorophyl pigments	fish, seaweed samples	SST				m to km		Echosounder, sonar, lidar, Aerial photography	m to km	minutes to days	Echosounder sonar, lidar, Aerial photography

25	tened species and vulnerabl Trends in extinction	NO							T	Т		1	· · · · · · · · · · · · · · · · · · ·		
	risk of target and bycatch aquatic species (A)							ļ							
26	of target and bycatch aquatic species (A)	YES	kg/km², mg/cu.m	echosounder echograms, fish school density, chlorophyl pigments	fish, seaweed samples	SST				m to km		Echosounder, sonar, LiDAR, Aerial photography	m to km	minutes to days	Echosound sonar, LiDA Aerial photograp
27	Trends in proportion of utilized stocks outside safe biological limits (A) (MDG indicator 7.4)														
28	Trends in catch per unit effort (C)	NO	· · · · · · · · · · · · · · · · · · ·												
29		YES	Number of Boats	Aerial images											Airborne
30	Trends in area, frequency, and/or intensity of destructive fishing practices (C)	NO													
31	of depleted target and bycatch species with recovery plans (B)	NO													
	020 areas under agriculture				suring conservation	on of biodiversit	٧.								
32	Trends in population of forest and agriculture dependent species in production systems (B)	YES	%, unit	species map									high res	year	ikonos, rapideye
33		YES	usd/unit	yield estimation						1		,	high res	year	ikonos, rapideye
34	Trends in proportion of products derived from sustainable sources (C)	YES	%, loss of vegetation	classification, land cover change									high res	year	ikonos, rapideye
	Trends in area of forest, agricultural and aquaculture ecosystems under	YES	area	land cover map		land tenure	low/medium		MODIS/ Landsat	low/medium	year	MODIS/Landsat	low/medium	year	MODIS/L
35	sustainable management (B)	I		·											
		m excess n	utrients, has been b	rought to levels that are	not detrimental	to ecosystem fui	nction and biodi	versity.		km <sup>2</sup>					

37	Trends in water quality in aquatic ecosystems (A)	YES	water constituents	Water leaving radiance	water samples		km <sup>2</sup>	weeks- month	MODIS, Sentinel 3(OLCI)	km <sup>2</sup>	weeks-month	MODIS, Sentinel 3	km <sup>2</sup>	weeks-month	MODIS, Sentinel 3
38	Impact of pollution on extinction risk trends (B)	NO													
39	Trends in pollution deposition rate (B)	YES	meters	bathymetry											airborne, bathymetr LiDAR
40	Trends in sediment transfer rates (B)	NO													
41	Trend in emission to the environment of pollutants relevant for biodiversity (C)	YES		SAR images, Ocean Color	wind speed under certain threshold	proper sun glint correction	10 cm to meters		SAR, Sentinel 1	10 cm to meters		SAR/Sentinel 1	10 cm to meters		SAR/Sentii
42	Trend in levels of contaminants in wildlife (C)	NO													
43	Trends in nitrogen footprint of consumption activities (C)	NO													
44	Trends in ozone levels in natural ecosystems (C)	YES	ppmv, Dobson unit	ozone concentrations				1 or 8 days	Total Ozone Mapping Spectrometer (TOMS), the Solar Backscatter Ultraviolet Spectrometer (SBUV), and the Global Ozone Monitoring Experiment (GOME).						
45	Trends in proportion of wastewater discharged after treatment (C)	NO													
	Trends in UV- radiation levels (C)	YES	UV-A, UV-B	Ocean Color	use of a AERONET/OC network (CIMEL)	corection of aerosols			CIMEL sensors			CIMEL sensors			CIMEL ser
By 202	0, invasive alien species a	and pathway	s are identified and	d prioritized, priority spec	cies are controlled	d or eradicated,	and measures a	are in place t	o manage pathwa	ays to prevent	their introduction a	nd establishment.			
47	Trends in the impact of invasive alien species on extinction risk trends (A)	YES	area%	time series, land cover map	population dinamics model								medium/high	year	rapideye, ikonos
48	Trends in the economic impacts of selected invasive alien species (B)	YES	usd/output	time series, land cover map		econometric model							medium/high	year	rapideye, ikonos
49	Trends in number of invasive alien species	YES	area%	land cover, species distribution maps									medium/high	year	rapideye, ikonos

50	Trends in incidence of	NO	,			$\top$	, <u> </u>						
,	wildlife diseases	1	,				,	1	1			1	
I	caused by invasive	1	,				·   ·	- [ · · · ·	1			1	
· · · · ·	alien species (C)	1					'	_ <b>_</b> '					
51	Trends in policy	NO	,				'	1	1			1	
J	responses, legislation	1	,				,	1	1			1	
,	and management	1	,				'   '	1	1			1	
,	plans to control and	1	,				'   '	1	1			1	
I	prevent spread of	1	,				' 	1	1			1	
,	invasive alien species	1	,				,	1	1			1	
ا ا	(B)	<u> </u>	'				'	<u> </u>			'		
52	Trends in invasive	YES	area	land cover map			'	1	1		medium/high	year	rapideye
,	alien species	1	,				'	1	1			1	ikonos
I	pathways	1	,				'   '	1	1			1	
	management (C)	<u> </u>	'					'			'		
-			ires on coral reets,	-	ecosystems impacted by climate ch						-		
53	Extinction risk trends	YES	,	SST, Ocean Color		10 cm to	days to	MODIS, SAR	10 cm to km <sup>2</sup>		10 cm to km <sup>2</sup>	-	
I	of coral and reef fish	1	,			km <sup>2</sup>	months	1	1	months		months	
'	(A)	<u> </u>	'				'	<u> </u>	<u> </u>			<u> </u>	
54	Trends in climate	YES		SST, Ocean Color	wind speed	10 cm to	days to	MODIS, SAR	10 cm to km <sup>2</sup>		10 cm to km <sup>2</sup>	days to	
	change impacts on	1	2 nm -1,			km <sup>2</sup>	months	1	1	months		months	
'	extinction risk (B)	4	'				'						
55	Trends in coral reef	YES		SST, Ocean Color,	wind speed	10 cm to	days to	MODIS, SAR	10 cm to km <sup>2</sup>		10 cm to km <sup>2</sup>		
,	condition (B)	1	2 nm -1,	Insolation, SAR,		km <sup>2</sup>	months	1	1	months		months	
I		1	,	Ocean Surface			' 	1	1			1	
I		1	,	Vector Winds			· · ·	1	1			1	
	1	1	,				'	1				1	
		1	,				,	'	1				
56	Trends in extent, and	YES	area	land cover		10 cm to	days to	MODIS, SAR	10 cm to km <sup>2</sup>	days to	10 cm to km <sup>2</sup>	days to	
, <sup>2</sup> ,	rate of shifts of					km <sup>2</sup>	months			months		months	
, 1	boundaries, of	1	,				,	1	1				
ر ۱	vulnerable	1	,				,	1	1			1	
, 1	ecosystems (B)	1	,				' 	1	1			1	
57	Trends in climatic	NO		†		+		+	<u> </u>	+	· + · · · · · · · · · · · · · · · · · ·	<u> </u>	
, , , , , , , , , , , , , , , , , , ,	impacts on		,				,	1	1			1	
, ,	community	1	,				·   ·	1	1			1	
	composition (C)	1	,				' 	1	1			1	
· ,			·	-				· +'	+	+	· +	+	
58		NO				I		- I	1			1	
58	Trends in climatic impacts on	NO											

Table 4.4C. Mapping of the adequacy of the use of remote sensing for the development of the indicators contained in Decision XI/3, for the strategic Goal C of the Strategic Plan for Biodiversity 2011-2020

	Code					Addition	Other		Global			Region	al		National	
Target	Code	Operational Indicator	Mesurable by RS	Metrics / Proxy	EO product	al non- RS data	requirement s / standards	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor
11 represer	•	, at least 17 per cent of terres d well-connected	strial and inlar	nd water, and 10 p	er cent of coastal a	nd marine ar	reas, especially a	areas of particul	ar importance	for biodiversity and	ecosystem servi	ces, are consei	ved through effectively	and equitably n	nanaged, ecolo	gically
represer		of protected areas and othe	r effective are	a-based conservat	ion measures, and	integrated in	nto the wider lar	ndscapes and se	ascapes.							
		Trends in coverage of	YES	area	landcover	cadastral		low/medium		MODIS/landsat/	low/medium	month/year	MODIS/landsat/	low/medium	month/year	MOIDS/landsat
		protected areas (A)				DB				sentinel2			sentinel2			/sentinel3
		Trends in extent of marine protected areas, coverage of key biodiversity areas and management effectiveness (A)	YES	area	time series			low/medium	month/year	MODIS/landsat/ sentinel3	low/medium	month/year	MODIS/landsat/ sentinel3	low/medium	month/year	MODIS/landsat /sentinel4
-		Trends in protected area condition and/or management effectiveness including more equitable management (A)	YES		soil moisture, phenology			low/medium	daily	amser-e, aviris, WindSat, AMSR- E, RADARSAT, ERS-1-2, Metop/ASCAT	low/medium	daily	amser-e, aviris, WindSat, AMSR-E, RADARSAT, ERS-1-2, Metop/ASCAT	low/medium	daily	amser-e, aviris, WindSat, AMSR-E, RADARSAT, ERS-1-2, Metop/ASCAT
		Trends in representative coverage of protected areas and other area based approaches, including sites of particular importance for biodiversity, and of terrestrial, marine and inland water systems (A)	YES	area	landcover			low/medium	month/year	MODIS/landsat/ sentines3	low/medium	month/year	MODIS/landsat/ sentinel3	low/medium	month/year	MODIS/landsat /sentinel4
	63	Trends in the connectivity of protected areas and other area based approaches integrated into landscapes and seascapes (B)	YES	area	landcover			low/medium	month/year	MOIDS/landsat/ sentinel3	low/medium	month/year	MODIS/landsat/ sentinel3	low/medium	month/year	MODIS/landsat /sentinel4
-	64	Trends in the delivery of ecosystem services and equitable benefits from protected areas (C)	YES			socio- economi c data	baseline data	low/medium	month/year	MODIS/landsat/ sentines3	low/medium	month/year	MODIS/landsat/ sentinel3	low/medium	month/year	MODIS/landsat /sentinel4
12		the extinction of known thre	eatened specie	es has been preven	ted and their conse	ervation statu	us, particularly o	of those most in	decline, has be	een improved and su	ustained.					
		Trends in abundance of selected species (A)	YES	mm	landcover		rainfall							1-30m	2-16d	casi, sentinel, LiDAR
-	66	Trends in extinction risk of species (A)	YES	mm	landcover, species composition		rainfall							1-30m	2-16d	casi, sentinel, LiDAR
-		Trends in distribution of selected species (B)	YES	area	land cover		canopy structure, collard							1-30m	2-16d	slicer/elvis

13 By 202	By 2020, the genetic diversity of cultivated plants and farmed and domesticated animals and of wild relatives, including other socio-economically as well as culturally valuable species, is maintained, and strategies have been developed and implemented for minimizing														
genetic erosion	enetic erosion and														
safeg	safeguarding their genetic diversity.														
68	Trends in genetic diversity	NO													
	of cultivated plants, and														
	farmed and domesticated														
	animals and their wild														
	relatives (B)														
69	Trends in genetic diversity	NO													
	of selected species														
70	Trends in number of	NO													
	effective policy														
	mechanisms implemented														
	to reduce genetic erosion														
	and safeguard genetic														
	diversity related to plant														
	and animal genetic														
	resources (B)														

Table 4.4D. Mapping of the adequacy of the use of remote sensing for the development of the indicators contained in Decision XI/3, for the strategic Goal D of the Strategic Plan for Biodiversity 2011-2020.

			Measurable	Metrics /		Additional non-RS	Other		Global	T		Regional	T		National		
	Code	Operational Indicator	by RS	Proxy	EO product	data	requirements / standards	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor	
E able	-	0, ecosystems that provide es	sential services,	, including sei	vices related to	water, and contribute	to health, livelih	oods and well-b	eing, are restored	and safeguarde	ed, taking into ac	count the needs of	women, indige	enous and local co	mmunities, and	the poor and	
7		Trends in proportion of total freshwater resources used (A) (MDG indicator 7.5)	NO			seasonal water levels of large catchments											
7		Trends in proportion of the population using improved water services (A) (MDG indicator 7.8 and 7.9)	NO			trends in national statistics											
7		Trends in benefits that humans derive from selected ecosystem services (A)	YES	e.g. pollination potential	land cover/land use	species/population modeling	food provision				medium/high	30d	ikonos, rapideye Landsat Sentinel2	medium/high	30d	ikonos, rapideye, Landsat Sentinel3	
7		Population trends and extinction risk trends of species that provide ecosystem services (A)	NO														
7		Trends in delivery of multiple ecosystem services (B)	YES	delta/rate of change	time series	socio-economic data		low/medium	15,30,180,365D	MODIS Landsat Sentinel2	low/medium	15,30,180,365D	MODIS Landsat Sentinel3	low/medium	15,30,180,3 65D	MODIS/ Landsat/ Sentinel4	
7		Trends in economic and non-economic values of selected ecosystem services (B)	YES	npp, area, fpar, par	Above ground biomass, seasonal productivity and carbon sequestration			low/medium	daily	modis	low/medium	daily	MODIS	low/medium	daily	MODIS	
7		Trends in health and wellbeing of communities who depend directly on local ecosystem goods and services (B)	NO			health and socio- economic indicators, nutrition measures, food availability											
7		Trends in human and economic losses due to water or natural resource related disasters (B)	YES	usd	Land cover	socio-economic data								vhr/high	1 day	aerial/ ikono	
7	79	Trends in nutritional contribution of biodiversity: Food composition (B)	YES	area	Land cover	agricultural output					medium	30d	Landsat/ Sentinel2	medium	30d	landsat/ sentinel2	
8	30	Trends in incidence of emerging zoonotic diseases (C)	YES	area	water bodies		malaria				medium	30d	radar				
8	31	Trends in inclusive wealth (C)	YES	area, unit	urbanization map	socio-economic data								high	year	ikonos, geoeye	

			1		1			1	1		1			1	
82	Trends in nutritional	YES	unit	agriculture,						medium	30d	Landsat	medium	30d	Landsat
	contribution of			yield								Sentinel2			Sentinel2
	biodiversity: Food														
	consumption (C)														
83	Trends in prevalence of	NO			time series of										
	underweight children				national statistics										
	under-five years of age (C)				on children weight										
	(MDG indicator 1.8)				measures										
84	Trends in natural resource	YES	unit, area	mining map,									medium	year	Landsat
	conflicts (C)			deforestation											Sentinel2
				тар											
85	Trends in the condition of	YES	area	land cover,									medium	year	Landsat
	selected ecosystem			time series											Sentinel2
	services (C)														
86	Trends in biocapacity (C)	NO													
87	Trends in area of degraded	YES	area	land cover,									medium	year	Landsat
	ecosystems restored or			time series											Sentinel2
ation ar	being restored (B) 2020, ecosystem resilience and th nd to combating	e contribution	of biodiversit	y to carbon stock	s has been enhanced, t	hrough conserva	ation and resto	ration, including r	restoration of at	least 15 per cent	of degraded e	ecosystems, thereby	contributing to cl	imate change	mitigation and
ation ar dese	2020, ecosystem resilience and th nd to combating ertification.		1			hrough conserva					_				
ation ar	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent	e contribution YES	npp, area,	land cover,	s has been enhanced, t carbon model	hrough conserva	ation and resto		restoration of at	least 15 per cent	of degraded e	ecosystems, thereby MODIS	contributing to cl	imate change daily	mitigation and
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats		1	land cover, species		hrough conserva					_				
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon		npp, area,	land cover, species composition,		hrough conserva					_				
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats		npp, area,	land cover, species composition, ground		hrough conserva					_				
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon		npp, area,	land cover, species composition, ground biomass,		hrough conserva					_				
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon		npp, area,	land cover, species composition, ground biomass, seasonal		hrough conserva					_				
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon		npp, area,	land cover, species composition, ground biomass, seasonal productivity		hrough conserva					_				
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon		npp, area,	land cover, species composition, ground biomass, seasonal productivity and carbon		hrough conserva					_				
ation ar dese 88	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A)	YES	npp, area, fpar, par	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration	carbon model	hrough conserva					_		low/medium	daily	MODIS
ation ar dese	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A) Population trends of forest-		npp, area,	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration time series,	carbon model	hrough conserva					_				MODIS rapideye,
ation ar dese 88	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A) Population trends of forest- dependent species in	YES	npp, area, fpar, par	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration time series, land cover	carbon model	hrough conserva					_		low/medium	daily	MODIS
ation ar dese 88	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A) Population trends of forest- dependent species in forests under restoration	YES	npp, area, fpar, par	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration time series,	carbon model	hrough conserva					_		low/medium	daily	MODIS rapideye,
ation ar dese 88 89	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A) Population trends of forest- dependent species in forests under restoration (C)	YES	npp, area, fpar, par area%	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration time series, land cover map	carbon model population dinamics model		low/medium	daily	MODIS	low/medium	daily	MODIS	low/medium	daily	MODIS rapideye,
ation ar dese 88 89	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A) Population trends of forest- dependent species in forests under restoration	YES	npp, area, fpar, par area%	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration time series, land cover map	carbon model population dinamics model		low/medium	daily	MODIS	low/medium	daily	MODIS	low/medium	daily	MODIS rapideye,
ation ar dese 88 88 89 89	2020, ecosystem resilience and th nd to combating ertification. Status and trends in extent and condition of habitats that provide carbon storage (A) Population trends of forest- dependent species in forests under restoration (C)	YES YES	npp, area, fpar, par area%	land cover, species composition, ground biomass, seasonal productivity and carbon sequestration time series, land cover map	carbon model population dinamics model		low/medium	daily	MODIS	low/medium	daily	MODIS	low/medium	daily	MODIS rapideye,

Table 4.4E. Mapping of the adequacy of the use of remote sensing for the development of the indicators contained in Decision XI/3, for the strategic Goal E of the Strategic Plan for Biodiversity 2011-2020.

	Code	Operational Indicator	Measurable	Metrics	EO		Other	Global				Regional			National			
0	Code	Operational Indicator	by RS (Yes/No)	/ Proxy	product	Additional non-RS data	requirements / standards	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor	Spatial	Temporal	Sensor		
7	By 201	y 2015 each Party has developed, adopted as a policy instrument, and has commenced implementing an effective, participate																
	91	Trends in implementation of national biodiversity strategies and action plans, including development, comprehensiveness, adoption and implementation (B)	YES	area	landcover	land tenure	REDD							low/medium	1у	MODIS Landsat Sentinel2		
vant ir	nternati	ional			-	ocal communities relevant for the conservat						iological reso	urces, are re	spected, subject to	o national legis	lation and		
	obligat	· · · · · · · · · · · · · · · · · · ·	· · · · ·	mentation	1	ntion with the full and effective participatio	_	d local com	munities, at all i	relevant le	vels.			1	1			
	92	Trends in land-use change and land tenure in the traditional territories of indigenous and local communities (B)	YES	area	landcover	land tenure, indigenous territories maps	REDD							low/medium	1у	MODIS Landsat Sentinel2		
_	93	Trends in the practice of traditional occupations (B)	YES	area	landcover	land tenure, land use change analysis, changes in proportion of population engaged in traditional occupations,	REDD							low/medium	1y	MODIS Landsat Sentinel2		
	94	Trends in which traditional knowledge and practices are respected through their full integration, safeguards and the full and effective participation of indigenous and local communities in the national implementation of the Strategic Plan (B)	NO			Presence of indigenous organizations and linkages to national level decision making, number of laws protecting indigenous rights and resources at national level												
	95	Trends of linguistic diversity and numbers of speakers of indigenous languages (B)	NO			National level statistics, Number of indigenous languages included in national primary education systems												
9	By 202		echnologies rel	lating to bio	odiversity, it	s values, functioning, status and trends, and	the consequences	of its loss, a	re improved, w	idely share	ed and transf	ferred, and ap	oplied.			L		
1	96	Trends in coverage of comprehensive policy-relevant sub-global assessments including related capacity-building and knowledge transfer, plus trends in uptake into policy (B)	NO															
	97	Number of maintained species inventories being used to implement the Convention (C)	NO															
	By 20		inancial resourc	ces for effe	ctively imple	ementing the Strategic Plan for Biodiversity 2	011-2020 from all	sources, an	d in accordance	with the c	consolidated	<u> </u>				1		
	98	Indicators in Decision X/3	NO															

Table 4.5. Existing satellites and remote sensing sensors and their potential applications to track progress towards the Aichi Biodiversity Targets

Aichi Target	Category	Satellite	Sensors	Data Products (eg raw data or derived)	Uses specific to Aichi Targets	Sources	Start Year / End Year (if completed)	Geographical Coverage	Repeat Viewing Frequency (days)	Spatial Resolution (meters)	Availability	Gaps/Limitations
4,15	Optical/Passive Low Spatial High Temporal	Greenho use Gas Observati on SAT (GOSAT)	Thermal And Near infrared Sensor for carbon Observation - Fourier Transform Spectrometer (TANSO-FTS) Thermal And Near infrared Sensor for carbon Observation - Cloud and Aerosol Imager (TANSO-CAI)	Radiance Cloud cover Mapped CO2 & CH4 (abundance, vertical mixing, concentrations and vertical profile) CO2 flux and 3-D distribution concentration map) Normalized Difference Vegetation Index (NDVI) Global Radiance distribution Clear sky reflectance	Monitoring Impacts of use of natural resource consumption and production by combining monitoring of carbon emission and vegetation condition Measuring carbon stocks	Japanese Aerospace Exploration Agency (JAXA)	2009 (expected to last 5 years)	Global - atmospheric	3	500 - 1,500	Freely Available: At present, only one ACOS product is publicly available - ACOS_L2S. It is a Level-2 product that contains full physics retrievals of column- averaged CO2 in units of dry- air mole fraction (Xco2). <u>Restricted:</u> Level 1B product (with calibrated radiances and geolocation), which is the input to the ACOS Level-2 production process, is currently restricted by cooperation agreements between JAXA and NASA.	-Not all data products are available -Primary objective is on atmospheric monitoring of GHGs, not Earth Observation; -Is not a stand-alone resource for biodiversity monitoring and needs to be used in conjunction with modelling and other RS and non-RS data
4,15	Optical/Passive Medium Spatial and Temporal Resolution	Orbiting Carbon Observat ory (OCO)	Three high-resolution grating spectrometers; specifics and other sensors TBA	Orbit granules of calibrated radiances Orbit granules of geolocated Xco2 Global Xco2 Global CO2 sources and sinks	Monitoring Impacts of use of natural resource consumption and production by combining monitoring of carbon emission and vegetation condition Measuring carbon stocks	National Aeronautics and Space Administration (NASA)	2014	Global - atmospheric	16	TBA - medium/mo derate	Freely Available	-Initial launch failed in 2009, second launch was delayed form 2011 to 2014
5,11	Optical/Passive Medium - High Spatial and Temporal Resolution	Satellite The Sino- Brazilian Earth Observati on (CBERS) 1, 2, 2b, 3, 4, &4b	<ul> <li>(1, 2 &amp; 3) Wide Field Imager</li> <li>Camera (WFI); Medium</li> <li>Resolution Camera (CCD);</li> <li>Infrared Multispectral Scanner</li> <li>Camera (IRMSS)</li> <li>(3) High Resolution</li> <li>Panchromatic Camera (HRC)</li> <li>(3 &amp; 4) Advanced Wide Field</li> <li>Imager Camera (AWFI); IRMSS;</li> <li>Panchromatic and</li> <li>Multiespectral Camera</li> <li>(PANMUX)</li> <li>(4b) TBA</li> </ul>	Multispectral Images	Broad-Fine Scale Habitat Mapping Protected Area Monitoring	Instituto Nacional de Pesquisas Espaciais (INPE) Chinese Academy of Space Technology, China National space and Brazilian Space Agency	(1) 1999- 2003; (2) 2003; (2b) 2007-2010; 3 (2013); 4 (2014); 4b (2016)	Global	3, 5 , 26	(1&2) 20 (2b) 2.7 (3&4) 5 (4b) TBA	Freely Available to all Chinese and Brazilian people	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring, needs to be used in conjunction with other data, modelling and field information;</li> <li>-Cloud cover and haze create also challenges for monitoring using optical sensor;</li> <li>-Very High Resolution (VHR) optical datasets have been exploited or tested to their full extent and even in cloud free images, present pixel mixing and shadowing challenges;</li> <li>-The lack of shortwave infrared band and provision of too much detail present noise in the data and challenges in extracting the desired metrics;</li> <li>-Limited availability, may be prohibitively expensive and time consuming to procure and process.</li> </ul>

5,6,9,10,11, 12,14,15	Optical/Passive Medium-High Spatial Resolution High Temporal Resolution	Landsat 1-5, 7-8	(1-7) Multispectral Scanner (4-5)Thematic Mapper (TM) (7) Enhanced Thematic Mapper Plus(TM) (MSS)(8) Operational Land Imager (OLI); Thermal Infrared Sensor (TIRS)	Climate Data Records (CDR) such as surface reflectance, land surface temperature Essential Climate Variables (ECV): leaf area index, burned area extent, snow covered area, surface water extent Normalised Difference Vegetation Index (NDVI) (4-5, 7) Bathymetry, ocean colour, SST	Protected Area Monitoring Habitat mapping and change detection -capturing broad extent -spatial patterns of fragmentation Assessing Habitat Degredation -desertification -ocean acidification Biodiversity Assessment -Indicators of overall species richness and diversity -Tracking species distributions Ecological Monitoring -Mapping ecosystems -Assessing the effectiveness of ecosystem Landcover / Landcover change -quantifying the rate and extent of forest disturbance and re-growth Tracking pressures and threats -identifying disturbance Restoration projects	US Geological Survey (USGS)/NASA/Global Land Cover Facility (GLCF)	(1) 1972 (4) 1982– 1993, (5) 1994 (7) 1999	Global	(4-7) 16 days	(4-5) 30 meter+ (8) 15 meter+
5,9,11,12	Active Medium - High Spatial and Temporal Resolution	Multi- Applicati on Purpose Synthetic Apeture Radar (MAPSAR )	L-band synthetic aperture radar (SAR)	Cloud free multi-spectral Images	Landscape Monitoring Monitoring Landscapes and Disaster Events Resource Surveying Protected Area monitoring Landscape Monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure -Retrieving above ground biomass and structure (e.g., height, cover) -Assessing habitat condition Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure Tracking pressures, threats and disturbance -Detecting dead standing trees -Patterns of clearing and other damage caused by fire	Instituto Nacional de Pesquisas Espaciais (INPE) & Deutsches Zentrum für Luft-un Raumfahrt eV (DLR)	ТВА	Global	7	3 - 20

Landsat 4-5: Freely Available Landsat 5 and 7: Commercially & Freely available Landsat 8: At least 400 scenes are collected daily, and placed into the USGS archive to become available for download within 24 hours after acquisition	<ul> <li>-The Landsat surface reflectance CDR products are considered provisional;</li> <li>-Less effective at capturing good imagery in hyper-arid or snow- covered regions, areas with low sun angle conditions, coastal regions where land area is small relative to adjacent water and areas with extensive cloud contamination;</li> <li>-Users are strongly cautioned against correcting data acquired over high latitudes (&gt;65 degrees North or South);</li> <li>-Less able to provide information on changes in habitat quality, species distribution and fine-scale disturbances, than spaceborne optical sensors Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs be used in conjunction with other data, modelling and field information;</li> <li>-Limited ecosystem monitoring capacity, using landcover as a surrogate and must be combined with other data.</li> </ul>
ТВА	-Unknown at this time but is likely to have similar limitations as other SAR sensors and will not be a stand-alone product for monitoring biodiversity but will need to be combined with other data, modelling and field information; -L-band SAR is incapable of simultaneously providing high resolution and wide coverage.

	Outline!/D			<b></b>			Tama 4000	Clabal			Even also Assetta la la	la vista stavil elevi
5, 6, 10,11,15	Optical/Passive Course Spatial, High Temporal Resolution	Aqua	Advanced Spaceborne Thermal Emission and Reflection Radiometer (ASTER) Clouds and Earth's Radiant Energy System (CERES) Multi-angle Imaging Spectroradiometer (MISR) Moderate-resolution Imaging Spectroradiometer (MODIS) Measurements of Pollution in the Troposphere (MOPITT)	Numerous data products measuring Land, Ocean, Atmospheric, Cryospheric and Calibrationi parameters from both Terra and Aqua Sensors:	Monitoring Earth's atmosphere, lands, oceans, and radiant energy including: -measuring levels of gas in the lower atmosphere and tracking its source -monitoring ocean parameters, circulation, temperature, colour, etc. Very Broad-scale Habitat Monitoring and Degredation -Early warnings of regional ecological change and climate change (photosynthetic activity) including: -coral reef monitoring -comparing plant productivity with carbon dioxide and other important greenhouse gases, as well as global temperature trends to better enable scientists to predict how changes in the climate will impact Earth's ecosystems. Tacking Pressures and Threats (fires and photosynthetic activity) -identifying and monitoring ocean acidification -measure how certain human activities, such as biomass burning and deforestation, may be contributing to climate change -Near real-time alerts of deforestation		Terra: 1999 Aqua: 2002	Global	16	ASTER (15- 90) MISR (250- 275) MODIS (250- 1,000) CERES (20,000) MOPITT (22,000 at nadir)	Freely Available	-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information; -Course resolution; -Cloud cover and haze create challenges for monitoring using optical sensors.
5,11,12	Active Moderate - High Spatial Resolution Moderate - Low Temporal Resolution	d Land Observin g Satellite - Phased	Panchromatic Remote-sensing Instrument for Stereo Mapping (PRISM); Advanced Visible and Near Infrared Radiometer type 2 (AVNIR-2); Phased Array type L-band Synthetic Aperture Radar (PALSAR)	PALSAR data are in dual Polarization, HH+HV, mode. Bands HH (red and green) and Band-HV (blue) can be used to visualize land use patterns. The backscattering coefficient or Normalized Radar Cross Section (NRCS) are also provided as gray scale images.	Protected Area Monitoring Monitoring Landscapes and Disaster Events Resource Surveying Protected Area monitoring Landscape Monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure Tracking pressures and threats -Detecting dead standing trees -Patterns of clearing and	Japanese Aerospace Exploration Agency (JAXA)	Around 2007; completed 2011	Global	46	10	Freely Available	-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs be used in conjunction with data, modelling and field information; -Incapable of simultaneously providing high resolution and wide coverage.

		<b></b>							<del></del>			1
					other damage caused by fire							
5,10,11,12, 14, 15	Active Low Spatial and Temporal Resolution	ENVISAT	Advanced Synthetic Aperture Radar (ASAR); The Medium Resolution Imaging Spectroradiometer (MERIS)	GlobCover Bathymetry Sea Surface Height (SSH) sea colour (can be converted to chlorophyll pigment concentration, suspended sediment concentration and aero loads over marine areas) Cloud type, top height, and albedo Top and bottom indices of atmosphere vegetation Photosynthetically available radiation Surface pressure Water vapor total column content for all surfaces Aerosol load over land and sea Vegetation indices Fractional Absorbed Photosynthetically Active Radiation (FAPAR)	Protected Area monitoring Landscape Monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure -Coral reef monitoring Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure Tracking pressures and threats -Detecting dead standing trees -Patterns of clearing and other damage caused by fire -Identifying and monitoring ocean acidification Ecosystem monitoring Disaster management -detecting oil spills -monitoring floods, landslides, volcanic eruptions -aiding forest fighting		2002/3-2012 Globcover 2005-2006; 2009		35	300 meter	Commercially available from Radarsat International	- Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs be used in conjunction with data, modelling and field information; -Incapable of simultaneously providing high resolution and wide coverage (swath width).
5,10,11,12, 14,15	Active High Temporal and Spatial Resolution	Light Detectio n and Ranging (LiDAR) Remote Sensing	Laser scanner and photodetector/optical receiver	Point Cloud: A 3- dimensional (3D) dense assemblage of points with precise location of individual points hit by the laser, height of the object in the lasers path and intensity of the laser return (similar to optical reflectance only more concentrated and not influenced by cloud or other atmospheric disturbance to as great	Protected Area monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure	Multiple	Various	Airborne	1+	0.1 - 10	<u>Commercially and Freely</u> <u>Available</u> on case-by-case basis. Sources of freely available data include USGS & university/institutional collections	-Not currently utilised widely, effectively or efficiently though it is growing in popularity around the world; -Not available at global scale; -Costly to obtain data if not already available as requires flying a plane and operating cameras, software, expertise, etc.; -Requires formatting, importing and process which can create huge transaction (computing) costs and technical challenges to process data, the larger the study area the more time consuming, costly and

5,11,12,14,	Active	Radarsat	Synthetic Aperture Radar (SAR)	an extent as optical sensors are).	Tracking pressures and threats -Detecting dead standing trees -Patterns of clearing and other damage caused by fire	Government of Canada /	(1) 1995-	Global	RS-1 &-2	(RS-1) 8-100	Commercially Available	otherwise prohibitive to utilize; -LIDAR data handling software packages are not keeping pace with the LiDAR technology advancements, especially in automated classification and vegetation mapping; -Intensity must be calibrated when doing the flight campaign with targets and/or utilising correction algorithms for existing data as most LiDAR sensors are not calibrated for intensity; without calibrating intensity LiDAR is less useful for habitat and species monitoring; -Is not a stand-alone resource for biodiversity monitoring; the point clouds are used to generate other geospatial products, such as digital elevation models, canopy models, building models, and contours for monitoring/predicting trends in species changes, needs be used in conjunction with modelling and field information. -Is not a stand-alone resource for
15	Low-High Spatial Resolution Moderate-High Temporal Resolution	1 & 2 Radarsat Constella tion Mission (RCM)		images with change detection capacity	Protected Area MonitoringResource management-Forestry-monitoring growth and otherchangesHydrology-monitoring wateruse/consumptionOceanography-mapping sea ice distribution-maritime surveillance -improving shippingnavigationGeologyMeteorologyEcosystem monitoringDisaster management-detecting oil spills-monitoring floods,landslides, volcanic eruptions-aiding forest fightingSustainable developmentFine to Broad HabitatMapping and changedetection-Discriminating structurallycomplex habitats (e.g.,forests) based on 3DstructureAssessing habitat degradation-within structuredenvironments (canopy)Biodiversity assessment-Floral and faunal diversity inhabitats (e.g.,forest) withcomplex three-dimensionalstructure	Canadian Space Agency	(1) 1995 2012 (2) 2007 (7 year minimum duration) Constellation scheduled for 2018 launch		(24) RCM (12)	(RS-2 & RCM) 3 -100 / 1 + in Spotlight Mode		monitoring/predicting trends in species changes, needs be used in conjunction with modelling and field information; -Often insufficient for the purpose of detailed habitat mapping over large areas b/c of a fundamental incapability to simultaneously providing high resolution and wide coverage VHR and high resolution datasets suffer from problems of shadowing from and within objects and mixed pixels, and can be expensive and time consuming to procure and process.

		T	Treating anosource and	T	T	<u> </u>			
			Tracking pressures and threats -Detecting dead standing trees -Patterns of clearing and other damage caused by fire						
5,9,10,11,1 2	Optical/Passive       IKONOS       High resolution stereo imaging sensor (satellite based camera)         Resolution       High remporal Resolution       Sensor (satellite based camera)		Protected Area monitoring Ecological monitoring Habitat mapping and change detection -Mapping successional fine scale homogeneous habitats, ecotones and mosaic areas (e.g. coral reefs) Assessing habitat degradation -Identifying fine scale degradation in forests Biodiversity assessment -Indicators of overall species richness and diversity -Delineation of tree crowns/clumps to species level Tracking pressures and threats -Detection of fine-scale disturbances -Identification and monitoring of ocean acidification	GeoEye	1999	Global 1–3	1 (PAN) - 4 (MS	Commercially Available	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information;</li> <li>-IKONOS imagery may incur a high purchasing cost to the user;</li> <li>-Specialist hardware/software for utilising data may be required;</li> <li>-IKONOS data needed lengthy processing;</li> <li>-Visual interpretation of the IKONOS image necessitated fieldwork;</li> <li>-IKONOS images are not great for creating accuracy of vegetation classes with high spectral variance (heterogeneous)</li> <li>-Often insufficient for the purpose of habitat mapping over large areas;</li> <li>-Cloud cover and haze create challenges for monitoring using optical sensors;</li> <li>-Very High Resolution (VHR) and high resolution datasets have not yet been tested or exploited to their full extent and suffer from problems of shadowing and mixed pixels;</li> <li>-Can be prohibitively expensive and</li> </ul>

					1	1	1	1	1		
											time consuming to procure and process.
5, 10, 11,12,15	Optical/Passive Indian and Remo Radar/Active Sensir High to Low Satelli Spatial (IRS) Resolution Syster Moderate Temporal Resolution	te based sensors on 11 satellites in operation - largest civilian remote sensing satellite constellation in the world	The main data products are images in a variety of spatial, spectral and temporal resolutions utilised for a variety of applications with climate monitoring & environmental monitoring among them. The latest satellite to add to the constellation, SARAL includes biodiversity protection as a focused use case, focused on oceanographic studies.	Landscape Monitoring Protected Area Monitoring Habitat mapping and change detection -broad extent and spatial patterns Assessing habitat degradation -broad scale loss (i.e., desertification) Biodiversity assessment -Indicators of overall species richness and diversity Tracking pressures and threats -Identifying disturbances -Monitoring desertification	Indo-French collaboration built by the French National Space Agency (CNES) and the Indian Space Research Organisation (ISRO)	First satellite launched in 1988, The first of the still operational satellites in the constellation was launched in 2003 SARAL is scheduled for 2013	Global	various	various	Commercially Available	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information;</li> <li>-Limitations vary with individual satellites/sensors;</li> <li>SARAL will likely only benefit marine biodiversity monitoring;</li> <li>-Can be prohibitively expensive and time consuming to procure and process.</li> </ul>
5,10,11,12	Active Europ Moderate Remo Spatial Sensir Resolution Satelli Low to High 1 & 2 Temporal Resolution	te ng ite	Radar Imagery	Protected Area monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure -coral reef monitoring Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure Tracking pressures and threats -Detecting dead standing trees -Patterns of clearing and other damage caused by fire -Identifying and monitoring ocean acidification	European Space Agency (ESA)	(1) 1991– 2001; (2)1995– 2001	Global	3/35/336	50	Freely Available	-ls not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs be used in conjunction with other data, modelling and field information; -Incapable of simultaneously providing high resolution and wide coverage (swath width).

5,9,10,11,1 2, 14	Optical/Passive High Spatial Resolution High Temporal Resolution	QuickBird	Panchromatic (PAN) and multispectral (MS)	Three levels of imagery ranging from least processed/corrected to orthorectified, GIS ready. 1) Basic Imagery - black and white or multispectral imagery available by scenes (not georeferenced) 2) Standard Imagery - black and white, multispectral or pan sharpened imagery (is georeferenced) available by area of interest 3) Orthorectified Imagery - in addition to the Standard Imagery corrections it is terrain corrected and comes GIS ready as an Image basemap in black and white, multispectral or pan sharpened option; available by area of interest.	Protected Area monitoring Ecological monitoring Habitat mapping and change detection -Mapping successional fine scale homogeneous habitats, ecotones and mosaic areas Assessing habitat degradation -Identifying fine scale degradation in forests -rapid detection of clearing and degradation Biodiversity assessment -Indicators of overall species richness and diversity -Delineation of tree crowns/clumps to species level Tracking pressures and threats -Detection of fine-scale disturbances -identify and monitor ocean acidification	DigitalGlobe	2001	Global	4	<1 (PAN) - 2.4 -2.8 (MS)	Commercially Available	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information;</li> <li>-Often insufficient for the purpose of habitat mapping over large areas;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors:</li> <li>-Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges;</li> <li>-Can be prohibitively expensive and time consuming to procure and process.</li> </ul>
5,11,12,14, 15	Optical/Passive Medium-High Spatial Resolution High Temporal Resolution	Système Pour l'Observa tion de la Terre (SPOT)	Panchromatic (PAN) and multispectral (MS) , infrared and SWIR	A range of high resolution, multipspectral NIR and SWIR imagery with or without orthorectification	Protected Area Monitoring Ecological Monitoring Fine-scale Habitat Monitoring -rapid detection of habitat and degradation Biodiversity assessment -Indicators of overall species richness and diversity Tracking pressures and threats -Identifying disturbances -Monitoring droughts and desertification Agricultural monitoring -crop yields Oceanography Climatology	Astrium	SPOT 1 (1986-1990) SPOT 2 (1990-2009) SPOT 3 (1993-1997) SPOT 4 (1998-2013) SPOT 5 (2002) SPOT 6 (2012) SPOT 7 scheduled for 2014	Global	1-4 Tasking optional with 1 day revisit	SPOT 1-4 (10- 20) SPOT 5 (2.5- 5) SPOT 6-7 (1.5)	Commercially Available	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges;</li> <li>-Can be prohibitively expensive and time consuming to procure and process.</li> </ul>
5,6,10	Optical/Passive Low Spatial Resolution High Temporal Resolution	Sea- viewing Wide Field-of- view Sensor (SeaWiFS )	Optical scanner	Angstrom Exponent Aerosol Optical Thickness Chlorophyll-chromophoric dissolved organic matter (CDOM) proportion index Chlorophyll a Photosynthetically Available Radiation Particulate Inorganic/Organic Carbon concentration Sea Surface Temperature Quality Sea surface Reflectance Sea Surface Temperature	monitor coral reefs and ocean acidification	GeoEye	1997–2010	Global	1-2	1,100	Freely Available	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information;</li> <li>-Ocean focused;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges.</li> </ul>

5,10,11,14	Optical/Passive	Advance	AVHRR 1 included a 4 channel	Imagery available in four		National Oceanic and	1978-?	Global	6	1,100	Freely Available	-Not particularly useful for habitat
	Low Spatial Resolution High Temporal Resolution	d Very High Resolutio n Radiomet er (1-3)	radiometer AVHRR 2 include 5 channel radiometer AVHRR 3 includes a 6 channel radiometer	data sets: The Global Area Coverage (GAC) data set The Local Area Coverage (LAC) data set High Resolution Picture Transmission (HRPT) is real-time downlink data Full Resolution Area Coverage (FRAC )	Very Broad-scale Habitat Monitoring and Degredation -Early warnings of regional ecological change and climate change (photosynthetic activity) -Near real-time alerts of deforestation Tacking Pressures and Threats (fires and photosynthetic activity) Protected Area Monitoring Ecological Monitoring -coral reefs and ocean acidification	Atmospheric Association (NOAA)	1981-? 1998-?					<ul> <li>mapping;</li> <li>-Not useful for change detection or biodiversity assessment;</li> <li>-Limited ecosystem monitoring capacity, using landcover as a surrogate and must be combined with other data;</li> <li>-Early data products suffered from difficulties with sensor calibration, orbital drift, limited spectral and directional sampling;</li> <li>-Is not a stand-alone resource for biodiversity monitoring needs to be used in conjunction with other data, modelling and field information;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and</li> </ul>
5,10, 15	Optical/Passive Low Spatial Resolution High Temporal Resolution	Aquarius	Specialised radiometer	Sea Surface Salinity (SSS)	monitor coral reefs and ocean acidification supplements observations of precipitation, evaporation, soil moisture, atmospheric water vapor, and sea ice extent	National Aeronautics and Space Administration (NASA)	2011	Global	7	150	Freely Available	<ul> <li>mixed pixel challenges.</li> <li>-Is not a stand-alone resource for biodiversity monitoring needs to be used in conjunction with other data, modelling and field information;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Ocean focused</li> </ul>
5,6, 10, 11	Optical/Passive Moderate Spatial Resolution High Temporal Resolution	Seawinds : Quikscat	Specialised radiometer	Surface Wind Vector (SWV)	monitor coral reefs and ocean acidification ocean response air-sea interaction mechanisms annual and semi-annual rainforest vegetation conditions daily or seasonal ice edge/ice pack movement and changes	National Oceanic and Atmospheric Association (NOAA)	1999-2009	Global	1	12.5-25	Freely Available	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Ocean focused</li> </ul>
5,9,11,12	Optical/Passive - Hyperspectral High Spatial Resolution High Temporal Resolution	WorldVie w-2	Multispectral sensor (MS)	high resolution Panchromatic band and eight (8) Multispectral bands; four (4) standard colors (red, green, blue, and near-infrared 1) and four (4) new bands (coastal, yellow, red edge, and near-infrared 2), full- color images	Protected Area monitoring Ecological monitoring Habitat mapping and change detection -Mapping successional fine- scale homogeneous habitats, ecotones and mosaic areas Assessing habitat degradation -Identifying fine scale degradation in forests Biodiversity assessment -Indicators of overall species richness and diversity -Delineation of tree crowns/clumps to species level	DigitalGlobe	2009	Global	1	0.46 (PAN) 1.84 (MS)	Commercially Available	<ul> <li>-Ocean rocused</li> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs be used in conjunction with other data, modelling and field information;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges;</li> <li>-Can be prohibitively expensive and time consuming to procure and</li> </ul>

					•					
					Tracking pressures and threats -Detection of fine-scale disturbances					
5,9,11,12	Optical/Passive - Hyperspectral High Spatial Resolution High Temporal Resolution	Airborne	Airborne Hyperspectral imaging sensor (HyMAP)	Hyperspectral imagery spanning 126 spectral bands	Habitat mapping and change detection -Distinguishing habitat types in low-contrast environments, and identifying forest successioinal classes Assessing habitat degradation -based on changes in chemical composition of vegetation Biodiversity assessment -High precision classification of plant communities -Mapping top canopy trees to species or genus level - identifying invasive species -Relating spectral heterogeneity to species richness and diversity Tracking pressures and threats -Identifying disturbances based on changes in foliage color, and fine-scale modifications due to disturbance	Spectronics	1999	Airborne	Airborne	5
5,9,11,12	Optical/Passive - Hyperspectral High Spatial Resolution High Temporal Resolution	Airborne	Airborne Visible/Infrared Imaging Spectrometer (AVIRIS)	calibrated images of the upwelling spectral radiance in 224 contiguous spectral channels (bands) with wavelengths from 400 to 2500 nanometers.	Habitat mapping and change detection -Distinguishing habitat types in low-contrast environments, and identifying forest successioinal classes Assessing habitat degradation -based on changes in chemical composition of vegetation Biodiversity assessment -High precision classification of plant communities -Mapping top canopy trees to species or genus level	National Aeronautics and Space Administration (NASA)	First developed in 1983, updated in 2012	Airborne	Airborne	2

	process.
	le net e stend elene necessar fen
Commercially available	-Is not a stand-alone resource for
	biodiversity monitoring/predicting
	trends in species changes, needs be used in conjunction with other
	data, modelling and field
	information;
	-Often insufficient for the purpose
	of detailed habitat mapping over
	large areas;
	-Cloud cover and haze present
	challenges for monitoring with
	optical sensors;
	-Very High Resolution (VHR) optical
	datasets have not yet been
	exploited or tested to their full
	extent and even in cloud free
	images, present shadowing and
	mixed pixel challenges;
	-The shape and orientation of tree
	crowns, solar illumination, and
	sensor geometry, topography and
	spectral variation exert enormous
	influence over airborne
	spectroscopic signatures;
	-Very high-performance airborne
	HiFIS are needed at spatial
	resolutions that can resolve
	individual tree crowns, which is
	necessary for species-level
	determinations;
	-Can be prohibitively expensive and
	time consuming to procure and
	process.
Freely and commercially	- Is not a stand-alone resource for
available	biodiversity monitoring/predicting
	trends in species changes, needs be
	used in conjunction with other
	data, modelling and field
	information;
	-Only data from 2006-2013 is
	currently downloadable, pre 2006
	data is processed on request if
	nassihla
	possible;
	-Often insufficient for the purpose
	-Often insufficient for the purpose of detailed habitat mapping over
	-Often insufficient for the purpose of detailed habitat mapping over large areas;
	-Often insufficient for the purpose of detailed habitat mapping over large areas; -Cloud cover and haze present
	-Often insufficient for the purpose of detailed habitat mapping over large areas; -Cloud cover and haze present challenges for monitoring with
	-Often insufficient for the purpose of detailed habitat mapping over large areas; -Cloud cover and haze present

					<ul> <li>identifying invasive species</li> <li>Relating spectral</li> <li>heterogeneity to species</li> <li>richness and diversity</li> <li>Tracking pressures and</li> <li>threats</li> <li>-Identifying disturbances</li> <li>based on changes in foliage</li> <li>color, and fine-scale</li> <li>modifications due to</li> <li>disturbance</li> </ul>							<ul> <li>-Very High Resolution (VHR) and High Resolution optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges;</li> <li>-The shape and orientation of tree crowns, solar illumination, and sensor geometry, topography and spectral variation exert enormous influence over airborne spectroscopic signatures;</li> <li>-Very high-performance airborne HiFIS are needed at spatial resolutions that can resolve individual tree crowns, which is necessary for species-level determinations;</li> <li>-Can be prohibitively expensive and time consuming to procure and process.</li> </ul>
5,11,12	High - Moderate	TerraSAR -X and Tandem- X	Synthetic Aperture Radar (SAR)	WorldDEM: a homogenous, worldwide digital elevation model data (DEM) Additional individual image products	Protected Area monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure Tracking pressures and threats -Detecting dead standing trees -Patterns of clearing and other damage caused by fire	German Aerospace Center (DLR) and EADS Astrium	TerraSAR - 2007 TandemX - 2010	Global	11 (3-4 at poles) Tasking 1-3	1-18 for individual products 2-10 for WorldDEM	Commercially Available	<ul> <li>Often insufficient for the purpose of detailed habitat mapping over large areas;</li> <li>VHR and high resolution datasets suffer from problems of shadowing from and within objects and mixed pixels;</li> <li>Incapable of simultaneously providing high resolution and wide coverage (swath width);</li> <li>Can be expensive and time consuming to procure and process.</li> </ul>
5,9,11,12	Optical/Passive - Hyperspectral Moderate Spatial and Temporal Resolution	E0-1	High resolution hyperspectral imager capable of resolving 220 spectral bands (Hyperion) Advanced Land Imager (ALI) Linear Etalon Imaging Spectrometer Array (LEISA) Atmospheric Corrector (LAC)	Hyperion - High resolution hyperspectral images ALI - panchromatic and multispectral	Habitat mapping and change detection -Distinguishing habitat types in low-contrast environments, and identifying forest successioinal classes Assessing habitat degradation -based on changes in chemical composition of vegetation Biodiversity assessment -High precision classification of plant communities -Mapping top canopy trees to species or genus level - identifying invasive species -Relating spectral heterogeneity to species	National Aeronautics and Space Administration (NASA)	2000	Global	16	30	Freely available	-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information; -Cloud cover and haze present challenges for monitoring with optical sensors.

					richness and diversity Tracking pressures and threats -Identifying disturbances based on changes in foliage color, and fine-scale modifications due to disturbance					
5,11,12	Active Radar Moderate Spatial Resolution Low Temporal Resolution	JERS-1 SAR	An L-band (HH polarization) synthetic aperture radar (SAR); A nadir-pointing optical camera (OPS); A side-looking optical camera (AVNIR).	Radar and optical Imagery data available spanning seven bands from the visible region to short wave infrared band and is capable of stereoscopic data in NIR	Protected Area monitoring Habitat mapping and change detection -Discriminating structurally complex habitats (e.g., forests) based on 3D structure Assessing habitat degradation -even within structured environments (canopy) Biodiversity assessment -Floral and faunal diversity in habitats (e.g.,forest) with complex three-dimensional structure Tracking pressures and threats Land surveys Agricultural-forestry-fisheries Disaster prevention and monitoring Coastal surveillance Locating natural resources.	Japanese Aerospace Exploration Agency (JAXA)	1992-1998	Global	44	18
5,9,11,12	Optical/Passive - Hyperspectral High Spatial and Temporal Resolution	Airborne	Compact Airborne Spectrographic Imager (CASI)	Multispectral imagery	Habitat mapping and change detection -Distinguishing habitat types in low-contrast environments, and identifying forest successioinal classes Assessing habitat degradation -based on changes in chemical composition of vegetation Biodiversity assessment -High precision classification of plant communities -Mapping top canopy trees to species or genus level - identifying invasive species -Relating spectral heterogeneity to species richness and diversity Tracking pressures and threats -Identifying disturbances based on changes in foliage color, and fine-scale modifications due to disturbance	Itres Research Ltd. of Calgary, Canada	Various	Airborne	Airborne	1+

Freely available	-No longer operational -Cannot easily differentiate between species in high heterogeneity habitats, shadowing and mixed pixels can present challenges for mapping detailed habitats over large areas; -Not great for change detection due to inactivity, low temporal resolution and inconsistency in classifying heterogeneous images; -May have difficulty finding complementary/supporting data sets (e.g. DEMs) in tropics; -The L-band is incapable of simultaneously providing high resolution and wide coverage.
Publically Available (may not be free)	<ul> <li>-Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data modelling and field information;</li> <li>-Often insufficient for the purpose of detailed habitat mapping over large areas;</li> <li>-Cloud cover and haze present challenges for monitoring with optical sensors;</li> <li>-Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges;</li> <li>-The shape and orientation of tree crowns, solar illumination, and sensor geometry, topography and spectral variation exert enormous influence over airborne spectroscopic signatures;</li> <li>-Very high-performance airborne HiFIS are needed at spatial resolutions that can resolve</li> </ul>
	Publically Available (may not

5,9,11,12	Optical and	Airborne	High-fidelity Imaging	two-dimensional image,	Habitat mapping and change	Carnegie Airborne Observ	Various	Airborne	Airborne	<1+
	Chemical Passive High Spatial and Temporal Resolution		Spectrometers (HiFIS)	but with a third dimension containing a detailed spectroscopic signature of plant canopies.	detection -Distinguishing habitat types in low-contrast environments, and identifying forest successioinal classes Assessing habitat degradation -based on changes in chemical composition of vegetation Biodiversity assessment -High precision classification of plant communities -Mapping top canopy trees to species or genus level - identifying invasive species -Relating spectral heterogeneity to species richness and diversity Tracking pressures and threats -Identifying disturbances based on changes in foliage color, and fine-scale modifications due to disturbance	atory	2012			100.250
4,5,10,11, 12, 14, 15	Optical/Passive Low Spatial Resolution High Temporal Resolution	Proba V	Vegetation Instrument	multispectral images: VNIR: -Blue(438-486nm) -Red(615-696nm) -NearIR(772-914nm) SWIR(1564-1634nm)	-Land observation with focus on vegetation -Environmental & agro- climatic conditions -Effects of extreme events as drought and floods -Natural resources (soil, water, rangeland) -Crop and livestock production; -Prevalence of diseases	European Space Agency (ESA)	2013	Global	1-2	100-350

individual tree crowns, which is necessary for species-level determinations; -Can be prohibitively expensive and time consuming to procure and process.
-Although HiFIS has come of age technologically, the theories and algorithms required to extract taxonomic information from the spectra remain in the early stages of development; -Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information; -Often insufficient for the purpose of detailed habitat mapping over large areas; -Cloud cover and haze present challenges for monitoring with optical sensors; -Very High Resolution (VHR) optical datasets have not yet been exploited or tested to their full extent and even in cloud free images, present shadowing and mixed pixel challenges; -The shape and orientation of tree crowns, solar illumination, and sensor geometry, topography and spectral variation exert enormous influence over airborne spectroscopic signatures; -Very high-performance airborne HiFIS are needed at spatial resolutions that can resolve individual tree crowns, which is necessary for species-level determinations; -Can be prohibitively expensive and time consuming to procure and process.
Process. -Primarily a technology test -Expected to have s short life span of 2.5 years -Is not a stand-alone resource for biodiversity monitoring/predicting trends in species changes, needs to be used in conjunction with other data, modelling and field information; -Cloud cover and haze present challenges for monitoring with 120

-Desertification		optical sensors;
		-Can be prohibitively expensive and
		-Can be prohibitively expensive and time consuming to procure and
		process.

# Annex 5. Relative costs of using remote sensing for biodiversity monitoring

## 5.1. Data production

Data can be produced by public institutions, such as space agencies and national geo-spatial agencies, or via commercial companies. Some spaces agencies have adopted an open access data policy, offering free data to virtually all users. Nonetheless, a full and open access data policy does not necessarily mean easy and fast data access, and sometimes distribution of imagery can be subject of a fee depending on the type of user agreement in place. For more details see section 4.1.2.

High resolution imagery is usually available via commercial companies and costs vary depending on the remote sense technology used, amount of imagery requested, and specific agreement with the data provider.

Costs of the most common and popular satellite products are summarized in table 3.1. Prices are in USA dollars (\$) per image as estimated in mid- $2013^4$ .

aDI	ble 3.1. Costs of the most common and popular satellite products as of mid-2013							
	Satellite (sensor)	Pixel size (m)	Minimum order area	Approx. cost (\$)				
			(sq. km)					
	NOAA (AVHRR)	1100	Free	No cost				
	EOS (MODIS)	250, 500, 1000	Free	No cost				
	SPOT-VGT	1000	Free	No cost				
	LANDSAT	15, 30, 60, 100, 120	Free	No cost				
	ENVISAT (MERIS)	300	Free	No cost				
	ENVISAT (ASAR)	150	Free	No cost				
	SRTM (DEM)	90	Free	No cost				
	EO-1 (Hyperion)	30	Free	No cost				
	EOS (ASTER)	15, 30, 90	3600	100				
	SPOT-4	10, 20	3600	1,600 - 2,500				
	SPOT-5	2.5, 5, 10	400	1,300 – 4,000				
	SPOT-6	1.5, 6.0	500	1,000 – 3,000				
	RapidEye	5	500	700				
	IKONOS	1, 4	100	1,000 - 2,000				
	QuickBird	0.6, 2.4	100	2,500				
	GeoEye	0.25, 1.65	100	2,000 – 4,000				
	WorldView	0.5, 2, 4	100	2,600 – 7,400				

#### Table 3.1. Costs of the most common and popular satellite products as of mid-2013

Source. IKONOS, QuickBird, GeoEye, WorldView and RapidEye: Landinfo. SPOT 4 & 5: Astrium EADS. Aster: GeoVAR. SRTM DEM, Landsat, Hyperion, MERIS, ASAR, AVHRR, SPOT-VGT and MODIS: NASA, ESA and Land Cover Facility

<sup>&</sup>lt;sup>4</sup> This price is for the buying of a sinlge image. If large amount of images are bought, price per single image may decrease.

# 5.2. Data analysis

Data can be analysed either *in house* or be outsourced. Space Agencies most often analyse their own data as they have the required expertise. Agencies at the national, provincial and local level might outsource the process to commercial companies offering the service, which they cost according to the amount of work and level of complexity.

## 5.3. Data validation

Companies or institutions creating the data would verify it as part of the creation process, but verification and updating may also be done by those experts who have knowledge of the specific area. The cost are usually incurred at the point of data editing, or in the case of the expert being requested for their input the cost incurred could be equal to that of their hourly rate.

### 5.4. Other costs

Besides the above costs, there are a number of other costs associated with the use of Earth Observation for biodiversity mapping and monitoring that need to be taken into account. The key categories to consider are:

- Hardware and software costs
- Training and support costs
- Age and frequency of the EO data required
- Type of EO product to purchase

The following examples illustrate the broad costs for each of the above categories in USA dollars (\$), as estimated in mid-2013. However, it is an estimate, and advice from suppliers of services and products should be foreseen to refine the estimates. The estimates provided below reflect the basic versions of commercial products which could be used to support the various image processing and analysis requirements.

### 5.4.1. Hardware and software costs

Hardware requirements can/should include:

- Production based computer: \$2,000 \$4,000
- Plotter (or large format color printer) \$4,500 \$13,500

Software requirements can include:

- Image processing package
  - ERDAS Imagine Professional \$13,500 for 1 license
  - Exelis ENVI (no versioning) \$4,500 for 1 license
- Desktop GIS package to allow integration of datasets, GIS analysis functions
  - ArcGIS 10 \$3,000
  - MapInfo \$2,000
- Free and open Source GIS software
  - ILWIS 3.8 Open source and free of charge, <u>http://52north.org/</u>
  - GRASS GIS <u>http://grass.osgeo.org/</u>

- o gvSIG <u>http://www.gvsig.com/</u>
- OpenJUMP GIS <u>http://www.openjump.org/</u>
- MapWindow GIS <u>http://www.mapwindow.org/</u>
- o QGIS <u>http://www.qgis.org/en/site/</u>
- uDig <u>http://udig.refractions.net/</u>

# 5.4.2 Training and support costs

Depending on the complexity of the earth observation monitoring using remote sensed data with support of field data should be 2-4 person weeks of effort (also depending on size of area). In addition:

- GIS and Remote Sensing expertise would be required
- Training can be provided, or personnel can be hired

A key factor influencing the decision to hire specialists or to invest in-house is whether the inventory and future monitoring is going to be done frequently or not. For short duration work perhaps only performed every three years, it is likely that consistent product quality will not be possible using in-house personnel that are infrequently using their skills. Instead, hiring external services and working with them closely to ensure the quality will yield the best results.

# 5.4.3. Age and frequency of the EO data required

Data costs are affected by:

- Urgency emergency services the faster you need it, the higher the cost.
- Age of the data the older the data, the less expensive it is.
- Spatial resolution the higher the spatial resolution, the higher the cost.
- Level of the product the higher level image processing, the higher the cost.